

Conceptual Models of Stressors and Limiting Factors for San Joaquin River Chinook Salmon



These conceptual models were prepared by the San Joaquin River Restoration Program (SJRRP) Team as a draft document in support of preparing a Program Environmental Impact Statement/Report (PEIS/R). The models define the role of, and expectations, strategy, and selection of water operations modeling for, the SJRRP.

The purpose for circulating this document at this time is to facilitate early coordination regarding initial concepts and approaches currently under consideration by the Program Team with the Settling Parties, the Third Parties, other stakeholders, and interested members of the public. Therefore, the content of this document may not necessarily be included in the PEIS/R. These draft conceptual models do not present findings, decisions, or policy statements of any of the Implementing Agencies. Additionally, all information presented in this document is intended to be consistent with the Settlement. To the extent inconsistencies exist, the Settlement should be the controlling document and the information in this document will be revised before its inclusion in future documents. While the Program Team is not requesting formal comments on this document, all comments received will be considered in refining the concepts and approaches described herein to the extent possible. Responses to comments will not be provided and this document will not be finalized; however, refinements will likely be reflected in subsequent SJRRP documents.

Executive Summary

This document describes the life history requirements and the environmental factors most likely affecting the abundance of San Joaquin River spring-run and fall-run Chinook salmon (*Oncorhynchus tshawytscha*) in the Restoration Area (for the purpose of this document includes the San Joaquin River between Friant Dam and the Merced River confluence) and downstream from the Restoration Area in the lower San Joaquin River, the Sacramento-San Joaquin Delta (Delta), San Francisco Estuary, and the Pacific Ocean. This document also describes (1) the historical status of Chinook salmon in the San Joaquin River prior to the construction of Friant Dam, (2) the life history and habitat requirements of Chinook salmon in the Central Valley, (3) potential stressors of Chinook salmon in the San Joaquin River basin, (4) a limiting factors assessment of the fall-run Chinook salmon populations in the Stanislaus and Tuolumne rivers, (5) conceptual models of the likely mechanisms by which controllable and uncontrollable environmental factors affect the abundance and recovery of spring-run and fall-run Chinook salmon populations in the San Joaquin River basin, and (6) data needs (i.e., knowledge gaps) for spring-run and fall-run Chinook salmon in the San Joaquin River Basin. The limiting factors assessment assumes that all restoration actions prescribed in the Settlement will be implemented. The conceptual models will be used to assist in evaluating programmatic alternatives, guiding flow management, and identifying key habitat restoration needs. As part of an adaptive management process, monitoring data will be used to continually refine the conceptual models and revise management and restoration priorities. The conceptual models will also be used to help develop quantitative population models that will in turn help refine the hypotheses to be tested under an Adaptive Management Plan. As new information becomes available and restoration actions begin, the conceptual models will be revised accordingly.

Historical Population Dynamics

From 1942 to 1947, adult spring-run Chinook salmon escapement to the San Joaquin River ranged between 5,000 and 56,000 during years when spring flows below Sack Dam were at least 83 cubic feet per second (cfs). Escapement dropped from about 2,000 fish in 1948 to 36 fish in 1950, as spring flows below Sack Dam declined to less than 3 cfs. Escapement surveys were not conducted after 1950 when the run was considered to be extirpated from the San Joaquin River. It is assumed that the San Joaquin River supported relatively few fall-run Chinook salmon after diversions began at Sack Dam sometime between 1860 and 1880.

Life History Requirements

Central Valley Chinook salmon exhibit two general freshwater life history types, “stream-type” and “ocean-type.” Central Valley spring-run Chinook salmon are generally classified as stream-type because the adults migrate into mid-elevation watersheds where they spend several months while maturing sexually. In contrast, Central Valley fall-run Chinook salmon are considered ocean-type, because the adults spawn in the lower watersheds within a few weeks of entering fresh water, and juveniles spend only a short time in the river, typically migrating to the ocean within a few months after emerging from the gravel in the redd (i.e., nest).

Stressors

Stressors are physical, chemical, or biological perturbations to a system that adversely affect ecosystem processes, habitats, and species. Examples include altered flows, blocked passage, blocked sediment recruitment, instream habitat alteration, invasive species, contaminants, and excessive salmon harvest. Stressors are discussed according to each life history stage of Chinook salmon, including egg survival and emergence, juvenile rearing, smoltification and downstream migration, ocean survival, adult migration, adult holding for spring-run salmon, and spawning. In addition, the potential effects of releasing hatchery-reared juvenile salmon and climate change are discussed in terms of recovering naturally spawning populations.

Limiting Factor Analysis

Limiting factors are the stressors that substantially influence the abundance and productivity of San Joaquin River adult salmon. Analyses of limiting factors examine the relationships between fish production (either adults or juveniles), parental stock (spawner) abundance, and key environmental conditions over time. These analyses are used to identify the critical life history stage that affects the production of adults, and to determine the primary environmental factor(s) affecting that life history stage. There is often a lack of data on many of the environmental factors that might affect the critical life history stage(s); thus, the analysis cannot identify all potential limiting factors.

Therefore, it was assumed that any stressor that affects a critical life history stage should be considered a potential limiting factor.

A population-based analysis of environmental factors that limit the production of Chinook salmon in the Restoration Area cannot directly assess existing conditions in the Restoration Area because of the current absence of Chinook salmon. The spring-run Chinook salmon escapement data collected in the Restoration Area during the 1940s were not used because the existing habitat in the Restoration Area and the Delta has been extensively degraded since the late 1940s. An alternative presented here utilized the population trend analyses conducted for fall-run Chinook salmon populations in the lower Stanislaus, Tuolumne, and Merced rivers (hereafter referred to as the “Tributaries”). Habitats that affect population dynamics in the Tributaries are similar to those that will affect spring-run and fall-run populations in the San Joaquin River. For example, the Restoration Area and all of the Tributaries have been degraded by extensive gravel mining and loss of functional floodplain habitats. As the SJRRP moves forward, data collected on Chinook salmon populations in the Restoration Area will be used to refine limiting factors analyses and conceptual models.

A few substantial differences between spring-run and fall-run Chinook salmon and between the San Joaquin River and the Tributaries cannot be assessed with a limiting factors analysis based on the fall-run Chinook salmon populations in the Tributaries. Each of these other factors was treated as if it would limit the production of Chinook salmon in the Restoration Area. Six potential limiting factors were identified that may affect salmon populations in the San Joaquin River but not the Tributaries:

- Flows, water temperature, and water quality in the Delta and lower San Joaquin River may affect migrating adult spring-run salmon from April through June. In contrast, adult fall-run salmon migrate from September through November.
- An absence of adult holding habitat with suitable water temperatures throughout the summer may affect the spring-run population. In contrast, fall-run Chinook salmon spawn during the fall within a few weeks after arriving at their spawning beds.
- High water temperatures in the spawning reach in September and October may prevent spring-run adults from spawning, or result in high rates of egg mortality. In contrast, some fall-run spawn in late November and December, when water temperatures are usually suitable regardless of flow.
- The Tributaries have little functional floodplain habitat in their downstream reaches, whereas spring-run fry successfully rear in downstream floodplains in some Central Valley rivers, such as the Sutter Bypass in Butte Creek. It is possible that Chinook fry will rapidly grow to a smolt size in restored floodplain and wetland habitats in the downstream reaches of the Restoration Area. If true, then many would migrate into the Delta before water temperatures become unsuitable in the lower Restoration Area during May.

- The relatively long and highly degraded migration corridor for San Joaquin River smolts will result in greater exposure to stressors, such as a greater number of predators and warmer spring water temperatures, than occurs for Stanislaus and Tuolumne river fish.
- Spring-run redds may be more susceptible to redd superimposition than fall-run redds because the spring-run Chinook salmon spawn earlier than fall-run Chinook salmon.

The Tributary limiting factors analysis provided evidence supporting three hypotheses regarding the production of adult spring-run and fall-run Chinook salmon in the San Joaquin River: (1) the most critical life history stages are the rearing juveniles and outmigrating smolts, (2) critical life history stages will be strongly affected by conditions in the Restoration Area, and (3) the most important environmental factor that affects the critical life history stages is stream flow during the late-winter and spring. Population analyses for fall-run Chinook salmon in the Stanislaus and Tuolumne rivers indicate that when relatively few adults spawn in highly degraded and sparse spawning habitats, many more fry are produced than can be supported by the rearing habitat. These analyses also indicate that the quality of the juvenile rearing and migratory habitats controls the production of adult salmon in these rivers, and it is likely that the same will be true for both spring-run and fall-run Chinook salmon in the Restoration Area. Since the 1940s, the production of fall-run Chinook salmon production in the Tributaries has been highly correlated with flows from March through June when the juveniles were rearing and migrating. The same should be true for San Joaquin River spring-run salmon because both spring-run and fall-run salmon migrate at about the same time.

Conceptual Models

The conceptual models describe the potential factors limiting each life history stage of spring-run and fall-run Chinook salmon. The conceptual models assume that all actions prescribed in the Settlement, such as screening bypass channels and improving passage conditions, will be implemented. However, it is likely that there are limiting factors that have not been identified in these conceptual models.

Spring-Run Chinook Salmon

The abundance of adult spring-run Chinook salmon that return to spawn in the Restoration Area will probably be affected by numerous factors, only some of which will be under the SJRRP control. The most important potential limiting factors under the SJRRP control include the following:

- **Restoration Flows** – Truncated spring pulse flows specified in the Settlement's Restoration Hydrographs are expected to protect no more than 83 percent of the migrating smolt-sized juveniles (≥ 70 millimeters (mm) fork length) and no more than 50 percent of the migrating adults during all but wet years, assuming that the timing of the hydrographs developed under the SJRRP can be shifted by up to 4 weeks (Settlement) and that the Chinook salmon reintroduced into the San

Joaquin River will behave similarly to those that rear in the upper reaches of Butte Creek in the Sacramento River watershed. A primary concern is whether it will be possible to shift the Restoration Hydrographs into May to protect a greater number of migrating adults and juvenile salmon, while providing at least periodic floodplain inundation during the March through May rearing period, maintaining suitable water temperatures for juvenile and adult salmon (target ≤ 68 degrees Fahrenheit ($^{\circ}\text{F}$) (20 degrees celcius ($^{\circ}\text{C}$)), without exhausting the cold water pool in Millerton Lake. Extending the high flow period into May and June would probably increase smolt production and survival by improving or ameliorating a combination of factors, which include food availability, predation, disease, water temperatures, contaminants, water quality, harvest, and entrainment. However, it is also possible that many fry would migrate to the downstream reaches of the Restoration Area where they would rapidly grow to a smolt size in restored floodplain and wetland habitats prior to May. If true, pulse flows between February and April may produce a sufficient number of smolts to sustain the salmon populations.

- **Cold Water Pool** – The volume of the cold water pool in Millerton Lake may be insufficient to extend spring pulse flows into May, provide the prescribed summer and fall flow releases, and maintain suitable water temperatures for migrating fish in the spring (target $\leq 68^{\circ}\text{F}$ (20°C)), holding adult spring-run salmon during the summer (target $\leq 70^{\circ}\text{F}$ (21.1°C)), and incubating salmon eggs during the fall (target $\leq 58^{\circ}\text{F}$ (14.4°C)).
- **Habitat Degradation Within the Restoration Area** – The highly degraded channel and floodplain morphology, loss of native riparian vegetation, and the presence of exotic species may result in high rates of mortality for juvenile salmon primarily by reducing food resources, increasing water temperatures, and increasing predation. In addition, the lack of gravel recruitment from lateral and upstream sources and high flow events (e.g., 1997) have gradually scoured away the spawning gravels immediately downstream from the dam.
- **Spawner Abundance** – Legal and illegal harvest of yearling juveniles and spawning adults may reduce adult recruitment, particularly if escapements are low. In addition, conditions that result in low production of juvenile salmon would reduce the number of adult fish that return to spawn 2 – 4 years later.

Important factors outside of the SJRRP control include:

- **Stream Flow Releases Outside the Restoration Area** – Stream flow releases in the Stanislaus, Tuolumne, and Merced rivers that contribute to flows in the mainstem San Joaquin River, Delta, and San Francisco Estuary are expected to affect the survival of rearing and migrating juvenile salmon and the survival and homing ability of adult salmon.
- **Habitat Degradation Outside The Restoration Area** – The highly degraded channel and floodplain morphology, loss of native riparian vegetation, and exotic species below the confluence with the Merced River, the Delta, and San Francisco Estuary are expected to substantially reduce the survival of rearing and migrating juvenile salmon.
- **Water Quality** – Pesticides and other contaminants may substantially reduce the food resources needed by juvenile salmon within and below the Restoration Area, and to a lesser degree, result in direct mortality of juvenile salmon. In addition, poor water quality (e.g., low dissolved oxygen and high ammonia concentrations) in the mainstem channel may affect the survival of juvenile, and to a lesser degree, adult salmon.
- **Delta Exports** – Spring Delta exports at the Federal and State pumping facilities result in entrainment of juvenile fish, reduced flow in the Stockton deep-water ship channel, and reduced freshwater outflow into the ocean, all of which affect the survival of juvenile salmon.
- **Ocean Productivity** – Ocean productivity, which affects food resources and the survival of juvenile and adult salmon, is affected by upwelling, coastal currents, El Niño events, and the amount of freshwater outflow from the San Francisco Estuary.
- **Climate Changes** – Climate changes are expected to affect inland water temperatures, hydrographs (i.e., floodplain inundation), and ocean productivity, and therefore affect the survival of juvenile and adult salmon.
- **Harvest and Predation in the Ocean** – The harvest of adults and the predation of juvenile and adult salmon in the ocean affects the number of adults that return to spawn, which may affect the number of juveniles produced during the following spring.

Fall-Run Chinook Salmon

The environmental factors that are likely to affect the production of fall-run Chinook salmon are nearly identical to those that affect spring-run Chinook salmon, with a few exceptions. The primary difference is that adult fall-run do not require summer holding habitat, because they mostly migrate in October and November and then spawn shortly thereafter. The key management issues are whether the cold water pool in Millerton Lake will be sufficient to restore naturally reproducing populations of both salmon runs.

Data Needs

To effectively manage the recovery of naturally reproducing Chinook salmon populations, the following information should be considered:

- Potential source populations of spring-run Chinook salmon that have not hybridized with fall-run Chinook salmon, migrate early, tolerate high water temperatures, and are sufficiently robust to provide fish for reintroduction to the San Joaquin River.
- Effects of flow, water temperature, and Delta exports on adult fish passage.
- Distribution and quality of spawning habitat below Friant Dam.
- Effects of flow and water temperatures on the suitability of holding habitat.
- Effects of the Restoration Flows and water diversions on the size of the cold water pool in Millerton Lake, and the suitability of release temperatures for spawning habitat.
- Effects of flow and water temperature on spawning and egg incubation habitats.
- Effects of redd superimposition from fall-run Chinook salmon spawners on the production of juvenile spring-run Chinook salmon.
- Impacts of harvest on the abundance of spawners.
- Effects of flow and water temperature on the number of juvenile salmon that survive to a smolt size in the San Joaquin River.
- Effects of flow and water temperature on the survival of smolts migrating from the San Joaquin River.
- Relative effects of flow, water temperature, floodplain inundation, exotic species, contaminants, and fine sediments on food availability for juvenile salmon.

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- Relative importance of unscreened diversions, predators in captured gravel mine pits and other degraded habitats, starvation, contamination, and disease to juvenile mortality in the San Joaquin River.
- Effects of flow, water temperature, exports, the Head of the Old River Barrier, water quality and ocean-vessel traffic in the deep-water ship channel, and conditions in the Old River channel on the survival of San Joaquin River salmon smolts in the Delta.
- Effects of ocean conditions on the survival of San Joaquin River salmon smolts.
- Predicted abundance of adult spring-run and fall-run salmon in the San Joaquin River below Friant Dam using quantitative models.
- Effects of flows established under the SJRRP on fall-run salmon populations in the Merced, Tuolumne, and Stanislaus rivers.

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Appendices

Appendix A Potential Stressors, Restoration Targets, Existing Conditions, Likely Impact Levels, Certainty of Impact on Spring-Run Chinook Salmon

Appendix B Potential Stressors, Restoration Targets, Existing Conditions, Likely Impact Levels, Certainty of Impact on Fall-Run Chinook Salmon

Definitions

Alevin	The life stage of a salmon between hatching from the egg and emergence from stream gravels as a fry. Alevins are characterized by the presence of a yolk sac, which provides nutrition while the alevin develops in the redd.
Apparent Velocity	The horizontal vector of interstitial flow that is a function of permeability and hydraulic gradient.
Conceptual Model	Conceptual models are verbal or graphic depictions of how scientists believe that ecological, hydrological, and managerial systems in the San Joaquin River basin will function and respond to SJRRP actions. They will be used to help identify actions that should have a high likelihood of achieving SJRRP objectives and help identify key knowledge gaps and hypotheses that will be addressed by an adaptive management process. The conceptual models will also be used to help develop quantitative models that will facilitate the development of testable hypotheses.
D ₅₀	The median diameter of gravel at a site (e.g., spawning bed).
Diel	A daily cycle, usually encompassing 1 day and 1 night.
Escapement	The number of adult salmon that successfully “escape” the ocean fishery and return to freshwater habitats to spawn.
Fry	Fry are salmon that have absorbed their yolk sac and emerged from the redd. They typically use low velocity, shallow habitats near the river banks. In the Central Valley, fry are frequently defined as juveniles smaller than 50 millimeters in fork length.
Grilse	A precocious salmon or anadromous trout that has matured at a much smaller size and usually younger age (2-year-old) than that of the fully grown adult fish (3-year-old and older).
Limiting Factors	Stressors that significantly influence the abundance and productivity of Chinook salmon populations.
Parr	The life stage for salmon that is distinguished by its dark parr marks, and when the salmon is large enough to use mid-channel habitats. In the Central Valley, parr are defined as juveniles between 50 and 70 millimeters in fork length.

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Permeability	The ease with which water passes through gravel, depending on the composition and degree of packing of the gravel and viscosity of the water.
Restoration Area	The San Joaquin River between Friant Dam and the Merced River confluence.
Redd	A nest prepared by a female salmon in the stream bed gravel where she deposits her eggs.
Restoration Hydrographs	The schedule of flow releases from Friant Dam as prescribed in the Settlement.
Smolt	A young salmon that is undergoing physiological and morphological changes for life in seawater. Subyearling smolts are generally between 70 and 120 millimeters in fork length, whereas yearling smolts are usually larger than 180 millimeters in fork length.
Stressors	Physical, chemical, or biological perturbations to a system that adversely affect ecosystem processes, habitats, and species. Examples include altered flows, blocked passage, blocked sediment recruitment, instream habitat alteration, invasive species, contaminants, and excessive salmon harvest.

Abbreviations and Acronyms

°F	degrees Fahrenheit
°C	degrees Celsius
µg/L	microgram per liter
Ache	acetylcholinesterase
ACID	Anderson-Cottonwood Irrigation District
BKD	bacterial kidney disease
CalEPA	California Environmental Protection Agency
CALFED	CALFED Bay-Delta Program
CDFG	California Department of Fish and Game
CDPR	California Department of Pesticide Regulation
cfs	cubic feet per second
cm	centimeter
CO ₂	carbon dioxide
CVI	Central Valley Index
CVP	Central Valley Project
CVPIA	Central Valley Project Improvement Act
CVRWQCB	Central Valley Regional Water Quality Control Board
CWT	coded-wire-tag
DDT	dichloro-diphenyl-trichloroethane
Delta	Sacramento-San Joaquin Delta
DO	dissolved oxygen
DWR	Department of Water Resources
ENSO	El Niño Southern Oscillation
ESU	Evolutionarily Significant Unit
FL	fork length
H ₂ S	hydrogen sulfide
hsp	heat shock protection
IWM	instream woody material
LWD	large woody debris
MEI	Multivariate El Niño Southern Oscillation Index
mg/L	milligram per liter
MID	Modesto Irrigation District
mL/sec	millimeters per second
mm	millimeter

San Joaquin River Restoration Program

NAWQA	National Water Quality Assessment Program
NH ₃	ammonia
NMFS	National Marine Fisheries Service
NO ₂	nitrogen dioxide
NO ₃	nitrate
NPDES	National Pollutant Discharge Elimination System
OP	organophosphorus
PDO	Pacific Decadal Oscillation
PEIS/R	Program Environmental Impact Statement/Report
PKD	proliferative kidney disease
ppt	parts per thousand
RBDD	Red Bluff Diversion Dam
RM	river mile
Settlement	Stipulation of Settlement
SJRRP	San Joaquin River Restoration Project
SWP	State Water Project
SWRCB	State Water Resources Control Board
TDS	total dissolved solids
TID	Turlock Irrigation District
TKN	total kjeldahl nitrogen
USEPA	U.S. Environmental Protection Agency
USFWS	U.S. Fish and Wildlife Service
USGS	U.S. Geological Survey
VAMP	Vernalis Adaptive Management Plan

These conceptual models were prepared by the San Joaquin River Restoration Program (SJRRP) Team as a draft document in support of preparing a Program Environmental Impact Statement/Report (PEIS/R). The models define the role of, and expectations, strategy, and selection of water operations modeling for SJRRP.

The purpose for circulating this document at this time is to facilitate early coordination regarding initial concepts and approaches currently under consideration by the Program Team with the Settling Parties, the Third Parties, other stakeholders, and interested members of the public. Therefore, the content of this document may not necessarily be included in the PEIS/R. These draft conceptual models do not present findings, decisions, or policy statements of any of the Implementing Agencies. Additionally, all information presented in this document is intended to be consistent with the Settlement. To the extent inconsistencies exist, the Settlement should be the controlling document and the information in this document will be revised before its inclusion in future documents. While the Program Team is not requesting formal comments on this document, all comments received will be considered in refining the concepts and approaches described herein to the extent possible. Responses to comments will not be provided and this document will not be finalized; however, refinements will likely be reflected in subsequent SJRRP documents.

1.0 Introduction

The SJRRP is intended to implement the Stipulation of Settlement (Settlement) in *Natural Resources Defense Council et al. v. Kirk Rodger et al.*, which was approved by the U.S. District Court in October 2006.

The Settlement is based on two parallel goals:

- **Restoration Goal.** To restore and maintain fish populations in “good condition” in the mainstem of the San Joaquin River below Friant Dam to the confluence of the Merced River, including naturally reproducing and self-sustaining populations of salmon and other fish.
- **Water Management Goal.** To reduce or avoid adverse water supply impacts to all of the Friant Division long-term contractors that may result from the Interim Flows and Restoration Flows provided for in the Settlement.

The Settlement calls for a combination of channel and structural improvements along the San Joaquin River below Friant Dam, and releases of additional water from Friant Dam to the confluence of the Merced River, to be implemented in stages. Spring-run and fall-run Chinook salmon will be reintroduced to the San Joaquin River by December 31, 2012. After all Phase 1 priority construction activities have been completed by December 2013, full Restoration Flows will be released from Friant Dam as specified in Exhibit B, Restoration Hydrographs, of the Settlement.

1.1 Purpose and Document Organization

The Fishery Management Work Group has prepared this document to describe the life history requirements and the environmental factors most likely affecting the abundance of San Joaquin River spring-run and fall-run Chinook salmon (*Oncorhynchus tshawytscha*) in the Restoration Area (for the purpose of this document includes the San Joaquin River between Friant Dam and the Merced River confluence (Figure 1-1), and downstream from the Restoration Area, including the lower San Joaquin River, Sacramento-San Joaquin Delta (Delta), San Francisco Estuary, and ocean. This document includes Chinook salmon conceptual models and supporting information intended to serve as key components of the Fishery Management Plan for the SJRRP. The models assume that all restoration actions prescribed in the Settlement will be implemented.

The information herein is the result of a thorough and in-depth review of background literature, reports, and existing models describing the life history and biology of California Central Valley spring- and fall-run Chinook salmon. This document contains the following components:

- Historical population status of Chinook salmon in the San Joaquin River before and immediately after construction of Friant Dam (Chapter 2)
- Review of background literature on the basic life history and habitat requirements of Chinook salmon in the San Joaquin River basin, including the Merced, Tuolumne, and Stanislaus rivers, the greater Central Valley, and other Pacific Coast river systems, where appropriate (Chapter 3)
- Potential stressors of Chinook salmon in the San Joaquin River basin (Chapter 4 and Appendix A for spring-run salmon and Appendix B for fall-run salmon)
- Limiting factors assessment of the fall-run Chinook salmon populations in the Stanislaus and Tuolumne rivers (Chapter 5)
- Conceptual models of the mechanisms likely to influence the abundance and recovery of spring-run and fall-run Chinook salmon populations in the San Joaquin River (Chapter 6)
- Data needs (i.e., knowledge gaps) for spring- and fall-run Chinook salmon in the San Joaquin River basin (Chapter 7)
- Sources used to prepare this document (Chapter 8)

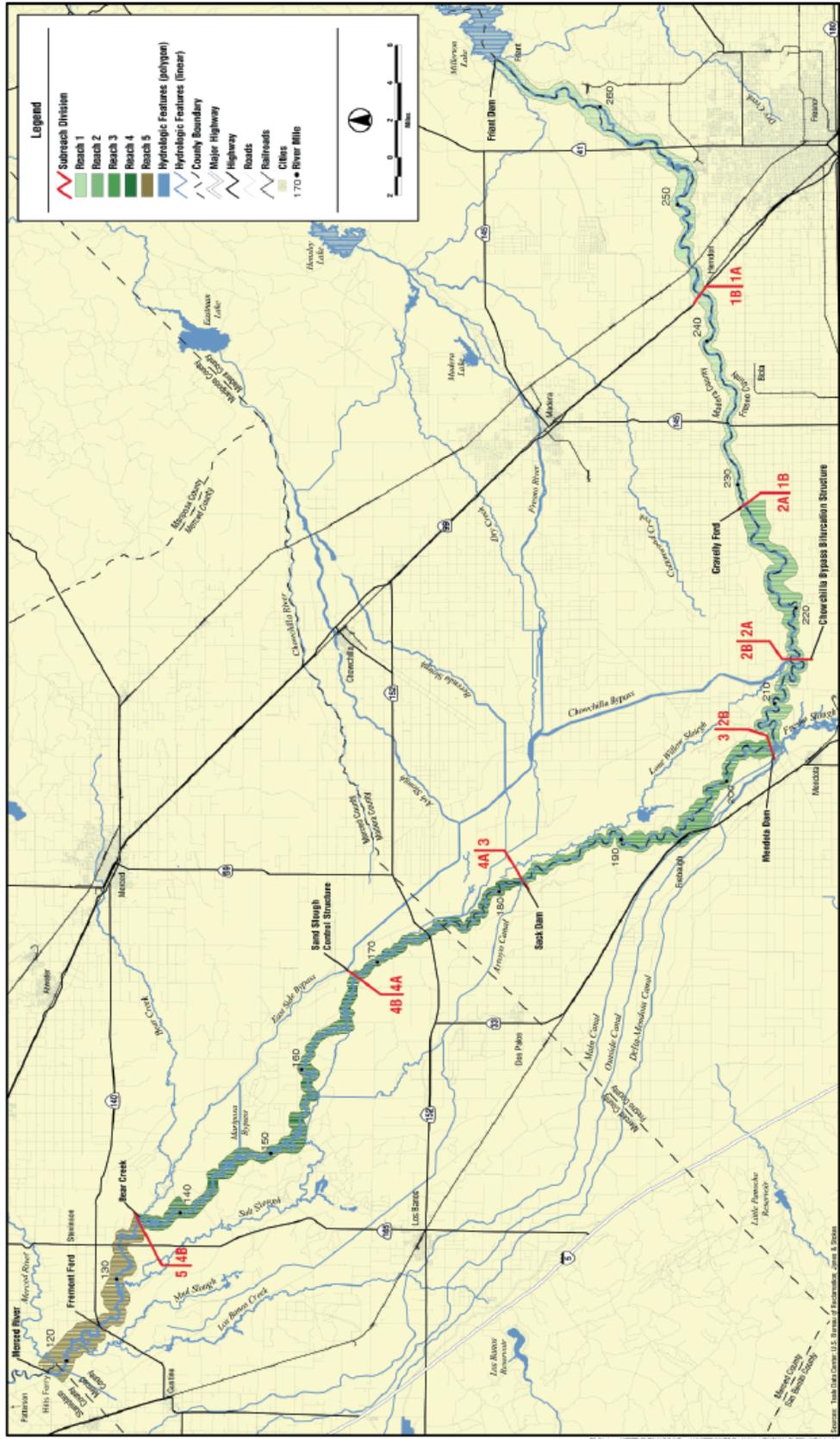


Figure 1-1.
San Joaquin River Restoration Area Showing Reach and Subreach Boundaries

The conceptual models will be used to assist in evaluating programmatic alternatives, guiding flow management, and identifying key habitat restoration needs. The models will also help identify key knowledge gaps to be addressed through a rigorous and comprehensive monitoring and adaptive management program. As part of the adaptive management process, monitoring data will be used to continually refine the conceptual models and management and restoration priorities. The conceptual models will also be used to help develop quantitative population models to refine the hypotheses to be tested under an Adaptive Management Plan. As new information becomes available and restoration actions begin, the conceptual models will be revised accordingly.

1.2 Scope

The Restoration Goal is to “restore and maintain fish populations in ‘good condition’ in the main stem of the San Joaquin River below Friant Dam to the confluence with the Merced River, including naturally producing and self-sustaining populations of salmon and other fish...” (Settlement). While many fish species will benefit from actions to meet the Restoration Goal, such as the incorporation of Restoration Flows, the emphasis of the Restoration Goal primarily is on spring-run Chinook salmon, and secondarily fall-run Chinook salmon. Therefore, the scope of this document is limited to spring- and fall-run Chinook salmon.

1.3 Coordination

This document and the conceptual models herein are based on existing salmonid models for the California Central Valley, scientific literature, and the opinions of experts working in the San Joaquin River basin. It will be further developed through extensive coordination and collaboration with various salmonid experts, restoration ecologists, modelers, as well as groups working in the basin, and Work Groups of the SJRRP. The Chinook salmon conceptual models are intended to aid in the facilitation, negotiation, and coordination of quantitative salmon population models, monitoring metrics, potential adaptive management strategies, and various regulatory review processes.

2.0 Historical Population Dynamics in the San Joaquin River

Considerable historical documentation exists regarding the presence of salmon in the San Joaquin River and its tributaries, although the identification of race is often difficult to ascertain. The first documentation of the presence of Chinook salmon in the San Joaquin River comes from Spanish explorers and missionaries of Old California (Yoshiyama et al. 2001). Large schools of adult salmon were observed in the pools near Friant during May, June, and the first part of July by the U.S. Fish Commission (Yoshiyama et al. 2001). The oral history of Native American inhabitants contains references to salmon being harvested seasonally upstream to Graveyard Meadows (Lee 1998). Salmon were also encountered in upper San Joaquin River tributaries such as the North San Joaquin River, Fine Gold Creek, Cottonwood Creek, and Whiskey Creek (Yoshiyama et al. 2001) and in valley floor tributaries such as the Chowchilla and Fresno rivers.

The California Fish and Game Commission had already noted dramatic salmon declines in the late 1800s (Yoshiyama et al. 2001). Gold mining, agricultural development, deforestation, and water development such as dam construction and flood conveyance activities adversely impacted salmon habitat. By the late 1800s and early 1900s, numerous impediments to anadromous fish passage were present in the San Joaquin River. These included Mendota Pool (RM 205) and Kerckoff Dam (approximately RM 291) After Kerckoff Dam was constructed in 1920, it permanently blocked spring-run salmon from spawning areas upstream and seasonally affected the flow in 14 miles of river with pools that provided over-summering habitat.

Clark (1929) reported that in the early 1900s there were primarily spring-run fish and relatively few fall-run. He said that the spring-run was “very good” in 1916 and 1917, “fairly good” in 1920 and 1926, but in 1928, very few salmon were seen in the river. By the 1920s, reduced autumn flows in the mainstem San Joaquin River nearly eliminated the fall run, although a small run did persist.

2.1 Spring-Run Chinook Salmon

Spring-run Chinook salmon once occupied all major river systems in California where there was access to cool reaches that would support over-summering adults. Historically, spring-run Chinook salmon were widely distributed in streams of the Sacramento-San Joaquin basin, spawning and rearing over extensive areas in the upper and middle reaches (elevations ranging 1,400 to 5,200 feet (450–1,600 meters)) of the San Joaquin, American, Yuba, Feather, Sacramento, McCloud, and Pit rivers (Myers et al. 1998). Only two evolutionarily significant units (ESU) of spring-run Chinook salmon remain in California: a Central Valley population and a Klamath-Trinity population (Moyle et al. 1995). Spring-run Chinook salmon in the San Joaquin River were extirpated in the mid-

to late 1940s following the construction of Friant Dam and diversion of water for agricultural purposes to the San Joaquin Valley.

After Friant Dam was constructed, numerous spring-run Chinook salmon returned to the river below the dam during the years when the river flowed below Sack Dam (Table 2-1) (DFG 1946, Warner 1991). Clark (1943) noted that Friant Dam first prevented upstream passage in 1942, although the dam did not begin storing water until February 21, 1944. Clark (1943) estimated that there were about 5,000 spring-run fish in a holding pool immediately below the dam in 1942, but no complete count was made that year. There was a “poor” run in 1944, when flows below Sack Dam were low and many fish were killed by “spearing” (DFG 1946). In 1945, daytime counts indicated that at least 56,000 spring-run fish passed through the Mendota Dam fish ladder or jumped over the dam (DFG 1946); it is likely that the Mendota Dam counts were low because many adults migrate at night. Flows below Sack Dam were low from spring 1948 through 1950 (Table 2-1) when only a portion of the runs were salvaged (Warner 1991). Escapement surveys were not conducted after 1950.

**Table 2-1.
Spring-Run Chinook Salmon in the San Joaquin River
from 1943 to 1950**

Year	Number Counted	Counting Method	Flows at Sack Dam
1943	35,000	Mendota Dam Ladder	4,086 cfs
1944	5,000	Mendota Dam Ladder	83 cfs
1945	> 56,000	Mendota Dam Ladder	3,066 cfs
1946	30,000	Mendota Dam Ladder	1,138 cfs
1947	6,000	Mendota Dam Ladder	98 cfs
1948	> 1,915	Hills Ferry Weir Trap	23 cfs
1950	36	Ladder from Salt Slough	3 cfs

Key:
cfs = cubic feet per second

2.2 Fall-Run Chinook Salmon

It is assumed that the San Joaquin River supported relatively few fall-run Chinook salmon after diversions began at Sack Dam sometime between 1860 and 1880.¹ Clark (1929) reported that there were few fall-run Chinook in the San Joaquin River since the early 1900s because of inadequate fall flows. During all but wet years, the river was nearly completely dewatered below Sack Dam until late November (Hatton 1940, Clark 1943), by which time it was too late for most fall-run salmon to migrate upstream in the San Joaquin River basin (Section 3.3). However, Hatton (1940) reported that in some years, fall-run fish migrated through natural sloughs and irrigation canals to the San Joaquin River above the Mendota weir. No escapement surveys were made to document the abundance of fall-run fish in the San Joaquin River.

Since the 1950s, significant numbers of San Joaquin River fall-run Chinook salmon have continued up the mainstem San Joaquin River into Salt and Mud sloughs, and their tributaries on the west side of the valley (DFG 2001). These sloughs conveyed poor quality water and had no suitable salmon spawning habitats (DFG 2001). In response to these events, the California Department of Fish and Game (DFG) has installed and operated a temporary fish barrier just upstream from the confluence with the Merced River since 1992 (DFG 2001, 2005). This barrier, called the Hills Ferry Barrier, prevented upstream access by 2,300 fall-run Chinook salmon in 1988, 322 in 1989, and 280 in 1990 (McBain and Trush 2002). Adult salmon were observed at the barrier and above the barrier between late October and mid-November in 2000 and 2004 (DFG 2001, 2005).

¹ <http://are.berkeley.edu/courses/EEP162/spring2007/documents/SJRCcasehistory.pdf>

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3.0 Life History Requirements

Central Valley Chinook salmon exhibit two general freshwater life history types, “stream-type” and “ocean-type” (Healey 1991). The evolution of stream-type and ocean-type life histories is an adaptation to the seasonal flow and temperature regimes in the rivers where Chinook salmon spawn and rear. Central Valley spring-run Chinook salmon are generally classified as stream-type because the adults migrate into mid-elevation watersheds where they spend several months while they mature sexually, and because juveniles typically spend at least 1e year rearing in fresh water. However, in the Central Valley and Oregon, spring-run juveniles typically migrate to the ocean within a few months after emerging from the gravel in the redd. In Butte Creek, California, the contribution of the subyearling life stage to adult production is approximately four times that of the yearling life stage (Ward et al. 2002). In contrast, Central Valley fall-run Chinook salmon are considered ocean-type, because the adults spawn in the lower watersheds within a few weeks of entering fresh water, and juveniles typically migrate to the ocean within a few months.

3.1 Egg Survival and Emergence

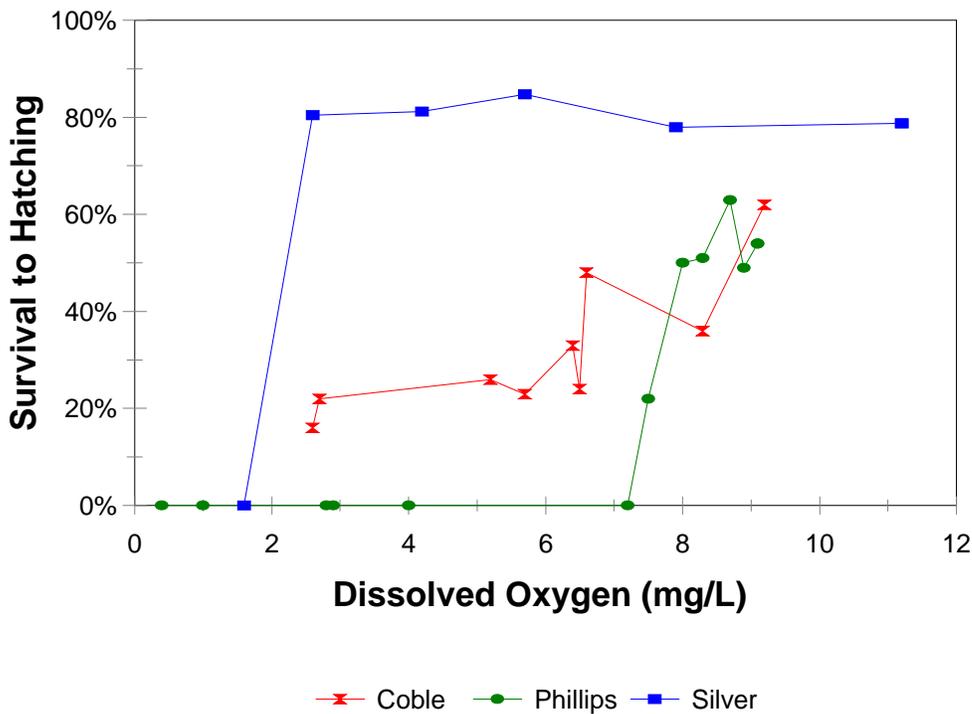
Salmon eggs incubate in nests called redds in gravel beds at depths of 12 to 18 inches under the surface of the bed until the alevins hatch in 40 to 50 days at a water temperature of 50 degrees Fahrenheit (°F) (10 degrees celcius (°C)). Normal embryo development and emergence of the fry from the gravel require suitable water temperatures, high concentrations of dissolved oxygen (DO), sufficient intragravel flow to deliver oxygenated water and flush metabolic wastes from the egg pocket, and a minimal amount of fine sediments that would otherwise block their emergence. In the Sacramento River, the egg incubation period for spring-run Chinook salmon extends from August to March (Fisher 1994, Ward and McReynolds 2001), whereas the incubation period for fall-run salmon in the San Joaquin River basin extends from late October through February.

This discussion focuses on factors that affect egg survival to the hatching stage and the factors that affect the ability of fry to emerge from the gravels. Gravel type, velocities, and specific spawning preferences of Chinook salmon are described in Section 3.5 Spawning.

3.1.1 Dissolved Oxygen and Turbidity

Numerous field and laboratory studies indicate that egg survival to hatching is greatly dependent on high concentrations of DO (Chapman 1988, Kondolf 2000). Excessive concentrations of substrate fines smaller than 1 mm in diameter are usually correlated with reduced DO (Chapman 1988, Kondolf 2000). There is a strong possibility that turbidity also affects egg survival as a result of clay-sized particles adhering to an egg’s membrane (Stuart 1953), reducing the egg’s ability to absorb DO. This effect provides a good explanation of why salmonid eggs survive at high rates under low DO

concentrations under clean laboratory conditions but not under natural settings with higher turbidity levels. When steelhead eggs were incubated on clean, porous ceramic plates under highly controlled levels of DO and flow in a laboratory, survival was high (about 80 percent) at DO levels as low as 2.5 milligrams per liter (mg/L) (Silver et al. 1963) (Figure 3-1). In contrast, a field study by Coble (1961), during which steelhead eggs were placed in plastic mesh sacks with gravel, indicates that egg survival gradually declined as DO declined from 9.2 mg/L to 2.6 mg/L (Figure 3-1). Another field study by Phillips and Campbell (1962), during which eggs were placed in perforated metal boxes with glass beads, indicates that no eggs survived at DO levels at or below 7.2 mg/L (Figure 3-1).



Sources: Silver et al. 1963 Coble 1961, and, Phillips and Campbell 1962.

Figure 3-1.
Relationship Between Dissolved Oxygen Concentration and Survival to Hatching of Steelhead Trout Eggs During Laboratory and Field Studies

Studies with other salmonid species show similar results. Eggs of chum salmon (*O. keta*; Alderdice et al. 1958), Chinook salmon (Silver et al. 1963), and coho salmon (*O. kisutch*; Shumway et al. 1964) incubated under clean laboratory conditions hatched at high rates at DO concentrations as low as 2.0 to 2.5 mg/L. Chum salmon eggs that were deposited in natural redds in an experimental stream channel with washed gravels also exhibited relatively high survival rates (50 percent) at DO levels as low as 2.5 mg/L (Koski 1975). Conversely, the survival of coho salmon eggs incubated in natural streams either in natural redds (Koski 1966) or in experimental chambers (Phillips and Campbell 1962) were reduced at DO concentrations below 9.0 mg/L and 8.3 mg/L, respectively. Although the adhesion of fines to the egg’s membranes was not evaluated in the field

studies, it is the most likely explanation for why eggs require greater concentrations of DO in natural streams than in a laboratory or in washed gravel.

The DO requirement of Chinook salmon eggs has not been accurately determined under natural field conditions. Gangmark and Bakkala (1960) studied the hatching survival of Chinook salmon eggs in artificial redds in Mill Creek, California, relative to DO concentrations. Their results were questionable, however, because individual test results were not presented and the authors referred to their earlier studies for a description of the methods (Gangmark and Broad 1955). The egg handling mortalities averaged 53 percent, possibly because the eggs were not allowed to water-harden before handling and because fungal infections caused by egg contact with the plastic mesh net bag resulted in mortality (Gangmark and Broad 1955). Furthermore, an evaluation of a portion of their raw data presented in Gangmark and Bakkala (1958) indicated that they obtained a poor relationship between survival and DO concentration, possibly due to variable rates in handling mortality among replicates. Without better direct evidence, it is assumed that Chinook salmon eggs have a relatively high DO requirement compared to coho and chum salmon and steelhead trout because Chinook salmon produce relatively large eggs. Large eggs generally require high DO concentrations because they have a relatively small ratio of surface to volume (Beacham and Murray 1985).

In addition to the effects of low DO concentrations on survival of eggs to hatching, any reduction in DO below the saturation level results in slowly developing embryos that emerge at a small size and before the absorption of all yolk (Phillips and Campbell 1962, Silver et al. 1963, Shumway et al. 1964, Mason 1969, Wells and McNeil 1970, Koski 1975). It is likely that small alevins are relatively weak and less able to emerge through sand layers covering the egg pocket than are large relatively healthy alevins incubated at high DO concentrations. Furthermore, Mason (1969) reported that small coho salmon fry subjected to low DO levels during incubation could not compete successfully with larger fry and emigrated from experimental channels. Chapman (1988) suggested that any reduction in DO levels from saturation probably reduces survival to emergence or postemergent survival.

3.1.2 Intragravel Flow

Intragravel flow is also correlated with egg survival. Intragravel flow is measured as either permeability or apparent velocity during egg survival studies. Permeability is the ease with which water passes through gravel, and depends on the composition and degree of packing of the gravel and viscosity of the water (Pollard 1955). Apparent velocity is the horizontal vector of interstitial flow and is a function of permeability and hydraulic gradient (Pollard 1955, Freeze and Cherry 1979). It is measured as the rate of flow through a standpipe, which is called apparent yield, divided by the porosity of the surrounding gravel. The actual velocity of flow through interstitial spaces, which is called the true or pore velocity, is faster than the apparent velocity because flow travels around substrate particles whereas apparent velocity assumes that the flow path is linear. Laboratory studies, such as Silver et al. (1963) that incubate eggs without a gravel medium, measure true velocity, whereas all field studies measure apparent velocity with standpipes.

The survival of steelhead and coho salmon egg to hatching in natural streams has been correlated with apparent velocity but not as strongly as with DO concentration, whereas there were no correlations with permeability (Coble 1961, Phillips and Campbell 1962). The size of coho salmon and steelhead embryos at hatching was reduced at low velocities, regardless of DO concentration in the lab (Shumway et al. 1964), whereas Chinook salmon and steelhead egg survival was not correlated with true velocity under the same laboratory conditions (Silver et al. 1963). Koski (1966) reported that survival to emergence of coho salmon eggs in natural redds was not correlated with a permeability index (milliliters per second (mL/sec)). Sowden and Power (1985) reported that rainbow trout egg survival in a groundwater-fed stream was strongly correlated with DO and apparent velocity, but not with the percentage of fines less than 2 mm, the geometric-mean particle size, or the fredle index.

Although egg survival and apparent velocity have been highly correlated in several studies, there is no consistent critical apparent velocity for egg survival, possibly because of the influence of different levels of DO and the adhesion of clay-sized particles to the egg's membrane among the studies. The results of five studies are listed below as evidence that the critical apparent velocity necessary for high rates of egg survival can vary from 0.65 feet per hour (ft/hr) (20 centimeters per hour (cm/hr)) to 50.9 ft/hr (1,550 cm/hr) depending on the DO concentration.

- Gangmark and Bakkala (1960) reported that the mean survival to hatching for newly fertilized Chinook salmon eggs planted in 220 artificial redds in Mill Creek, California exceeded 87 percent where apparent velocity was at least 1.5 ft/hr and DO exceeded 5 mg/L. Mean survival was 67 percent at 14 sites where apparent velocity ranged between 0.5 and 1.0 ft/hr during the same study. However, the results of their study are questionable because individual test results were not presented and the methods were not described (see the above discussion on egg DO requirements).
- Coble (1961) reported that steelhead egg survival to hatching was high, 48 to 62 percent, at artificial redds with mean apparent velocities that exceeded 1.52 ft/hr (46.5 cm/hr) and mean DO levels greater than 6.4 mg/L.
- Phillips and Campbell (1962) reported that steelhead egg survival was high, 49 to 63 percent, in artificial redds with apparent velocities that exceeded 0.65 ft/hr (20 cm/hr) and mean DO levels that exceeded 8.3 mg/L.
- Reiser and White (1988) reported that Chinook salmon egg survival to hatching was highly correlated ($r = 0.797$) with apparent velocity and the percentage of two size classes of substrate fines during laboratory tests that maintained DO levels between 6.2 and 7.7 mg/L. These results suggest that at low DO levels tested, apparent velocity less than 50.9 ft/hr (1,550 cm/hr) resulted in reduced egg survival. They also reported that fines less than 0.84 mm in diameter affected survival to a much greater degree than did sediment between 0.84 and 4.6 mm in diameter, presumably due to greater influence of intragravel flow.

- Deverall et al. (1993) reported apparent velocities in natural Chinook salmon redds exceeded 16.4 ft/hr (500 cm/hr) at 45 of 49 redds in the Waitaki River, New Zealand and that egg survival to hatching was between 75 and 98 percent at three redds where apparent velocity ranged between 6.56 ft/hr (200 cm/hr) and 9.84 ft/hr (300 cm/hr) and DO levels were near saturation.

3.1.3 Water Temperature

A review of numerous studies suggests that 42 to 55°F (5.5 to 12.8°C) is the optimum temperature range for incubating Chinook salmon (Donaldson 1955, Combs and Burrows 1957, Combs 1965, Eddy 1972, Bell 1973, Healey 1979, Reiser and Bjornn 1979, Garling and Masterson 1985). Results of laboratory investigations conducted by Seymour (1956) showed a rapid increase in Chinook salmon egg mortality as temperatures increased above 57°F (13.9°C). A more recent thermal tolerance study of Sacramento River fall-run Chinook salmon eggs found that egg mortality began to occur as water temperature rose above 54°F (12.2°C) but was insignificant at temperatures from 52 to 56°F (11.1 to 13.3°C) (USFWS 1999). Alderdice and Velsen (1978), who reviewed the available literature, estimated that the upper temperature limit for 50 percent mortality of Chinook salmon eggs was near 61°F (16°C). Healey (1991) suggested that although Chinook normally begin spawning in late summer when temperatures are near 61°F (16°C), temperatures are falling rapidly at this time of year and the eggs are probably not exposed to near lethal temperatures for long. On a previous study, Healey (1979) found that water temperatures higher than 57°F (13.9°C) caused greater than 82 percent mortality of Chinook salmon eggs in the Sacramento River. These eggs appear to be no more tolerant of high water temperatures than the more northern California races. Myrick and Cech (2001) likewise concluded that there appears to be very little variation in thermal tolerance of Chinook salmon eggs among geographic races.

Chinook salmon egg survival also declines at water temperatures below 42°F (5.6°C) and mortality is about 100 percent at a constant temperature of 35°F (1.7°C) (Leitritz 1959). Eggs can tolerate temperatures below 42°F (5.6°C) for about 6 days without mortality (Leitritz 1959). Gangmark and Bakkala (1958) reported water temperatures between 34 and 36.5°F (1.1 and 2.5°C) in January 1957 in artificial redds with planted eggs in Mill Creek, the North Fork of Mill Creek, and the Sacramento River. The duration of the cold temperatures was not reported but there was no indication that egg survival rates were affected.

3.1.4 Emergence

After hatching, alevins remain buried in the gravel for an additional period of development during which time nutrition is provided by absorption of the yolk sac. After yolk sac absorption by the alevins has been completed, fry begin the process of emerging from the gravel. In the Sacramento River basin, spring-run Chinook salmon alevins remain in the gravel for 2 to 3 weeks after hatching and emerge from the gravels into the water column from November to March (Fisher 1994, Ward and McReynolds 2001). In the Tuolumne River, the period of fall-run Chinook salmon alevin development has been estimated to last between 35 and 55 days (mean 47 days) at 50 to 55°F (10 to 13°C), based on the timing from redd completion to peak emergence at five fall-run Chinook salmon redds monitored in fall 1988 (TID and MID 1992).

3.2 Juvenile Rearing and Migration

Upon emergence, salmon fry swim or are displaced downstream (Healey 1991). Downstream movement of fry primarily occurs at night along the margins of the river. After this initial dispersal, fry may continue downstream to the estuary and rear, or may take up residence in the stream for a period of time from weeks to a year (Healey 1991). Although juvenile spring-run Chinook salmon primarily exhibit a stream-type life history pattern wherein they remain in freshwater until the spring following their emergence from the gravel in the redd, they are also known to migrate from spawning areas in their first year. Populations in Oregon (Healey 1991) and California (e.g., Butte Creek) primarily migrate to the ocean as subyearling smolts within a few months after emergence. The duration of juvenile freshwater residency may be influenced by water temperature and river outflow. Nicholas and Hankin (1989) found that the duration of freshwater rearing in Oregon coastal streams is tied to water temperatures, with juvenile Chinook salmon remaining longer in rivers with cool water temperatures. Moyle (2002) suggests that the propensity for Chinook salmon fry and smolts to emigrate to the ocean increases as high flows cause reduced water temperatures and increased turbidity.

River-rearing salmon fry occupy low velocity, shallow areas near stream margins, including backwater eddies and areas associated with bank cover such as large woody debris or large substrate (Lister and Genoe 1970, Everest and Chapman 1972, McCain 1992). Juvenile Chinook salmon often seek refuge in low velocity habitats where they can rest and feed on drifting invertebrates with minimum expenditure of energy. Because of the energetic demands of both retaining position within the water column and obtaining prey items, as well as the metabolic demands on ectotherms as water temperatures increase, fish feeding and growth in rivers depend on a number of factors working in concert. Energy required to maintain position within the water column is generally a function of body size (Chapman and Bjornn 1969, Everest and Chapman 1972). For example, small fish and newly emerged fry typically inhabit slower water habitats, often found at the margins of mainstem channels, backwaters, or side channels. Larger fish typically move into swifter flowing habitats, where larger prey are usually available (Lister and Genoe 1970). This shift is also energetically more economical, since larger fish would require more prey items, and capturing one prey item is energetically more efficient than capturing many.

Juvenile salmonids larger than 2 inches (50 mm) in length in the Sacramento-San Joaquin system also rear on seasonally inundated floodplains. Sommer et al. (2001) found higher growth and survival rates of Chinook salmon juveniles that reared on the Yolo Bypass than in the mainstem Sacramento River, and Moyle (2000) observed similar results on the Cosumnes River floodplain. Sommer et al. (2001) found that drifting invertebrates, the primary prey of juvenile salmonids, were more abundant on the inundated Yolo Bypass floodplain than in the adjacent Sacramento River. Bioenergetic modeling suggested that increased prey availability on the Yolo Bypass floodplain was sufficient to offset increased metabolic demands from higher water temperatures (9°F (5°C)) higher than in the mainstem). Gladden and Smock (1990) estimated that annual invertebrate production on two Virginia floodplains exceeded river production by one to two orders of magnitude. In the Virginia study, annual production on the floodplain continuously

inundated for 9 months was 3.5 times greater than on the floodplain inundated only occasionally during storms (Gladden and Smock 1990).

Sommer et al. (2001) suggested that the well-drained topography of the Yolo Bypass may help reduce stranding risks when floodwaters recede. Most floodplain stranding occurs in pits or behind structures (e.g., levees or berms) that impede drainage (Moyle et al. 2005). Additionally, research in the Cosumnes River (Moyle et al. 2005) and Tuolumne River (Stillwater Sciences 2007) suggests that flow-through of water on inundated floodplains appeared to be more important for providing suitable habitat for Chinook salmon and other native fish species than the duration of inundation or other physical habitat characteristics. Thus, configuration of restored floodplains to promote active flow-through of river water (i.e., creation of conveyance floodplains) would likely maximize habitat value for juvenile Chinook salmon.

Considering the historical extent of floodplain inundation in the San Joaquin system, and tule (*Scirpus acutus*) marsh habitat along the San Joaquin River before land development, it is possible that juvenile Chinook salmon and steelhead reared on inundated floodplains in the San Joaquin River in Reaches 2 through 5. These downstream reaches were inundated for a good portion of the year in normal and wetter years, providing suitable water temperatures for juvenile rearing from January to at least June or July in most years, and perhaps extending into August in wetter years. As snowmelt runoff declined, and ambient temperatures increased, water temperatures in slow-moving sloughs and off-channel areas probably increased rapidly. The extent to which juvenile salmonids would have used the extensive tule marshes and sloughs historically found in Reaches 2, 3, 4, and 5 is unknown.

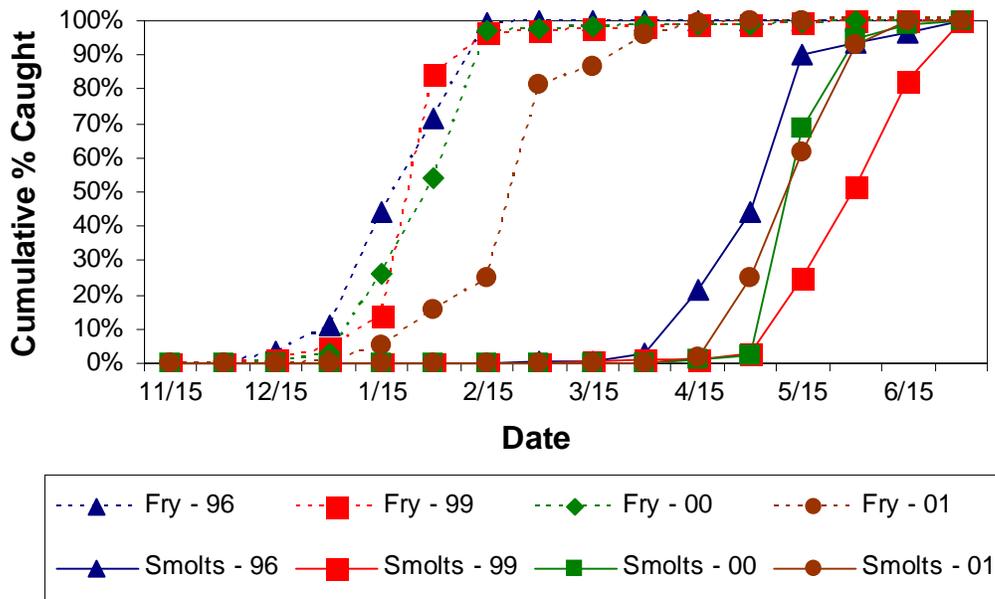
The quality of juvenile rearing habitat is highly dependent on riparian vegetation. Riparian vegetation provides shading that lowers river temperatures, provides allochthonous organic matter that drives the salmon's food web, contributes woody debris for aquatic habitat complexity, bank stability through root systems, and filtration of sediments and nutrients in storm runoff (Helfield and Naiman 2001).

3.2.1 Migration Timing

Juvenile salmon in the Central Valley move downstream at all stages of their development: most as newly emerged fry dispersing to downstream rearing habitats and others that migrate toward the ocean as they undergo smoltification. Smoltification is the physiological process that increases salinity tolerance and preference, endocrine activity, and gill Na^+ - K^+ ATPase activity. It usually begins in late March when the juveniles reach a fork length between 70 and 100 mm; however, a few fish delay smoltification until they are about 12 months old (yearlings) when they reach a fork length between 120 and 230 mm. Environmental factors, such as stream flow, water temperature, photoperiod, lunar phasing, and pollution can affect the onset of smoltification (Rich and Loudermilk 1991).

Rotary screw trap studies at the Parrott-Phelan Diversion Dam in Butte Creek probably provide the best available information on the migratory behavior of a natural spring-run salmon population in the Central Valley, because hatchery fish are not planted in Butte Creek and the fall-run salmon do not spawn above the study site. In Butte Creek, at least

95 percent of the juvenile spring-run salmon migrate as fry from the spawning areas upstream from Parrott-Phelan Diversion Dam into the Sutter Bypass where they rapidly grow (0.5 to 0.7 mm/day) to a subyearling smolt size (60 to 100 mm fork length (FL)) (Ward et al. 2002). The Butte Creek fry primarily disperse downstream from mid-December through February (Figure 3-2) whereas the subyearling smolts primarily migrate between late-March and mid-June (Figure 3-2). Spring-run yearlings in Butte Creek migrate from September through March (Hill and Webber 1999, Ward and McReynolds 2001, Ward et al. 2002). Juvenile emigration patterns in Mill and Deer creeks are very similar to patterns observed in Butte Creek, with the exception that Mill Creek and Deer Creek juveniles typically exhibit a later young-of-the-year migration and an earlier yearling migration (Lindley et al. 2004).



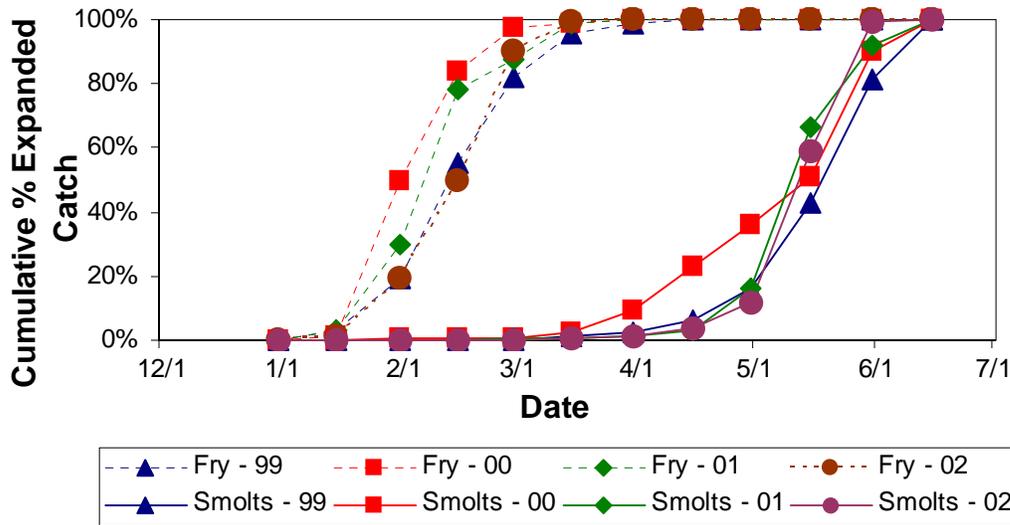
Sources: Hill and Webber 1999, Ward and McReynolds 2001, Ward et al. 2002.

Notes:

1. The data are plotted in 2-week intervals relative to the last date of capture in each interval.
2. Fry < 50 mm fork length.
3. Subyearling smolt \geq 70 mm fork length

Figure 3-2.
Cumulative Percent of Spring-Run Chinook Salmon Fry and Subyearling Smolt-Sized Fish Caught with Rotary Screw Trap at Parrott-Phelan Diversion Dam on Butte Creek, California, in 1996, 1999, 2000, and 2001

Fall-run salmon fry disperse downstream from early January through mid-March, whereas the smolts primarily migrate between late March and mid-June in the Stanislaus River (Figure 3-3), which is nearly identical to the timing of spring-run smolt outmigration in Butte Creek. Fall-run yearlings are caught during all months that the rotary screw traps are operating at Oakdale on the Stanislaus River; this occurs from December through June, regardless of flow (Fishbio Environmental, LLC, unpublished data).



Sources: Fishbio Environmental, LLC, unpublished data.

Notes:

1. The data are plotted in 2-week intervals relative to the last date of capture in each interval.
2. Fry < 50 mm fork length.
3. Smolt ≥ 70 mm fork length.

Figure 3-3.
Cumulative Percent of Expanded Number of Fall-Run Chinook Salmon Fry and Smolt-Sized Fish Passing Rotary Screw Trap at Oakdale on the Stanislaus River, California, in 1999, 2000, 2001, and 2002

3.2.2 Delta and Estuary Rearing

In many systems, an important life history strategy of juvenile salmonids is to leave freshwater soon after emergence and take up residence in tidally functioning estuaries. While this is a common life history strategy among salmon on the Pacific Coast, fry often appear most abundant 2 to 3 months earlier in the Delta than in other Pacific Coast estuaries, perhaps in response to the warmer temperatures in the Delta (Healey 1980, Kjelson et al. 1982). Juvenile salmon less than 70 mm fork length (FL) are abundant in the Delta from February to April (MacFarlane and Norton 2002). Work in other West Coast estuaries indicates estuarine rearing by fry is an important and critical life stage of salmon development (Levy and Northcote 1981). Fyke trapping and trawling studies conducted by the U.S. Fish and Wildlife Service (USFWS) in the Sacramento River and in the Delta suggest small juvenile Chinook salmon use the shoreline and larger juveniles typically use the center of the channel (USFWS 1994a). Other studies along the Pacific Coast also indicate a preference for nearshore areas by less mature juvenile salmon (Dauble et al. 1989, Healey 1991). The diet of fry and juvenile Chinook salmon in the San Francisco Estuary consists of dipterans, cladocerans, copepods, and amphipods (Kjelson et al. 1982). Thus, the nearshore habitats in the Delta and San Francisco Bay are probably valuable to juvenile salmon for rearing purposes, whereas the main deepwater channels are used for migratory purposes.

Numerous spring-run Chinook fry from the San Joaquin River entered the estuary before and shortly after Friant Dam was constructed. Prior to construction of Friant Dam, seasonal downstream migrations of juvenile Chinook salmon occurred following heavy outflow events (Hallock and Van Woert 1959). Peak migration at Mendota was between late January and June, peaking in February 1944. Additional sampling at Mossdale also found the greatest numbers emigrating during January and February (Hallock and Van Woert 1959). Juveniles captured at Mendota before 1949 were “for all practical purposes the progeny of spring-run Chinook salmon adults only, since very few fall-run fish spawned in the upper San Joaquin” (Hallock and Van Woert 1959). Based on this information, it is highly likely that fry-sized spring-run Chinook salmon from the San Joaquin River basin historically used the lower San Joaquin River, Delta, and San Francisco Bay for rearing purposes.

3.2.3 Smoltification and Estuary Presence

Juvenile salmon undergo complex physiological changes, called smoltification, in preparation for their life in saltwater (summarized in Quinn 2005). These include changes in osmoregulation (salt balance), body shape and color, energy storage, and migratory behavior. A change in osmoregulation is critical because in the freshwater environment, juvenile salmon must keep from losing their essential electrolytes (salts that regulate body functions) and absorbing too much water. To do this, they minimize water intake, excrete dilute urine, and actively acquire salts with their gills. In saltwater, which is saltier than their body fluids, they drink, but must excrete salts from their gills and produce concentrated urine. The smolting process is metabolically demanding and juveniles release hormones that trigger the utilization of their energy reserves. One of these hormones, cortisol, also inhibits their immune system, making smolts more vulnerable to disease and other stress. Also, the juveniles change their appearance from

the green-brown backs and vertical parr marks that camouflage them in streams to the blue-green backs, silver sides, and white bellies that are typical of pelagic marine fishes. The smolting process is triggered by a combination of conditions, including body size, rate of growth, increasing day length, and increasing water temperatures. There is a smoltification window during spring, after which slow growing, small individuals lose their ability to smoltify.

As Chinook salmon begin smoltification, they prefer to rear further downstream where ambient salinity is up to 1.5 to 2.5 parts per thousand (ppt) (Healy 1980, Levy and Northcote 1981). Smolts enter the San Francisco Estuary primarily in May and June (MacFarlane and Norton 2002) where they spend days to months completing the smoltification process in preparation for ocean entry and feeding (Independent Scientific Group 1996). Within the estuarine habitat, juvenile Chinook salmon movements are dictated by the tidal cycles, following the rising tide into shallow water habitats from the deeper main channels, and returning to the main channels when the tide recedes (Levy and Northcote 1981, Healey 1991). Kjelson et al. (1982) reported that juvenile Chinook salmon demonstrated a diel migration pattern, orienting themselves to nearshore cover and structure during the day, but moving into more open, offshore waters at night. The fish also distributed themselves vertically in relation to ambient light. During the night, juveniles were distributed randomly in the water column, but would school during the day into the upper 3 meters of the water column.

Decaying marsh vegetation forms the basis of the juvenile Chinook salmon's food web in the Columbia River (Bottom 2007). Juveniles, 40 to 60 mm fork length, primarily used shallow, nearshore, and wetland habitats. They fed on insects (adult dipterans), amphipods (*Corophium salmonis*, *C. spinicome*), and water fleas (Cladocera) that were produced in wetland habitats. Juveniles spent an average of 73 days (10 to 219) in the Columbia River estuary growing an average of 0.5 mm per day in 2004 (Bottom 2007).

In the San Francisco Estuary, insects and crustaceans dominate the diet of juvenile Chinook salmon (Kjelson et al. 1982, MacFarlane and Norton 2002). Larval fish become increasingly important in the diet as juvenile Chinook salmon approach and enter the ocean (MacFarlane and Norton 2002). Juvenile Chinook salmon spent an average of about 40 days migrating through the Delta to the mouth of San Francisco Bay in spring 1997, but grew little in length or weight until they reached the Gulf of the Farallon Islands (MacFarlane and Norton 2002). After passing through Suisun Bay, juvenile Chinook primarily fed on the hemipteran *Hesperocorixa* sp., the calanoid copepod *Eucalanus californicus*, the mysid *Acanthomysis* sp., fish larvae, and other insects (MacFarlane and Norton 2002). In San Pablo Bay, marine crustaceans in the order Cumacea were the dominant prey of juvenile salmon. In the Central Bay, the juvenile Chinook salmon fed on insects, fish larvae, *Ampelisca abdita* (a gammaridean amphipod), and cumaceans (MacFarlane and Norton 2002). Based on the mainly ocean-type life history observed (i.e., fall-run Chinook salmon), MacFarlane and Norton (2002) concluded that unlike other salmonid populations in the Pacific Northwest, Central Valley Chinook salmon show relatively little estuarine dependence and may benefit from expedited ocean entry. It is possible that the absence of extensive marsh habitats outside Suisun and San Pablo bays, and the introduction of exotic species of zooplankton, limit

important food resources in the San Francisco Estuary that are present in other Pacific Northwest estuaries (MacFarlane and Norton 2001).

3.2.4 Ocean Phase

All Chinook salmon utilize the ocean to achieve maximum growth, although this growth is a tradeoff with high mortality, and all races of salmon deal with this tradeoff differently (Pearcy 1992). Central Valley Chinook salmon typically spend between 2 and 4 years at sea (Mesick and Marston 2007a). Most mortality experienced by salmonids during the marine phase occurs soon after ocean entry (Pearcy 1992, Mantua et al. 1997). Typically, salmon time their ocean entry to minimize predation and maximize growth; however, Chinook salmon appear to use the “bet-hedging” strategy, adopting diverse ocean entry patterns that do not correspond to major ocean events (Pearcy 1992).

Because of the small size of juveniles entering the ocean, their movements are greatly influenced by currents during this time. Most head in a northerly direction along the coastal shelf during the first year of their life (Pearcy 1992). Williams (2006) notes that in the summer, juveniles are found in slow eddies at either side of the Golden Gate, but that their distribution shifts north beyond Point Reyes later in the fall. Knowledge of California salmon life in the ocean is extremely limited. MacFarlane and Norton (2002) were the first to describe their physiology and feeding behavior in coastal waters of central California. They compared the feeding rates and condition of fall-run Chinook salmon in the lower end of the Delta (Chippis Island), at the Golden Gate Bridge (representing the end of the San Francisco Bay), and in the Gulf of the Farallones. Results indicated that feeding and growth were reduced in the estuary, but increased rapidly in the coastal shelf in the Gulf of the Farallones (MacFarlane and Norton 2002). Fish larvae were the most important prey of juvenile Chinook salmon in the coastal waters of the Gulf of the Farallones (MacFarlane and Norton 2002). Euphausiids and decapod early life stages were also consumed in significant numbers.

Maturing Chinook salmon are abundant in coastal waters ranging from southeastern Alaska to California and their distribution appears to be related to their life history type (stream-type or ocean-type), race, and physical factors such as currents and temperature (Healey 1991, Williams 2006). Unfortunately, little information exists on the geographic distribution of Chinook salmon in the sea. Williams (2006) reported coded-wire-tag recoveries by fisheries management area from the Regional Mark Information System database. Results indicated that Central Valley Chinook salmon are primarily distributed between British Columbia and Monterey, California, with the highest percentages found off the coasts near San Francisco and Monterey.

Subadults feed on northern anchovy, juvenile rockfish, euphausiids, Pacific herring, osmerids, and crab megalopae along the Pacific Coast (Hunt et al. 1999). Northern anchovies and rockfish appear to be the most important prey items off the San Francisco coast (Hunt et al. 1999). It is likely that prey items change seasonally, and salmon take advantage of such changes with opportunistic feeding (Williams 2006).

3.3 Adult Migration

As Chinook salmon near sexual maturity, they attempt to return to their natal stream to spawn. Adults, particularly the stream-type fish that migrate long distances in the ocean to feed, use geomagnetic orientation in ocean and coastal waters to locate the mouth of their natal stream, where they switch to olfactory clues (Quinn 1990). The mechanism of compass orientation and the transition from compass orientation in coastal waters and estuaries to olfactory-based upriver homing appear to be very complicated and not well understood (Quinn 1990). Furthermore, ocean-type populations of Pacific salmon, such as the fall-run Chinook populations in the San Joaquin River tributaries, may not have a well-developed means of navigation by compass orientation since they do not migrate far from the coast to feed.

Adult Pacific salmon primarily rely on olfactory cues to guide their upriver migration to their natal stream, although other factors may be involved (Quinn 1990). It is generally believed that as juveniles rear and migrate downriver, they imprint on the olfactory cues at every major confluence and retrace the sequence as adults when they return to spawn (Harden-Jones 1968, Quinn et al. 1989, Quinn 1990). Few adult coho (Wisby and Hasler 1954) and Chinook salmon (Groves et al. 1968) that had their olfactory pits plugged (to prevent them from sensing waterborne odors) were able to home to their natal stream. Most (67 percent and 89 percent) of the control fish in those studies were able to home to their natal stream. Blinded fish were able to home more successfully than were fish with occluded olfactory pits. Experiments have also shown that juvenile coho salmon exposed to artificial waterborne odors while they were reared in hatcheries homed to waters that contained those artificial odors (Cooper et al. 1976, Johnsen and Hasler 1980, Brannon and Quinn 1990, Dittman et al. 1994, Dittman et al. 1996). Normal homing rates for Chinook salmon are not precisely known, but probably range between 84 percent and 99 percent, which are the homing rates calculated for hatchery-reared Chinook salmon in New Zealand (Unwin and Quinn 1993) and the Cowlitz River Hatchery, Washington (Quinn and Fresh 1984).

There is contradictory evidence that hereditary factors may also influence homing behavior. Bams (1976) and McIsaac and Quinn (1988) provided proof that a high proportion of displaced Chinook salmon offspring homed to their ancestral spawning area even though the juvenile fish were never exposed to their ancestral waters. However, Donaldson and Allen (1957) provided evidence that coho juveniles relocated to two different locations prior to smolting would home to their release sites and not to their original hatchery site. The scent from siblings (population-specific odors) did not affect adult coho salmon homing behavior in Lake Washington (Brannon and Quinn 1990), and no other mechanism to account for a hereditary factor has been discovered.

When adult Pacific salmon do not return to their natal stream, they appear to select a new river for spawning based on the magnitude of stream flow. Two field studies conducted by Quinn and Fresh (1984) in Washington and Unwin and Quinn (1993) in New Zealand determined that adult Chinook salmon strays selected rivers with the highest stream flow. An experimental study conducted by Wisby and Hasler (1954) also showed that when the

scent of the fishes' natal river was not present, coho salmon moved into the arm of a forked channel with the greatest flow.

3.3.1 San Francisco Bay and Sacramento-San Joaquin Delta

Runs are designated on the basis of adult migration timing as the fish enter San Francisco Bay; however, runs also differ in the degree of maturation at the time of river entry, thermal regime and flow characteristics of their spawning site, and the actual time of spawning (Myers et al. 1998). Spring-run Chinook salmon migrate upstream during the spring before they have fully reached sexual maturity, whereas fall-run Chinook salmon are sexually mature when they enter fresh water between June and December (Moyle 2002) and spawn shortly thereafter. Adult spring-run Chinook salmon begin entering San Francisco Bay in late January and early February (DFG 1998). Adult San Joaquin River basin fall-run Chinook salmon have been collected in the Delta near Prisoners Point primarily during September and October (Hallock et al. 1970).

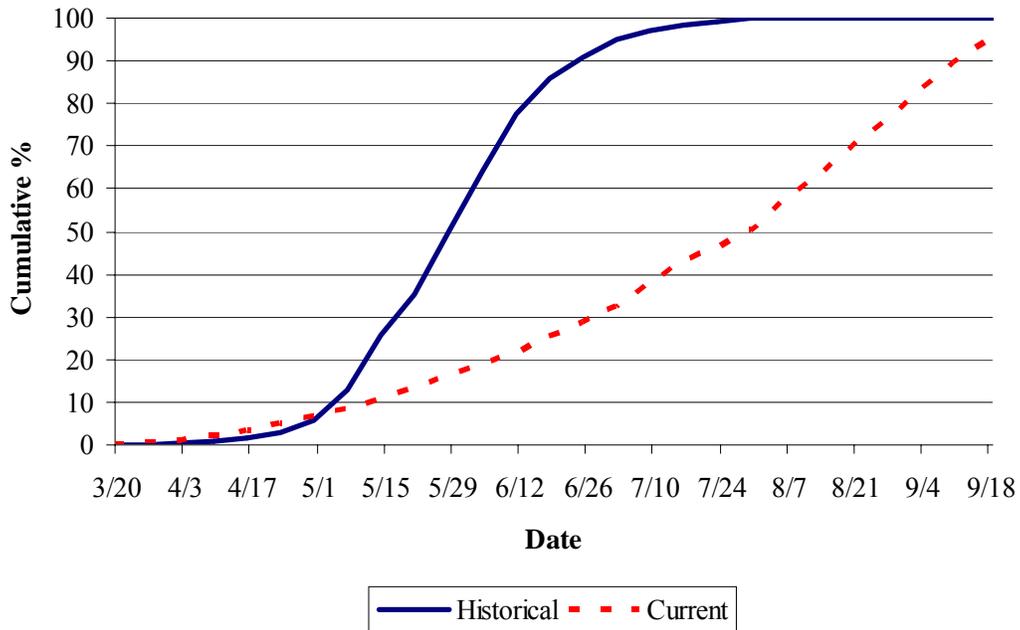
As adult Chinook salmon migrate through the Delta, they cease feeding (Higgs et al. 1995). Merkel (1957) found a high percentage of empty stomachs of salmon captured in North San Francisco Bay, particularly during the beginning of the spring-run Chinook salmon migration period (February and March). Merkel found no Chinook salmon in North San Francisco Bay with immature gonads, and presumed that samples from the San Francisco Bay were farther along in sexual maturity as opposed to offshore samples and as a result, the fish had ceased feeding, unlike the offshore samples (Merkel 1957). Recent study continues to verify the cessation of feeding on estuary entrance and gonadal development (DFG 1998).

Adult Chinook salmon inserted with sonic tags and tracked throughout the Delta and lower Sacramento and San Joaquin rivers were observed exhibiting substantial upstream and downstream movement in a random fashion while migrating upstream (CALFED 2001) several days at a time. Adult salmonids migrating upstream are assumed to make greater use of pool and mid-channel habitat than channel margins, particularly larger salmon such as Chinook (Hughes 2004).

3.3.2 River

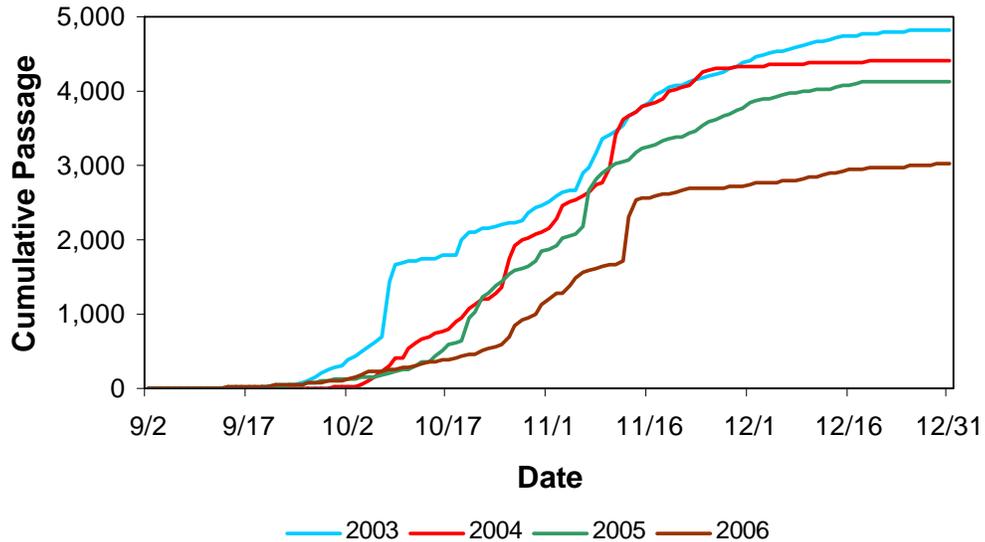
In the Sacramento River watershed (the closest population of spring-run Chinook salmon to the San Joaquin River), adult spring-run Chinook salmon historically returned to fresh water between late March and early July (Figure 3-4) (DFG 1998). The spring-run populations in Mill (Johnson et al. 2006) and Butte creeks (McReynolds 2005, personal communication) still exhibit this historical migration timing. However since 1970, most spring-run salmon in the Sacramento River at Red Bluff Diversion Dam (RBDD), the Feather River, and Yuba River migrate during the summer (Figure 3-4) (DFG 1998).

Weir counts in the Stanislaus River suggest that adult fall-run Chinook salmon in the San Joaquin River basin typically migrate into the upper rivers between late September and mid-November (Figure 3-5) (S.P. Cramer and Associates 2004, 2005; Cramer Fish Sciences 2006, 2007).



Source: DFG 1998.

Figure 3-4.
Timing of Adult Spring-Run Chinook Salmon Migrating Past Red Bluff Diversion Dam from 1970 to 1988 (Current) and Composite Data from Mill and Deer Creeks, Feather River, and Upper Sacramento River Before Construction of Shasta Dam (Historical)



Sources: S.P. Cramer and Associates 2004, 2005; Cramer Fish Sciences 2006, 2007.

Figure 3-5.
Cumulative Number of Adult Fall-Run Chinook Salmon Counted in Stanislaus

River near Riverbank (RM 31.4) with a Weir and Vaki RiverWatcher Digital Infrared Recording System from 2003 to 2006

3.4 Adult Holding

When adult spring-run Chinook salmon begin their migration to their natal streams, they are sexually immature, unable to spawn. After they arrive in their natal streams in the spring, they hold in deep pools through the summer, conserving energy until the fall when their gonads ripen and they spawn. In the Sacramento River system, adult spring-run Chinook salmon typically hold between April and July (Yoshiyama et al. 1998) or September (Vogel and Marine 1991) and then begin spawning in late August at the higher elevations, and in October at the lower elevations (DFG 1998). While holding during the summer, spring-run adults minimize their activity, which is thought to lower metabolic rates and therefore conserve energy for eventual reproductive activities (NRC 1992, as cited in Bell 1986).

Spring-run Chinook salmon adults generally require deep pools with relatively slow water velocities as holding habitat. Deep pools remain cooler during warm summer months, and provide refuge from avian and terrestrial predators. Instream cover (e.g., undercut banks, overhanging vegetation, boulders, large wood, and surface turbulence) also provides refuge from predation. For spring-run Chinook salmon in the Sacramento River system, Marcotte (1984) reported that the suitability of holding pools declines at depths less than 8 feet. Airola and Marcotte (1985) found that spring-run Chinook salmon in Deer and Antelope creeks avoided pools less than about 6 feet deep. In the John Day River in Oregon, adults usually hold in pools deeper than 5 feet that contain cover from undercut banks, overhanging vegetation, boulders, or woody debris (Lindsay et al. 1986). Marcotte (1984) reported that water velocities in holding pools used by spring-run Chinook in Deer and Antelope creeks ranged from 0.5 ft/s to 1.2 ft/s.

Water temperatures for adult Chinook salmon holding are reportedly optimal when less than 60.8°F (16°C), and lethal when above 80.6°F (27°C) (Moyle et al. 1995). Moyle et al. (1995) reported that spring-run Chinook salmon in the Sacramento River typically hold in pools that have temperatures below 69.8°F (21°C) to 77°F (25°C). In Butte Creek, prespawn adult mortalities were minimal when average daily temperatures were less than 66.9°F (19.4°C) with only brief periods of high temperatures up to about 70°F (21°C) in July between 2001 and 2004 (Ward et al. 2006).

In the Stanislaus River, fall-run Chinook salmon probably do not hold for more than 1 or 2 weeks before spawning, based on the time between when they pass the Riverbank weir (S.P. Cramer and Associates 2004, 2005; Cramer Fish Sciences 2006, 2007) and the initiation of spawning (DFG 1991-2005).

3.5 Spawning

Most Chinook salmon spawn in the mainstem of large rivers and lower reaches of tributaries, although spawning has been observed over a broad range of stream sizes, from small tributaries less than 10 feet wide (Vronskiy 1972) to large mainstem rivers (Healey 1991). The adults migrate upstream until they locate a bed of gravel where water temperatures and DO concentrations are suitable for egg incubation. Adult Chinook salmon typically spawn at the tails of pools (a.k.a heads of riffles), where the fish have access to both suitably sized gravel and refuge provided by the depth of the pool (Vronskiy 1972, Chapman 1943, Mesick 2001a). Pool tails may also provide optimum conditions for egg incubation, because surface water tends to downwell into the gravel at pool tails, thereby delivering high concentrations of DO to incubating eggs, and transporting metabolic wastes from the egg pocket.

Gravel that is suitable for spawning consists of a mixture of particle sizes from sand (0.1 to 6.0 inches (0.25 to 15.24 cm)) diameter cobbles, with a median diameter (D_{50}) of 1 to 2 inches (2.54 to 5.08 cm). D_{50} values of gravel for spring-run Chinook have been found to range from 0.4 inches to 3.1 inches (10.8 mm to 78 mm) (Platts et al. 1979, Chambers et al. 1954, 1955, all as cited in Kondolf and Wolman 1993).

Chinook salmon are capable of spawning within a wide range of water depths and velocities (Healey 1991). The water depths most often recorded over Chinook salmon redds range from 0.4 feet to 6.5 feet and velocities from 0.5 feet per second (ft/s) to 3.3 ft/s, although criteria may vary between races and stream basins. For example, fall-run Chinook salmon, because of their larger size, are generally able to spawn in deeper water with higher velocities (Healey 1991), than spring-run Chinook salmon, which tend to dig comparatively smaller redds in finer gravels (Burner 1951). Similarly, 4- and 5-year-old fish are generally larger than the average 3-year-old fish, and can spawn in deeper, faster water with larger gravels and cobbles.

On arrival at the spawning grounds, adult female Chinook salmon dig pits in the gravel bed that are typically 12 inches deep and 12 inches in diameter. During spawning, the female deposits about 1,500 eggs in a pit and then covers them with gravel. Over a period of 1 to several days, the female gradually digs several egg pits in an upstream direction within a single redd (Burner 1951, Healey 1991). By disturbing the gravel that surrounds the egg pocket, the female loosens the bed material and cleans some of the fine sediment from the gravel, thereby improving interstitial water flow. Females can remove from 2 percent to 15 percent of fine sediment smaller than 0.04 inches (<1 mm) during the redd-building process, depending on the initial proportion of fines in the gravel (Kondolf 2000). Before, during, and after spawning, female Chinook salmon defend the redd area from other potential spawners (Burner 1951). Defense of a constructed redd helps to prevent subsequent spawners from constructing redds in the vicinity of an egg pocket, which can dislodge the eggs and increase egg mortality. Adult Chinook salmon females generally defend their redd until they die, usually within 1 to 2 weeks of spawning.

3.6 Adult Carcasses

There is substantial evidence that adult salmon carcasses provide significant benefits to stream and riparian ecosystems. In the past, the large numbers of salmon that returned to streams contributed large amounts of nutrients to the ecosystem (Pearsons et al. 2007, Bilby et al. 1998, Hocking and Reimchen 2002). The carcasses provide nutrients to numerous invertebrates, birds, and mammals, and nutrients from decaying salmon carcasses are incorporated into freshwater biota (Helfield and Naiman 2001, Bilby et al. 1998), including terrestrial invertebrates (Hocking and Reimchen 2002). Helfield and Naiman (2001) found that nitrogen from carcasses is incorporated into riparian vegetation. Merz and Moyle (2006) found marine-derived nitrogen incorporated into riparian vegetation and wine grapes. Merz and Moyle (2006) also compared relative nitrogen contribution rates between salmon-abundant and salmon-deprived rivers. The results indicated that salmon-abundant rivers had much more marine-supplied nitrogen than nonsalmonid bearing rivers (Merz and Moyle 2006). This nutrient supply is a positive feedback loop in which nutrients from the ocean are incorporated into riparian growth that in turn provides ecosystem services by providing additional growth and development of the riparian system. Carcass nutrients are so important to salmonid stream ecosystems that resource managers spread ground hatchery salmon carcasses in Washington streams (Pearsons et al. 2007).

4.0 Stressors

A number of documents have addressed the history of human activities, current environmental conditions, and factors contributing to the decline of Chinook salmon in the Central Valley. The San Joaquin River Restoration Study Background Report, which was compiled by McBain and Trush, Inc. in 2002, describes the changes in habitat and likely stressors that will affect Chinook salmon in the Restoration Area. Another important document is the Final Restoration Plan that was adopted for the Anadromous Fish Restoration Program in 2001 (USFWS 2001). It identifies many stressors that affect spring-run and fall-run Chinook salmon in the Central Valley. The Final PEIS/EIR for the CALFED Bay-Delta Program (CALFED 2000) and the Final PEIS for the Central Valley Project Improvement Act (CVPIA) provide summaries of historical and recent environmental conditions for salmon and steelhead in the Central Valley. The National Marine Fisheries Service (NMFS) prepared range-wide status reviews for West Coast Chinook salmon (Myers et al. 1998). NMFS also assessed the factors for Chinook salmon decline in a supplemental document (NMFS 1996). The following summarizes the information from these documents as well as more recent research on Chinook salmon and their habitats in the Central Valley and other West Coast rivers.

Stressors are discussed according to each life history stage of Chinook salmon, including egg survival and emergence, juvenile rearing, smoltification and downstream migration, ocean survival, adult migration, adult holding for spring-run salmon (Section 4.5), and spawning. In addition, the potential effects of releasing hatchery-reared juvenile salmon and climate change are discussed in terms of recovering naturally spawning populations. The following discussion generally pertains to both spring-run and fall-run salmon, particularly for the juvenile stages, which generally use the same habitats at the same times. The discussion of stressors that affect adult stages will include issues specific for each run.

4.1 Egg Survival and Emergence

Stressors that may affect the survival of eggs and emergence of alevins in the San Joaquin River include high water temperatures, sedimentation (fines deposited in the substrate), turbidity (suspended clay-sized particles), and redd superimposition. Chinook salmon egg mortality rapidly increases as water temperatures exceed 57°F (13.9°C) (Section 3.1.3). High rates of sedimentation of the spawning gravels reduce intragravel flows (Section 3.1.2) and potentially entomb alevins. High levels of turbidity can coat the egg membrane with clay-sized particles that inhibit its ability to absorb oxygen or excrete metabolic wastes (Section 3.1.1).

Other potential stressors for incubating eggs, such as predation, anglers walking on redds, and streambed scour, are not expected to be significant within the Restoration Area. Eggs incubating in natural gravels in the San Joaquin River basin are probably protected

from large invertebrate (e.g., crawfish) or fish (e.g., sculpin) predators because the interstitial spaces in the gravel are too small for predators to reach the egg pockets. Sculpin and crayfish are capable of penetrating deeply into streambeds to feed on salmon eggs and alevins but only where the gravel is coarse and free of fine sediments (McLarney 1964, Phillips and Claire 1966, Vyverberg 2004, pers. comm.). It is also unlikely that walking on redds would harm incubating eggs because the eggs are typically 12 inches below the surface of the gravel and natural gravel beds do not shift easily or otherwise move when walked upon. Montgomery et al. (1996) reported that the tops of chum salmon (*O. keta*) egg pockets were below the level of scour depth that occurred during frequent, bankfull flows in a small West Coast stream. It is likely that Chinook salmon bury their eggs at greater depths than chum salmon (DeVries 1997), therefore, streambed scour should be an unlikely source of mortality for incubating eggs in the Restoration Area.

4.1.1 Excessive Sedimentation and Turbidity

Koski (1966) reported that a majority of mortality in redds was caused by the inability of alevins to emerge due to excessive amounts of fine sediments in the redd. He found numerous dead coho salmon alevins that were completely buttoned-up but extremely emaciated at a depth of 8 inches. Beschta and Jackson (1979) showed that in a flume, fines 0.5 mm in diameter tend to form a barrier in the upper 10 cm of the gravel bed that “seals” against intrusion of fines into the egg pocket but also creates a barrier to emergence. This barrier has been described in salmon redds as a mixture of coarse sand and fines 6 to 12 inches above the egg pocket (Hawke 1978) that has a geometric mean diameter (d_g) lower than the substrate above and below the middle layer (Platts et al. 1979). Bams (1967) reported that when sockeye salmon alevins confronted a sand barrier, they “butted” upward to loosen sand grains and form an open passage to the substrate surface. Koski (1966) reported that the number of days for the first coho salmon alevins to emerge was unaffected by the amount of fines, but that the total duration of emergence for all alevins was longer in redds with high percentages of fines.

Quantification of alevin entombment relative to the amount of fines has been difficult. Researchers who evaluated emergence rates by capping natural redds with nets, such as Koski (1966, 1975), Tagart (1976), and TID and MID (1991), cannot accurately estimate egg survival to emergence (Young et al. 1990) because they did not estimate egg viability, fertilization success, the loss of eggs during deposition in the egg pocket (Young et al. 1990), or escapement of fry that migrate under the trap’s netting (Garcia De Leaniz et al. 1993).

Laboratory studies suggest that alevin entombment occurs over a range of substrate particle sizes, including those < 0.85 mm (Shelton and Pollock 1966), < 3.3 mm (Koski 1966), < 4.67 mm (Tapple and Bjornn 1983), and < 6.4 mm (McCuddin 1977). However, these studies tested the ability of large, healthy alevins to emerge under high concentrations of sand, which is an abnormal condition considering that high concentrations of sand typically result in low DO levels and small, weak alevins.

Flood events, and land disturbances resulting from logging, road construction, mining, urbanization, livestock grazing, agriculture, fire, and other uses may contribute sediment directly to streams or exacerbate sedimentation from natural erosive processes (California Advisory Committee on Salmon and Steelhead Trout 1988, NMFS 1996). High permeability measurements in Reach 1A approximately 5 years ago suggest that sedimentation has not been a problem (Stillwater Sciences 2003). Furthermore, turbidity levels are usually low in the San Joaquin River basin until high rates of runoff occur in January or February, which is after a majority of the eggs have hatched.

4.1.2 Excessively High Water Temperatures

Target incubation temperatures for Chinook salmon are daily maximums of less than 58°F (14°C). Water released from Friant Dam should be less than 58°F (14°C) throughout the spawning period as long as the cold water pool in Millerton Lake is not exhausted. The HEC 5Q water temperature model developed for the Restoration Area (Smith and Deas, unpublished results) suggests that implementing the Restoration Hydrographs could result in maximum temperatures of the Friant release flows of 62°F (16.7°C) in October or November (Figure 4-1). Using hydrologic and climatic conditions from 1980 to 2005, the temperature of the release flows would exceed 60°F during 20 years of the 26-year period (Figure 4-1). It is possible that these temperatures could result in Chinook salmon egg mortality rates of about 50 percent (Section 3.1.3).

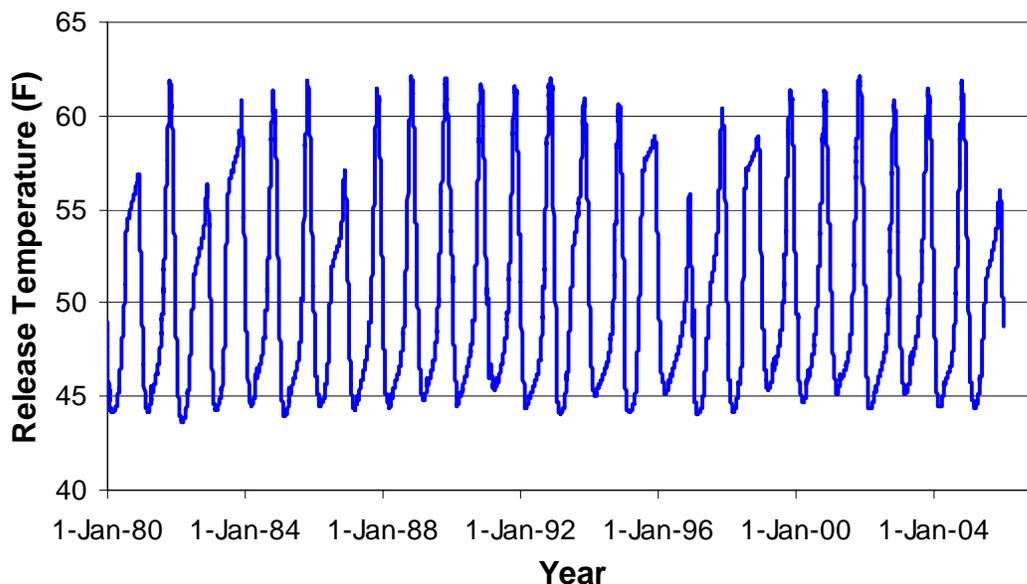


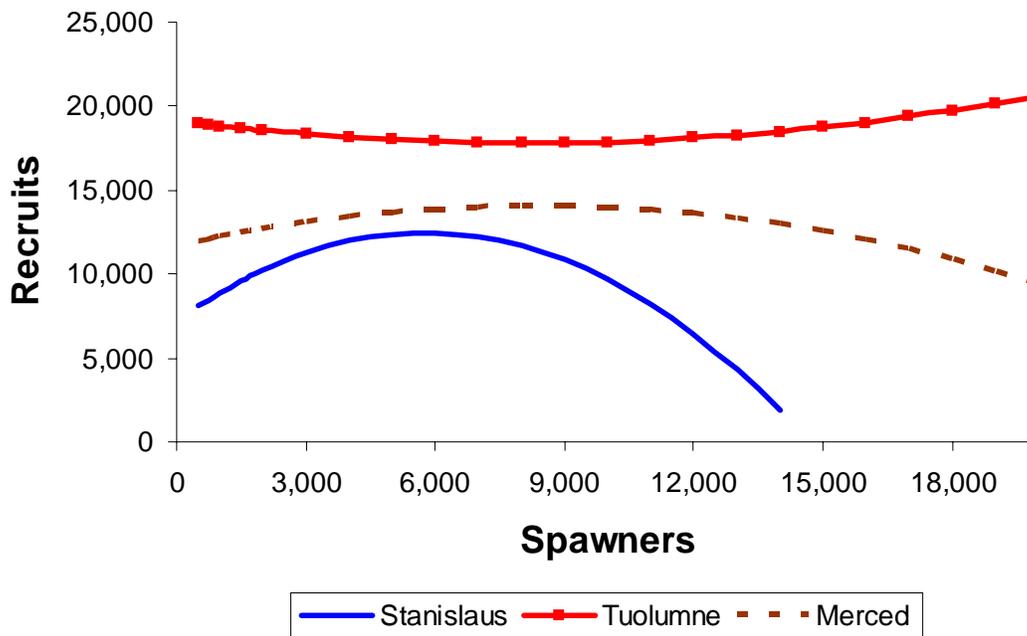
Figure 4-1.
Results of HEC 5Q Water Temperature Model Showing Predicted Water Temperatures of Releases from Friant Dam if Restoration Hydrograph Releases Were Made Under Hydrologic and Climatic Conditions from 1980 to 2005

4.1.3 Redd Superimposition

Redd superimposition occurs when spawning fish construct new redds on top of preexisting redds such that either the eggs in the preexisting redd are either destroyed (dug up) or buried under fines that prevent most of the fry from emerging. Redd superimposition has been reported for the Stanislaus River (Mesick 2001a), American River (Vyverberg 2004, pers. comm.), and the Tuolumne River (TID and MID 1991). Redd superimposition can occur at low escapements and in areas with ample high quality spawning habitat (Mesick 2001a), presumably because spawners prefer to dig redds in the loose gravels provided by preexisting redds that are no longer guarded by the original female. Redd superimposition does not necessarily kill the eggs or entomb the alevins in the original egg pocket, because most superimposing redds are not constructed exactly on top of preexisting redds but rather several feet to the side as well as several feet upstream or downstream from the original redd. Entombment would only occur in superimposed redds constructed in spawning beds where the concentration of fines was relatively high.

Carl Mesick Consultants (2002a) estimated redd superimposition rates in the Stanislaus River during fall 2000 when escapement was relatively high by monitoring superimposition at 82 artificial redds that were constructed in late October before most of the fall-run fish had begun to spawn. In this study, redd superimposition completely disturbed 15 percent of the artificial egg pocket areas (presumably with 100 percent egg mortality) and buried another 23 percent of the artificial egg pocket areas with gravel and fines that could entomb some or all of the alevins.

It is unlikely that redd superimposition limits adult recruitment in the Stanislaus, Tuolumne, and Merced rivers because many more fry are produced at high spawner densities than can be sustained by the quality of the rearing habitat (Section 5.1). Spawner-recruitment relationships for the Tuolumne and Merced rivers are relative flat (Figure 4-2) (Mesick and Marston 2007b), which suggests that high densities of spawners do not reduce adult recruitment to a significant degree. Although a high density of adult spawners has reduced adult recruitment in the Stanislaus River (Figure 4-2), rotary screw trap evidence (Section 5.1) indicates that many more fry were produced than the number of smolt outmigrants from 1998 to 2004 when spawner abundance ranged between 2,400 and 11,650 fish (Mesick and Marston 2007b).



Source: Mesick and Marston 2007.

Note:

A categorical variable called "Population Shift" was used for all three rivers to account for a shift in recruitment that occurred sometime between 1987 and 1994.

The relationships are based on regression models of recruits, quadratic spawner terms ($a_2 + a_1 + c$), and a mean Vernalis flow of 7,000 cubic feet per second from March 1 to June 15.

Figure 4-2.
Spawner-Recruit Relationships for Stanislaus, Tuolumne, and Merced Rivers

4.2 Juvenile Rearing and Migration

Likely stressors for juvenile Chinook salmon rearing in and migrating through the Restoration Area include inadequate food resources, high water temperatures, predation, entrainment at unscreened diversions, contaminated runoff from agriculture and housing development, and disease. These stressors are primarily influenced by flow diversions, agricultural practices, urban development, and gravel excavations.

A relatively small number of Chinook salmon fry that migrate into the lower San Joaquin River (below the confluence with the Merced River) from the San Joaquin River tributaries and Delta are thought to survive except during flood years. Ocean recovery rates of the fry obtained from the Coleman National Fish Hatchery and tagged with coded wire half tags indicate that fry survival was lower in the Central Delta near the mouth of the Mokelumne River than in the North Delta near Courtland, Ryde, or Isleton during dry years, although the difference was not statistically significant (Brandes and McLain 2001). However, during flooding in 1982 and 1983, tagged fry survived at similar rates in the Central Delta and South Delta in the Old River compared to the North Delta (Brandes and McLain 2001). The poor survival of juveniles rearing in the Delta in dry and normal water years may be caused by predation, entrainment at numerous small, unscreened diversions, unsuitable water quality, high water temperatures, inadequate food resources, and direct mortality at the Federal and State pumping facilities in the Delta. Entrainment at the Delta pumping facilities may be minimal during very wet years because tagged fry were collected at the pumping facilities only during the dry years whereas none were collected in wet years (Brandes and McLain 2001). Although the fry migration life stage does not appear to contribute as much to current production of the population in San Joaquin River tributaries and the Delta, it may be an important life stage in rivers with functional floodplain habitats in downstream reaches, such as Sutter Bypass on Butte Creek (Ward and McReynolds 2001, Ward et al. 2002) and possibly in restored floodplain and wetland habitats in the lower Restoration Area, where fry can rapidly grow to a smolt size because of warmer water temperatures and abundant food resources.

4.2.1 Food Resources

The survival of juvenile Chinook salmon to the adult stage partially depends on their ability to grow rapidly enough to begin their downstream migration as smolts early in the spring when their chances are highest to survive their migration through the Delta and estuary to the ocean. In addition, it is highly likely that large, healthy smolts will survive their migrations at higher rates than would smaller, poorly fed smolts.

It is likely that food resources in the Restoration Area will be adversely affected by a combination of factors:

- Reduced flows or dikes that substantially reduce the contribution of organic matter and prey-sized invertebrates from inundated floodplains
- Sedimentation and gravel extraction that affects the production of in-river, prey-sized invertebrates

- Lack of nutrients provided by low numbers of adult salmon carcasses
- Reduced native riparian and wetland vegetation that is the primary basis of the aquatic food web
- Lack of organic matter and prey-sized invertebrates from upstream reservoirs
- Pesticides and other contaminants that reduce the abundance of food organisms
- Competition for food with native and introduced species

Floodplain Inundation and River Connectivity

Most of the energy that drives aquatic food webs in rivers is derived from terrestrial sources (Allan 1995), and aquatic productivity is related to flood magnitude and the area inundated in some rivers (Large and Petts 1996). Flooding, particularly the rising limb of the hydrograph, typically results in high concentrations of both dissolved and particulate organic matter being released into the river (Allan 1995). High flows that inundate floodplains also provide food for juvenile fish that rear in floodplain habitats. Research in other river systems has shown that production of invertebrates, the most important prey resource for many fishes, on inundated floodplains can far exceed river production. Sommer et al. (2001) found that drift invertebrates, the primary prey of juvenile salmonids, were more abundant on the inundated Yolo Bypass floodplain than in the adjacent Sacramento River. As a result, growth, survival, and feeding success of juvenile Chinook salmon were higher in the Yolo Bypass, the primary floodplain of the lower Sacramento River, than in the adjacent mainstem channel in 1998 and 1999 (Sommer et al. 2001). Gladden and Smock (1990) estimated that annual invertebrate production on two Virginia floodplains exceeded river production by one to two orders of magnitude.

Floodplain habitats tend to produce small invertebrates with short life cycles such as chironomids and cladocerans (McBain and Trush 2002). However, the duration and frequency of floodplain inundation can be an important determinant of invertebrate production and community structure. In the Virginia floodplains studied by Gladden and Smock (1990), annual production on the floodplain continuously inundated for nine months was 3.5 times greater than on the floodplain inundated only occasionally during storms. On Cosumnes River floodplains, Grosholz and Gallo (2006) found that the invertebrate community structure was regulated by the timing and duration of inundation of the floodplain. Planktonic crustaceans emerged first, followed by insect macroinvertebrates. Importantly, juvenile fish diets tracked the species composition of the emerging invertebrate community subsequent to inundation of the floodplain.

Lateral connectivity of river channels to adjacent floodplains has been shown to be an important control on the timing, composition, and total invertebrate biomass in a river. In the Rhone River basin, Castella et al. (1991) have shown, using a series of connectivity indices, that invertebrate diversity and biomass in the river can be linked to the connectivity of the river to its floodplain. The mainstem San Joaquin River is bordered by San Joaquin River Flood Control District levees and individual landowner levees (McBain and Trush 2002) resulting in a separation of much of the river from its historic floodplain.

Invertebrate colonization of a rewatered river channel or newly inundated floodplain is regulated by three primary mechanisms: proximity to a source of colonists, the *in situ* invertebrate “seedbank” in the substrate, and the timing and duration of inundation. In Alabama’s Sipsey River, Tronstad et al. (2005) showed that invertebrate species composition and the timing of recolonization is controlled by the frequency of inundation of invertebrate “seedbanks” in floodplain soils: recently inundated soils had faster rates of emergence and greater species diversity than soils with a longer interval between periods of inundation. This disparity suggests that invertebrate production in newly rewatered reaches and adjacent floodplains of the San Joaquin River may be directly related to the length of time since they were last wetted. Constructed floodplains, for example, may take considerably longer to become productive than bypass channels that receive flood flows during periodic storm events. The invertebrate community in the upper Sacramento River recovered to a composition similar to undisturbed sections of the river within 1.5 years after sterilization by a chemical spill (Boullion 2006 as cited in Cantara Trustee Council 2007). The source of invertebrates from immediately upstream areas likely contributed to the rapid recolonization of the upper Sacramento River, and a similar situation can be expected when Restoration Flows are released into the formerly dewatered reaches of the San Joaquin River in the Restoration Area.

The physical habitat structure of the rewatered habitat also plays a role in the rate, composition, and maintenance of invertebrate communities. Hilborn (1975) demonstrated that habitat heterogeneity is a fundamental control on ecosystem community structure. A simple sand-bedded channel with no riparian habitat (i.e., homogeneous habitat) will typically have lower invertebrate diversity than a comparable channel that is more complex and includes substrate size variability and developed riparian vegetation. Fundamentally, channel heterogeneity equates to more niches for more types of invertebrates. For example, Benke (2001) found that invertebrate diversity and biomass in Georgia rivers was higher in a system with a well developed floodplain and abundant large woody debris (LWD) in the river, than in an otherwise similar system with lower habitat diversity. In the Restoration Area, channels and floodplains with existing habitat complexity (e.g., riparian vegetation, LWD) are likely to support higher invertebrate production and diversity than homogeneous channels or newly constructed floodplains.

Indirect Effects of Pesticides and Other Contaminants

It is likely that contaminants usually do not kill juvenile salmon directly, but instead substantially reduce their food resources or increase their susceptibility to disease or pathogens. However, the observed concentrations of organophosphate pesticides in water samples collected in the San Joaquin River at Vernalis and most other locations in the Delta in January through April in 2001 and 2002 shortly after rainfall events, when contaminant levels are highest (Werner et al. 2003), were seldom toxic to a cladoceran (*Ceriodaphnia dubia*), a resident cladoceran (*Simocephalus vetulus*), a chironomid larvae (*Chironomus tentans*), and an amphipod (*Gammarus daiberi*). Results of surveys conducted between 1992 and 2000 suggest that the amounts of organophosphate pesticides applied as dormant sprays in the San Joaquin River basin have steadily decreased over the past decade, although they still exceed criterion maximum concentration levels established by DFG (Orlando et al. 2003). Since 1993, there has

been a shift in insecticides in the Central Valley from organophates to permethrin and finally to the new compounds of pyrethroids, which are nearly 20 times more toxic to aquatic invertebrates and fish than permethrin (Amweg et al. 2005). Despite the fact that pyrethroids are now one of the most important and fastest growing insecticides applied in the Central Valley, primarily for agriculture and urban uses, only a limited number of studies and monitoring efforts are focusing on occurrence and toxicity (Oros and Werner 2005). Unfortunately, there are not enough field monitoring data on the spatial and temporal occurrences of pyrethroids for making risk assessments to date (Oros and Werner 2005).

Sedimentation and Gravel Extraction

Sedimentation, which is the deposition of fine sand (<0.2 mm), and gravel extraction, which created ditches and ponds in the riverbed and floodplain, have probably reduced the availability of food resources for juvenile salmon in the Restoration Area. Waters (1995) suggested that a change from gravel and cobble riffles to deposits of silt and sand results not only in a decrease in abundance of invertebrates that are important as fish foods, but also results in a change in invertebrate species from those inhabiting the interstitial spaces of large particles to small, burrowing forms less available to fish. However, captured mine pits in the San Joaquin River basin typically store large volumes of organic matter and contain dense growths of aquatic vegetation. There is an abundant “hatch” of adult aquatic insects from these ponds, and it is possible that these ponds provide more food than is produced in the main channels.

Nutrients from Adult Salmon Carcasses

After spawning, adult Chinook salmon carcasses remain in the stream corridor to decompose, and are an important food and nutrient source within a watershed (Cederholm et al. 1999). Decomposing salmon carcasses are recognized as a source of marine-derived nutrients that play an important role in the ecology of Pacific Northwest streams (Gresh et al. 2000). On the Olympic Peninsula in Washington, 22 different animal species were observed feeding on salmon carcasses (Cederholm et al. 1999). Carcass nutrients can affect the productivity of algal and macroinvertebrate communities that are food sources for juvenile salmonids. Decomposing salmon carcasses have also been shown to be vital to the growth of juvenile salmonids (Bilby et al. 1998; Bilby et al. 1996, as cited in Gresh et al. 2000).

The relatively low abundance of salmon and steelhead has significantly reduced this important nutrient source in the Central Valley, and throughout the Pacific Northwest. The study by Gresh et al. (2000) estimated that the annual biomass of salmon entering Pacific Northwest streams (California, Oregon, Washington, Idaho) was historically on the order of 352 million pounds, and has been reduced to only approximately 26 million pounds, a reduction of over 93 percent. Channelization and removal of LWD can also decrease the retention of salmon carcasses and reduce nutrient input.

Riparian Vegetation

Historically, canopy species within the riparian corridor in the upper reaches of the Restoration Area (Reaches 1 and 2A) consisted of a patchy band of cottonwoods, willows, and valley oaks on floodplain and terrace surfaces between the confining bluffs.

In the downstream reaches (downstream from Mendota), there were large flood basins (low-lying areas adjacent to the river channel) dominated by tule marsh on both sides of the river, often many miles wide. Riparian canopy species (cottonwood, willow, valley oak) were limited to relatively narrow bands (typically less than 1,000 feet wide based on 1914 maps) of mineral soil berms deposited along channels that dissected the vast tule marsh.

Conversion of native vegetation types to agriculture, aggregate mining, and urban development has strongly impacted the San Joaquin River's wetlands and riparian habitat. As of 1998, approximately 25,380 and 6,030 acres of riparian and wetland habitats have been converted to agricultural and urban uses, respectively (McBain and Trush 2002). Approximately 4,610 and 1,920 acres of riparian forest and riparian scrub, respectively, were present in 1998 (McBain and Trush 2002).

The San Joaquin riparian corridor, like most California landscapes, is host to many nonnative invasive plant species. In 2000, the California Department of Water Resources (DWR) mapped vegetation along the San Joaquin River from Friant Dam to the confluence with the Merced River (DWR 2002). DWR mapping identified 127 nonnative plant species, which was 50 percent of all plant species identified. The primary nonnative invasive species identified in the DWR mapping include tree-of-heaven, giant reed, pampas grass, eucalyptus, edible fig, white mulberry, Lombardy poplar, castor bean, Himalayan blackberry, scarlet wisteria, and tamarisk (DWR 2002). The DWR effort also recorded parrot's feather, a highly invasive aquatic plant. Nonnative invasive plant species cover 99 acres along the river corridor in nearly monospecific stands, and occur as a component of most, if not all, native vegetation types (Chapter 8 in McBain and Trush 2002). These plant species are particularly abundant in Reach 1, where high levels of disturbance may have aided their spread, as suggested by their distribution in and around aggregate mining pits (McBain and Trush 2002).

Exotic plant species can alter the structure and dynamics of natural ecosystems. Nonnative plant species can impact native wildlife by displacing native vegetation that is used for nesting or as a food source. Once established, nonnative plant species can alter nutrient cycling, energy fixing, food web interactions, and fire and other disturbance regimes, to the extent that the native landscape is changed. Habitat fragmentation contributes to the spread of nonnative species by increasing edge habitat, which provides greater opportunities for invasion by exotic species (Cox 1999). Ecosystem alterations resulting from nonnative plant species invasions can be exacerbated by activities such as overgrazing and vegetation clearing that create favorable conditions for further nonnative plant establishment (Cox 1999, Randall and Hoshovsky 2000). Alteration of historical flooding regimes by flow regulation further promotes invasions by nonnative species by eliminating processes necessary for recruiting and maintaining native plant species (Cox 1999).

Reservoir Productivity

The San Joaquin River basin upstream from Millerton Lake consists of granitic soils with low mineral nutrient content (Reclamation 2006). Partly as a result, Millerton Lake is low in total dissolved solids (TDS) and has low levels of chemical nutrients (Dale

Mitchell, 2006, pers. comm.). Little information is available regarding the plankton communities of Millerton Lake, but there is evidence that plankton production varies considerably on a seasonal basis. Cladocerans in the genus *Leptodora* (water fleas) have been observed to be abundant in Millerton Lake during summer months, with population crashes commonly occurring in September (Dale Mitchell 2006, pers. comm.). Threadfin shad in Millerton Lake are known to feed extensively on *Leptodora*, indicating that this organism may be seasonally available as a food source for fishes in the San Joaquin River downstream from Friant Dam.

Competition with Native and Introduced Species

Some nonnative fish species have habitat requirements that overlap with those of native species. These species may be more aggressive and territorial than native species, resulting in the exclusion of native species from their habitats. Many of the nonnative species, such as green sunfish, also tolerate extremely high water temperatures and appear better able to persist in water with low DO, high turbidity, and contaminants than native fishes.

The arrival of the Asiatic clams *Corbicula fluminea* and *Corbula amurensis* in the San Francisco Estuary disrupted the normal benthic community structure and depressed phytoplankton levels in the estuary due to the highly efficient filter feeding of these clams (Cohen and Moyle 2004). The decline in the levels of phytoplankton reduces the population levels of zooplankton that feed on them, and hence reduces the forage base available to salmonids transiting the Delta and San Francisco Estuary. This lack of forage base can adversely impact the health and physiological condition of these salmonids as they migrate through the Delta region to the Pacific Ocean.

Introductions of exotic zooplankton species have supplanted other zooplankton species that provided important food resources for fish in the upper San Francisco Estuary (Hennessy and Hieb 2007). In 1993, *Limnoithona tetraspina*, an introduced cyclopoid copepod, mostly replaced the historically common and larger *L. sinensis*. The introduced copepod, *Pseudodiaptomus forbesi*, along with the Asiatic clam, *C. amurensis*, contributed to the decline of the calanoid copepod, *Eurytemora affinis*, beginning in the late 1980s. *E. affinis* was an important food resource for juvenile fish. The introduced calanoid copepod, *Sinocalanus doerrii*, was first recorded in spring 1979. In contrast, the native cladocerans, *Bosmina*, *Daphnia*, and *Diaphanosoma*, and the native rotifer, *Synchaeta bicornis*, have gradually declined since the early 1970s. It is likely that relatively small exotic species, such as *L. tetraspina*, are not as important in the juvenile salmonid forage base as were the displaced native species.

4.2.2 Disease

USFWS conducted a survey of the health and physiological condition of juvenile fall-run Chinook salmon in the San Joaquin River and its primary tributaries, the Stanislaus, Tuolumne, and Merced rivers, during spring 2000 and 2001 (Nichols and Foott 2002). *Renibacterium salmoninarum*, the causative agent of bacterial kidney disease (BKD), was detected in naturally produced juveniles caught in rotary screw traps from the Stanislaus and Tuolumne rivers and juveniles caught with a Kodiak trawl at Mossdale in the San Joaquin River. No gross clinical signs of BKD were seen in any of the fish examined.

However, these low-level infections might remain active after the fish enters the ocean where the clinical symptoms might develop.

Proliferative kidney disease (PKD) was detected in both natural and hatchery juveniles from the Merced and mainstem San Joaquin rivers in 2000 and 2001 (Nichols and Foott 2002) and in natural juveniles from the Merced River in 2002 (Nichols 2002). The myxozoan parasite *Tetracapsula bryosalmonae*, which causes PKD, was detected in the kidney samples of only 2 percent of the juvenile Merced River fish in April 2000, but in 90 percent of the April 2001 samples, 100 percent of the May 2001 samples, and 51 percent of the April 2002 samples. Heavy infections were observed in 22 percent of the samples in 2002 (Nichols 2002). These data suggest that the incidence of pathogen infection is low in above normal water years such as 2000 compared to dry water years such as 2001 and 2002. PKD has been described at the Merced River Fish Hatchery since the 1980s and in California since at least 1966. It compromises the fish's performance in swimming, salt water entry, and disease resistance (Nichols and Foott 2002). Nichols and Foott (2002) suggested that PKD could be a significant contributor to mortality in natural fish.

Columnaris disease, caused by the bacterium *Flexibacter columnaris*, was observed in juvenile Chinook salmon caught in rotary screw traps in the Stanislaus River in spring 2007. The disease can rapidly increase in the population as water temperatures reach a mean daily temperature of 68 to 69.8°F (20 to 21°C). Along with the protozoan *Ichthyophthirius multifiliis* (Ich), columnaris was a leading cause of adult salmon mortality in the lower Klamath River in 2002.

There were no signs of infection from pathogenic species of bacteria, including *Aeromonas salmonicida*, *Yersinia ruckeri*, and *Edwardsiella tarda*, in the San Joaquin River basin during spring 2001 (Nichols and Foott 2002). Although *Myxobolus cerebralis*, the causative agent of whirling disease, was not detected in a pooled sample of 194 fish, the parasite has been detected in rainbow trout from the Stanislaus River. Tests were not conducted for *Flavobacterium columnare*.

The pathogen *Ceratomyxa* is present in the Central Valley and studies indicate that it causes a high mortality rate of Chinook smolts migrating through the lower Willamette River, Oregon (Steve Cramer 2001, pers. comm.). This disease relies on tubifix worms for an intermediate host and the worms flourish in organic sediments. It is likely that the worms multiply and the disease spreads in years when organic sediments are not flushed by high flows. There are indications that mortality of smolts due to this disease increases in drought years and decreases in wet years. *Ceratomyxa* disease is a particular concern for the San Joaquin River because there is a tubifix worm farm located in Reach 1A, at RM 261 (Jones and Stokes 2002a). It is also possible that organic sediments accumulate and produce tubifix worms in captured mine pits.

4.2.3 Predation

Fish species in the Restoration Area that will probably prey on juvenile salmon include largemouth bass (*Micropterus salmoides*), smallmouth bass (*M. dolomieu*), Sacramento pikeminnow (*Ptychocheilus grandis*), green sunfish (*Lepomis cyanellus*), warmouth (*L.*

gulosus), black crappie (*Pomoxis nigromaculatus*), and striped bass (McBain and Trush 2002). DFG (2007a) electrofishing surveys of the Restoration Area in 2004 and 2005 indicated that largemouth and spotted bass (*M. punctulatus*) were prevalent as far upstream as Reach 1B and were very common in the lower reaches of the river. Largemouth bass are adapted to low flow and high water temperature habitats and typically inhabit captured mine pits in the San Joaquin River basin. Smallmouth bass are adapted to riverine habitats but are also relatively inactive when water temperatures are low. Large salmonids, such as rainbow trout at least 140 mm in fork length, would also be expected to prey on juvenile Chinook salmon. Juvenile coho salmon, 140 mm in length, have eaten 64 mm Chinook salmon (Pearsons and Fritts 1999). Although planted catchable-sized rainbow trout might prey on juvenile Chinook salmon, it is DFG policy not to plant hatchery trout in rivers that contain anadromous fish populations, such as Chinook salmon.

Juvenile salmonids are also susceptible to avian predators. Species including California gulls, ring-billed gulls, Caspian terns, double-crested cormorants, and American white pelicans have been documented to prey on outmigrating steelhead and salmon as they pass through dams on the Columbia and Snake rivers (Bayer 2003). Fish-eating birds that occur in the California Central Valley include great blue herons (*Ardea herodias*), gulls (*Larus spp.*), osprey (*Pandion haliaetus*), common mergansers (*Mergus merganser*), American white pelicans (*Pelecanus erythrorhynchos*), double-crested cormorants (*Phalacrocorax spp.*), Caspian terns (*Sterna caspia*), belted kingfishers (*Ceryle alcyon*), black-crowned night herons (*Nycticorax nycticorax*), Forster's terns (*Sterna forsteri*), hooded mergansers (*Lophodytes cucullatus*), and bald eagles (*Haliaeetus leucocephalus*) (Stephenson and Fast 2005). These birds have high metabolic rates and require large quantities of food relative to their body size.

Predation in Central Valley Rivers

High predation rates are known to occur below small dams, such as RBDD and Sack Dam in the Restoration Area. As juvenile salmon pass over small dams, the fish are subject to conditions that may disorient them, making them highly susceptible to predation by other fish or birds. In addition, deep pool habitats tend to form immediately downstream from the dams where Sacramento pikeminnow (*Ptychocheilus grandis*), striped bass, and other predators congregate. Tucker et al. (1998) showed high rates of predation by Sacramento pikeminnow and striped bass on juvenile salmon below the RBDD.

EA Engineering, Science and Technology (TID and MID 1992), conducted river-wide electrofishing surveys in the Tuolumne River in spring 1989 and 1990, and found that few largemouth and smallmouth bass contained naturally produced juvenile Chinook salmon in their stomachs, whereas bass had numerous hatchery-reared juvenile salmon in their stomachs shortly after the fish were released for a survival study (Table 4-1). It is likely that there were numerous naturally produced juvenile salmon during both years because there was a moderate number of spawners present during both years: 5,779 and 1,275 present in fall 1988 and 1989, respectively (DFG Grand Tab estimates). The spring 1990 studies should have been particularly effective for evaluating predation because the electrofishing was conducted at night, shortly after the bass would have been

feeding and their stomachs would have still have contained undigested juvenile salmon. In addition, the study was conducted during a drought, when predation rates would be expected to be highest due to low flows and high water temperatures. These results suggest that bass prey on few naturally produced juveniles because they primarily migrate at night when predation rates are lowest, whereas hatchery fish typically migrate during the day (Roper and Scarnecchia 1996) and they are thought to be naïve at avoiding predators.

**Table 4-1.
Predation Studies in Lower Tuolumne River in 1989 and 1990**

Sampling Dates	La Grange Flows (cfs)	Percent Largemouth Bass with Juvenile Salmon in Stomachs	Percent Smallmouth Bass with Juvenile Salmon in Stomachs	Origin of Juvenile Salmon
4/19 to 5/17, 1989	40 – 121	3.6 (2 out of 56)	8.6 (5 out of 58)	Naturally Produced
1/29 to 3/27, 1990	142 – 174	2.1 (2 out of 97)	3.1 (1 out of 32)	Naturally Produced
4/25 to 4/28, 1990	187 – 207	2.6 (2 out of 76)	6.3 (1 out of 16)	Naturally Produced
5/2 to 5/4, 1990	299 -572	26 (40 out of 152)	33.3 (6 out of 18)	CWT Hatchery

Source: TID and MID 1992

Key:

cfs = cubic feet per second

CWT = coded wire tag

Striped bass, which primarily migrate into the San Joaquin River tributaries during the late-winter and spring (S.P. Cramer and Associates 2004, 2005; Cramer Fish Sciences 2006, 2007), were the primary predators of juvenile fall-run salmon fitted with radio tags in a Stanislaus River study (Demko et al. 1998). Although more than 90 percent of the radio-tagged fish appear to have been eaten by predators, there is uncertainty as to whether gastrically implanting the radio tags, which had 12-inch long external whip antennas, impaired the ability of the juvenile salmon to avoid predators.

Adult Sacramento pikeminnow, which form large schools in ditch-like channels 3 to 8 feet deep, are very abundant in the San Joaquin River basin and prey on salmon fry. Although none of the electrofishing studies conducted in the Tuolumne and Stanislaus rivers identified pikeminnow as predators of juvenile salmon, it is relatively difficult to capture schooling Sacramento pikeminnow with electrofishing gear, and they have complex stomachs that may be difficult to sample using flushing techniques.

Predation in the Delta

Striped bass, Sacramento pikeminnow, and largemouth bass are predators of juvenile salmon in some Delta habitats. Pickard et al. (1982) reported that juvenile salmon predation was high for both Sacramento pikeminnow and striped bass in the Sacramento River Delta between 1976 and 1978. Gill nets were set in Horseshoe Bend and near Hood to collect predators between February 1976 and February 1978. The results suggest that 150 to 1,050 mm fork length striped bass and 300 to 700 mm fork length Sacramento pikeminnow primarily fed on fry and relatively few smolts. Feeding rates for pikeminnow and striped bass were highest in winter (December through February), when

77.7 percent had fish in their stomachs, and low during spring (March through May), when only 23.3 percent had fish in their stomachs. However, stomach evacuation rates would be expected to be higher during the spring; therefore, an in-depth analysis is needed to determine the relative predation rates for fry and smolts. Relatively few steelhead, white catfish (*Ictalurus catus*), channel catfish (*I. punctatus*), and black crappie (*Pomoxis nigromaculatus*) were caught in the gill nets at Horseshoe Bend.

In contrast, Nobriga et al. (2003) used seines and experimental gill nets to sample age-0 and age-1 striped bass and largemouth bass in 3- to 13-foot-deep water in the Yolo-Bypass, lower Sacramento River, and in the Central Delta from March through June 2001. They reported that only 1 juvenile Chinook salmon was found in the stomach of 1 of 81 striped bass and another juvenile Chinook salmon was found in the stomach of 1 of 63 largemouth bass. These predators were primarily feeding on yellowfin goby (*Acanthogobius flavimanus*), gammarid amphipods, *Corophium*, and/or aquatic insects.

Densities of black bass and striped bass are about 3 times higher in the central Delta downstream from Rough and Ready Island near Stockton and in the Mokelumne River (eastern Delta) than in the northern or southern areas of the Delta based on a DFG resident fish study conducted from 1980 to 1983 (Table 4-2, DFG, unpublished data). DFG introduced Florida largemouth bass into the Delta in the early 1980s and again in 1989, and catch rates of black bass have increased since 1993 (Lee 2000). Although predation of juvenile salmon in the Delta has not been quantified, predation would contribute to the low survival rates of juvenile salmon migrating between Dos Reis and Jersey Point and to Sacramento River juveniles migrating into the Mokelumne River through the Delta Cross Channel.

Table 4-2.
Number and Mean Fork Length of Largemouth Bass, Smallmouth Bass, and Striped Bass per Kilometer Collected in DFG Electrofishing Surveys in Sacramento-San Joaquin Delta, 1980 to 1983

Location	Largemouth Bass 208 mm FL	Smallmouth Bass 225 mm FL	Striped Bass 140 mm FL
Central Delta	12.81	0.02	0.03
Eastern Delta	12.92	0.20	0.19
Southern Delta	4.42	0.36	0.03
Northern Delta	3.83	0.78	0.03
Western Delta	5.97	0.08	0.00

Note:

The sampling sites in each region of the Delta are shown in Figure 1 of Schaffter (2000).

Key:

DFG = California Department of Fish and Game

mm = millimeter

FL = fork length

4.2.4 Water Quality

Water quality in the valley floor of the San Joaquin River basin has been impaired as a result of contamination from a variety of sources, including (1) aquatic and terrestrial herbicide application, (2) urban and agricultural pesticide application, (3) trace elements from industrial activities as well as those naturally present in soils, and (4) effluent from wastewater treatment plants and livestock operations, particularly dairy farms. Point sources of pollution originate from single identifiable sources, whereas nonpoint sources that originate from many different sources. Examples of nonpoint sources are agricultural runoff (e.g., excess fertilizers, herbicides, and pesticides) and urban stormwater containing oil, grease, heavy metals, polycyclic aromatic hydrocarbons, and other organics (CVRWQCB 1998). Impervious surfaces (e.g., concrete) tend to reduce water infiltration and increase stormwater runoff (NMFS 1996).

In general, water contamination or degradation may cause chronic or sublethal effects that compromise the physical health of aquatic organisms and reduce their survival over an extended period of time beyond initial exposure. For example, a study conducted in Puget Sound, Washington (Arkoosh et al. 1998), indicates that emigrating juvenile Chinook salmon exposed to contaminants, polycyclic aromatic hydrocarbons, and polychlorinated biphenyls suffered increased susceptibility to the common marine pathogen (*Vibrio anguillarum*). Similarly, a laboratory study suggests that sublethal concentrations of pollutants can be acting synergistically with endemic pathogens of juvenile Chinook salmon, thus compromising survivorship through immunologic or physiologic disruption (Clifford et al. 2005). Although less common, high concentrations of particular contaminants (e.g., ammonia, hydrogen sulfide, herbicides, pesticides) may lead to acute toxicity and death after only short exposure times.

Recent studies suggest that chronic or sublethal effects of contaminants may be subtle and difficult to detect. For example, early experimental studies indicated that hatchery-reared juvenile Chinook salmon exposed to undiluted agricultural subsurface drainwater from the west side of the San Joaquin River had > 75 percent mortality, whereas there were no chronic detrimental effects on the growth and survival of the study fish exposed to agricultural return flows that were diluted by \geq 50 percent (Saiki et al. 1992). However, recent studies suggest that juvenile fall-run Chinook salmon died in the laboratory after eating selenium-contaminated invertebrates and prey fish over a 90-day period that were collected from the San Joaquin River basin (Beckon 2007). These two sets of studies suggest that bioassays of fathead minnows in water samples from the San Joaquin, Merced, Tuolumne, and Stanislaus rivers that showed little evidence of toxicity (Brown 1996) may not have detected chronic or sublethal effects that may affect salmon.

Herbicides

Chemicals containing ingredients such as diquat dibromide, free and complexed copper (e.g., copper ethylenediamine), fluridone, glyphosate, dimethylamine salt of 2,4-dichlorophenoxyacetic acid, and alkylphenoxyethoxylates are applied to control aquatic weeds such as *Egeria densa* and water hyacinth (*Eichhornia crassipes*) in the Delta (DFG 2004). The primary impacts of diquat dibromide and fluridone are sublethal to juvenile Chinook salmon causing narcosis, rheotropism, chemical interaction, and

immunotoxicity (NMFS 2006a). Exposure of juvenile Chinook salmon to these herbicides can increase their vulnerability to predation from both piscine and avian predators as well as reduce valuable invertebrate prey items (NMFS 2006a). In addition, the application of herbicides may result in low DO concentrations as the plants decompose (NMFS 2006a, 2006b).

Pesticides

Recent studies have indicated a serious potential risk of pesticides to exposed early life stages of Chinook salmon in the Central Valley of California (Viant et al. 2006). A large number of pesticides have been detected by water quality sampling programs in the San Joaquin basin, including aldrin, carbaryl, chlorpyrifos, diazinon, dieldrin, diuron, heptachlor, lindane, malathion, metribuzin, and trifluralin (Domagalski et al. 2000). Most problems occur in the lower Restoration Area (Reaches 3 through 5) where water quality is influenced by water imported from the Delta and by agricultural drainage, particularly from Mud and Salt sloughs. Reaches 1 and 2 have generally good water quality (Brown 1997). Domagalski's study (et al. 2000) and other multiyear studies (Brown 1997, Panshin et al. 1998) assessed a wide array of contaminants. More than half of the surface water samples from certain agricultural drainages in the Central Valley contain seven or more pesticides (Panshin et al. 1998). These pesticide mixtures include organophosphates and carbamates that are likely to have additive effects on the neurobehavior of salmon exposed in contaminated watersheds (Scholz et al. 2006). The growing number of chemical pesticides found in the San Joaquin Valley is too large to encompass in this review. Furthermore, accurately quantifying risks of individual pesticides or synergistic effects of multiple pesticides is not easily validated; most studies rely on comparing contaminant levels (from biota or the environment) to literature values, regional or national statistics, or suitable reference sites.

USGS NAWQA Toxicity Monitoring. The San Joaquin-Tulare study unit was among the first basins chosen for the U.S. Geological Survey (USGS) National Water Quality Assessment Program (NAWQA), and has recently focused considerable attention on pesticide contamination in the San Joaquin River basin (Dubrovsky et al. 1998, Panshin et al. 1998, Kratzer and Shelton 1998, Brown and May 2000). Generally, toxicity within the San Joaquin River has been attributed to pesticides from agricultural nonpoint sources, substantiated by the lack of detection of pesticide compounds in reference sites on the upper Kings River and Tuolumne River, situated above agricultural influences (Dubrovsky et al. 1998). In the NAWQA studies, available drinking water standards were not exceeded at San Joaquin River monitoring sites, but the concentrations of several pesticides exceeded the criteria for the protection of aquatic life. As mentioned previously, regional or national contamination levels are used to interpret San Joaquin River study results. Gilliom and Clifton (1990, from Brown 1998) reported that the San Joaquin River had some of the highest concentrations of organochlorine residues in bed sediments among the major rivers of the United States. Although the organochlorine pesticide DDT (dichloro-diphenyl-trichloroethane) was banned in the United States in 1973, DDT concentrations have continued to be detected in biota of the San Joaquin Valley streams at lower levels (Goodbred et al. 1997, Dubrovsky et al. 1998) as contaminated soils are transported to streams and sediment is resuspended from riverbeds.

Concentrations of organophosphate pesticides (*i.e.*, diazinon and chlorpyrifos) in runoff are high, and highly variable during winter storms (Kratzer and Shelton 1998). In winter, dormant-spray pesticides, including diazinon and chlorpyrifos are applied to fruit orchards and alfalfa fields in the San Joaquin River basin and Delta islands (Kuivila 1995, 2000). These pesticides are delivered to local watercourses and the Delta by overland runoff. Diazinon is the common name of an organophosphorus (OP) pesticide used to control pest insects in soil, on ornamental plants, and on fruit and vegetable field crops. Chlorpyrifos is also an OP pesticide and is used to kill insect pests by disrupting their nervous system. OP pesticides were originally developed for their water solubility and ease of application. After they have been applied, they may be present in the soil, surface waters, and on the surface of the plants that are sprayed, and may be washed into surface waters by rain.

Reaches 1 and 2 of the San Joaquin River have not been identified as problem areas by the NAWQA studies, but pesticides have been detected in groundwater samples from domestic water supply wells. However, concentrations of pesticides in groundwater supplies generally have not increased in the last decade (Dubrovsky et al. 1998). The extremely low levels of pesticides and herbicides, and ephemeral nature of their presence in surface waters, prompted the creation of the California Department of Pesticide Regulation (DPR) within the California Environmental Protection Agency (CalEPA), which tracks pesticide use. Data are available at the following web site:

<http://www.cdpr.ca.gov/dprdatabase.htm>

Basin Plan Objectives and CVRWQCB Monitoring. For most pesticides, numerical water quality objectives have not been adopted, but a number of narrative water quality objectives (e.g., no adverse effects) for pesticides and toxicity are listed in the Basin Plan (CVRWQCB 1998). The EPA criteria and other guidelines are also extremely limited, since numerical targets based on the anti-degradation policy would not allow pesticide concentrations to exceed natural “background” levels (*i.e.*, nondetectable levels or “zero”). For the San Joaquin River system, including the five reaches of this study area, the California State Water Quality Control Board (SWRCB) has set a goal of “zero toxicity” in surface water. This goal is intended to protect the beneficial uses of Recreation, Warm Freshwater Habitat, Cold Freshwater Habitat, and Municipal and Domestic Supply from potential pesticide impacts.

The most recent 303(d) list of impaired waterbodies presented by the Central Valley Region Water Quality Control Board (CVRWQCB) identifies Reaches 3, 4, and 5 of the San Joaquin River study area, Mud Slough, and Salt Slough as impaired due to pesticides and “unknown toxicity.” In addition to CVRWQCB, USGS and DPR are conducting cooperative synoptic and/or in-season sampling for pesticides, herbicides, and insecticides. The following stations are part of the ongoing studies: San Joaquin River at Vernalis (USGS 11303500), Maze (USGS 11290500), Patterson (USGS 11274570), Crows Landing (USGS 11274550), and Stevinson (USGS 11260815), Bear Creek at Bert Crane Road. (CVRWQCB MER007), Salt Slough at Lander/Hwy 165 (USGS 11261100), Mud Slough (USGS11262900), and Los Banos Creek at Hwy 140 (CVRWQCB MER554). Results of these sampling efforts will help characterize the distribution of pesticides and other toxins within these impaired waterbodies. Annual

reports discussing the results for DPR-funded studies can be found at the following Web site: <http://www.cdpr.ca.gov/docs/emppm/pubs/memos.htm>

Because of their importance as a marker of pesticide-use practices, DDT and two OP pesticides, diazinon and chlorpyrifos are focused on in this document. These compounds, and simazine and metolachlor, were some of the most frequently detected compounds in the NAWQA program studies (Dubrovsky et al. 1998). In addition to the well-known effects of DDT on egg shell thinning and deformities in birds, OP pesticides can affect survival or cause chronic physiological effects on exposed fish via acetylcholinesterase (AChE) enzyme inhibition and induction of heat shock proteins in response to stress. Juvenile Chinook salmon may be more vulnerable to predation and grow less as a result of only brief exposures to AChE-inhibiting pesticides (Eder et al. 2007, Scholz et al. 2000). Recently, there has been a general movement towards the use of pyrethroids instead of OP pesticides in agriculture. High doses of pyrethroid compounds, such as esfenvalerate can be acutely toxic to juvenile Chinook salmon (Wheelock et al. 2005). The ecological effects of increased use of pyrethroids on aquatic ecosystems and Chinook salmon populations are in need of further research (Phillips 2006). Despite the fact that pyrethroids are now one of the most important insecticides and increasingly applied in the Central Valley, primarily for agriculture and urban purposes, only a limited number of studies and monitoring efforts are focusing on occurrence and toxicity (Oros and Werner 2005). There are not enough field monitoring data to date on the spatial and temporal occurrences of pyrethroids for making risk assessments (Oros and Werner 2005).

Trace Elements

Selenium and mercury are two environmental contaminants of primary concern in aquatic environments, and the San Joaquin River is not an exception. Selenium and mercury are trace elements that can be harmful to aquatic life because they undergo biomagnification after being converted to organic forms in reducing (i.e., low oxygen) conditions by methylating bacteria. As a result of this conversion to an organo-metallic compound, methylated selenium and mercury are preferentially absorbed into fatty tissues and can biomagnify through the food chain despite low ambient concentrations. CVRWQCB water quality objectives for selenium are currently being exceeded for Mud Slough and downstream reaches. While the reported background concentrations for selenium for the San Joaquin River above Salt and Mud Sloughs are about 0.5 micrograms per liter ($\mu\text{g/L}$), selected sites along the river have selenium concentrations from 1 to 5 $\mu\text{g/L}$ (CVRWQCB 2001). The input of selenium from the Grasslands area into the San Joaquin River represents a major risk for larval fish, including salmon (Beckon 2007).

Effluent from Wastewater Treatment Plants and Livestock Operations

Free ammonia (NH_3), other nitrogen species (NO_2 , NO_3 , organic nitrogen), pH, chlorine, and DO are a concern in the Delta, particularly near the outflow from sewage treatment plants and dairy farms. One of the most significant water quality problems in the Delta is the low DO problem in the Deepwater Ship Channel near the Port of Stockton. The first 7 miles of the deepwater ship channel west of the Port of Stockton experiences DO concentrations below the CVRWQCB DO water quality standards (SJRDOTWG 2007). The low DO problem is due to poor water circulation and the oxygen demand exerted by

wastewater discharge from the Stockton Regional Wastewater Control facility and the decomposition of algal biomass produced upstream. In response to nutrients discharged by irrigated agriculture and dairy operations in the San Joaquin River basin, high concentrations of planktonic algae grow within 8 to 10 feet of the water's surface upstream from the ship channel and then settle below the sunlight zone and die when the water flows into the 35-foot-deep ship channel (Lee and Jones-Lee 2003). Minimum DO concentrations measured in the San Joaquin River ship channel at the DWR Rough and Ready Island station during April and May typically range between about 3 mg/L during low flows (e.g., 1987) and 7 mg/L during flood conditions (e.g., 1998). DO levels below 3.3 mg/L are considered lethal for salmon whereas levels below 5.0 mg/L may reduce growth rates of juvenile salmon (Spence et al. 1996). Nitrification of even low levels of ammonia as well as decomposition of algal detritus and residual wastewater use large amounts of DO. Other factors that affect DO concentrations in the ship channel include water temperature, atmospheric aeration, and sediment oxygen demand (Jones and Stokes 2002b).

Observed Salmon Mortalities During the 2007 VAMP Studies. It is possible that impaired water quality in the San Joaquin River near Stockton was responsible for the mortality of about 20 percent of tagged juvenile fall-run Chinook salmon during the May 2007 Vernalis Adaptive Management Plan (VAMP) studies. A total of 152 of about 780 juvenile salmon that had surgically inserted acoustic tags and were released in the mainstem San Joaquin River stopped their migrations and presumably died adjacent to a railroad bridge and the Stockton Regional Wastewater Control Facility outfall (Natural Resource Scientists 2007). Initially, 116 dead fish were located on May 17 and 18 (Natural Resource Scientists 2007), whereas another 36 dead fish were located after May 20, 2007. The cause of the mortality remains uncertain because few of the dead fish were recovered, no bioassay studies were conducted in the river near the wastewater facility, and there were no water quality monitoring stations where the dead fish were found. Because of the high concentration of fish tags at this location, either unusually high predator activity or some toxicity event was hypothesized to have resulted in the localized fish mortality.

Potential water quality constituents that may be associated with fish toxicity or mortality of the VAMP study fish in May 2007 include NH_3 , hydrogen sulfide (H_2S), and low levels of OP pesticides (e.g., chlorpyrifos and diazinon). Monitoring of the wastewater control facility's effluent indicated that pH, DO, turbidity, chlorine, and ammonia were within compliance conditions of the facility's permits shortly after the fish had been released.² Monitoring in the river approximately 0.5 miles upstream and downstream from the site also suggest that pH (7.75 to 8.25) and DO (>9 mg/L) levels would not account for the mortality (Mueller-Solger 2007). However, although unionized ammonia levels in the river were less than 0.02 mg/L, well below the U.S. Environmental Protection Agency (1999) critical levels for salmon (e.g., 0.21 mg/L NH_3 at 20°C, and a pH of 8), final effluent grab samples collected by CVRWQCB staff at the Stockton

² Letter from Patricia Leary, California Regional Water Quality Control Board, Central Valley Region, June 20, 2007, to Mark Madison, Director, Department of Municipal Utilities, City of Stockton Regional Wastewater Control Facility.

Regional Wastewater Control Facility contained total ammonia and total Kjeldahl nitrogen (TKN) at levels of 4.4 mg/L and 6.2 mg/L, respectively. Since average daily pH at the Port of Stockton approaches levels (pH 8 or above) that produce acute and chronic ammonia toxicity, and algal photosynthesis in the lower San Joaquin likely produces diel pH swings due to scavenging of carbon dioxide and alkalinity, it is possible that ammonia toxicity to fish occurs at some time of day for several months from spring through fall of each year. Lastly, since low levels of chlorpyrifos and diazinon have been shown to impair olfactory function and predator avoidance in Chinook salmon (Eder et al. 2007, Scholz et al. 2000), and have also been detected in samples collected under the City of Stockton/San Joaquin County Stormwater Program (National Pollutant Discharge Elimination System (NPDES) Permit No. CAS083470), it is possible that organophosphate pesticide exposure played a role in the observed fish mortality due to increased susceptibility to predation.

4.2.5 Entrainment

In 2001, DFG (McBain and Trush 2002) inventoried 95 riparian diversions in the Restoration Area between RM 209 and 267 that were mostly unscreened pumps. The estimated maximum diversion capacity ranged between less than 1 cfs to 63 cfs. Three of these diversions are weir structures just downstream from Friant Dam. The Big Willow Unit Diversion (RM 261.3) is a cobble-type weir that diverts a small amount of water to the Fish Hatchery. The Rank Island Unit is a cobble weir located at RM 260. The Rank Island Unit diverts approximately 5 cfs to property on the north side of the river. The Milburn Unit Diversion is a small concrete-rubble weir located at RM 247.2. A small pump is located just upstream. In addition, Herren and Kawasaki (2001) found 298 and 2,209 diversions in the San Joaquin basin and Delta respectively. More than 95 percent of these diversions were unscreened, and the impacts of these diversions on juvenile Chinook salmon are unknown. No studies have been conducted to determine the entrainment rates at the pumps and weirs in the Restoration Area or downstream in the Delta.

Below the Restoration Area

The irrigation season in the San Joaquin River between Stockton and the Merced River between 1946 and 2002³ has been principally between March and October, with some water diverted in February and November (Hallock and VanWoert 1959, Quinn and Tulloch 2002). DFG estimated that an average of 127,000 acre-feet of water was diverted annually from all diversions in this reach from 1946 to 1955 (Hallock and VanWoert 1959). Quinn and Tulloch (2002) estimated that from 1999 to 2001, annual pumping rates increased to an average of about 154,500 acre-feet at the four largest diversions, which include the Banta-Carbona Irrigation District, West Stanislaus Irrigation District, Patterson Water Company, and El Solyo Water Company.

During 1955, nets were fished in the Banta-Carbona Irrigation District pumps (RM 67.5), El Solyo pumps (RM 82.0), and Patterson Water Companies pumps (RM 104.4) (Hallock and Van Woert 1959). The highest entrainment rates were measured at the Banta-Carbona site in 1955 at about 12 fish per hour. In summer 2002, screens were installed at

Banta-Carbona that appear to be effective at protecting juvenile salmon³. In comparison, the Patterson Water Company pumps entrained about 1.6 juvenile salmon per hour and the El Solyo pumps entrained about 5.2 salmon per hour in 1955. There are no screens at the West Stanislaus Irrigation District, Patterson Water Company, or El Solyo Water Company pumps, although screens are proposed for the Patterson pumps.

Entrainment of juvenile salmon at the Federal (Central Valley Project (CVP)) and State (State Water Project (SWP)) pumping facilities in the Delta is not directly measured but instead estimated as a function of the expanded number of fish salvaged, fish size, and water velocity through the louvers (Foss 2003). For a 2,000 cfs export flow, the efficiency of the louvers for fish larger than 100 mm in length is estimated to be 70 percent and 68 percent at the CVP facilities and SWP facilities, respectively. Louver efficiencies are about 6 percent higher for salmon up to 100 mm in length compared to larger fish. The number of fish salvaged at the louvers is estimated with samples taken at least every 2 hours while water is pumped (Foss 2003). When tagged juvenile fall-run Chinook salmon were released in the San Joaquin River near Mossdale in spring 1992 and 1993, means of 3.3 percent and 0.3 percent were salvaged at the CVP and SWP facilities, without and with a barrier at the Head of the Old River, respectively (Table 4-3).

**Table 4-3.
Number of Tagged Fall-Run Chinook Salmon Smolts from the Feather River Hatchery Released in San Joaquin River at Mossdale in 1992 and 1993, and Salvage Rates**

Release Date	Vernalis Flow (cfs)	CVP and SWP Export Rates (cfs)	HORB Installed	Number Released	Expanded Salvage		Percent Salvaged	
					CVP	SWP	CVP	SWP
04-May-93	4,730	1,494	No	51,937	931	102	1.79%	0.20%
12-May-93	3,770	1,585	No	52,616	1,332	113	2.53%	0.21%
07-Apr-92	1,620	5,682	No	107,103	5,380	71	5.02%	0.07%
13-Apr-92	1,530	1,185	No	103,712	3,385	106	3.26%	0.10%
24-Apr-92	1,070	1,009	Yes	104,739	28	28	0.03%	0.03%
04-May-92	1,480	2,777	Yes	99,717	28	8	0.03%	0.01%
12-May-92	1,020	1,757	Yes	105,385	0	6	0.00%	0.01%

Source: USFWS 2000a.

Key:

cfs = cubic feet per second

CVP = Central Valley Project

HORB = Head of the Old River Barrier

SWP = State Water Project

³ http://www.jonesandstokes.com/news/Tracy_percent20Press.PDF

Most of the juvenile mortality at the Delta pumping facilities is probably due to predation in Clifton Court Forebay and the canals leading to the pumps by nonnative predators such as striped bass, largemouth bass, and sunfishes (*Centrarchidae* spp.). It is assumed that pre-louver predation losses are 15 percent from the trash racks to the louvers at the CVP facilities and 75 percent in Clifton Court Forebay which leads to the SWP facilities (Foss 2003). Some of the acoustically tagged juvenile fall-run Chinook salmon released for the spring 2007 VAMP studies were preyed on by large fish congregated near the trash racks at the CVP pumping facilities (Vogel, unpublished studies).

4.2.6 Degraded In-River Physical Habitat

In Pacific Northwest and California streams, habitat simplification has led to a decrease in the diversity of anadromous salmonid species habitat (NMFS 1996). Habitat simplification may result from blocked gravel recruitment by upstream dams as well as various land-use activities, including gravel extraction, bank revetment, timber harvest, grazing, urbanization, and agriculture.

Gravel Recruitment Blocked by Dams and Levees

Friant Dam has eliminated sediment supply from the upper watershed, and combined with the modified flow regime and land use downstream from Friant Dam, varying degrees of sediment budget imbalance have occurred in downstream reaches. The current paradigm of dam impacts to sediment supply downstream from the dams is that periodic high flow releases from the dam transport sediment stored in the stream bed, and because the sediment supply from the upper watershed is blocked, channel degradation occurs downstream from the dam as alluvial features (bars and riffles) slowly diminish (Collier et al. 1996). Instream gravel mining has exacerbated this sediment deficit in the Restoration Area (McBain and Trush 2002). Local imbalances in sediment supply and transport have caused primarily incision and channel widening with some local aggradation (sedimentation) in the Restoration Area (Cain 1997). Loss of alluvial features in the Restoration Area has contributed to the reduction in frequency of floodplain inundation, which has probably caused a substantial reduction in potential food resources and refuge from predators for juvenile salmonids in the Restoration Area. In addition, channel incision reduces the availability of alternating bars and riffles that juvenile salmon use for feeding and predator avoidance during low flow periods.

Lack of Large Woody Debris

Large quantities of downed trees are a functionally important component of many streams (NMFS 1996). LWD influences channel morphology by affecting longitudinal profile, pool formation, channel pattern and position, and channel geometry. Downstream transport rates of sediment and organic matter are controlled in part by storage of this material behind LWD. LWD affects the formation and distribution of habitat units, provides cover and complexity, and acts as a substrate for biological activity (NMFS 1996). Wood enters streams inhabited by salmonids either directly from adjacent riparian zones or from riparian zones in adjacent nonfish-bearing tributaries. Removal of riparian vegetation and instream woody material (IWM) from the streambank results in the loss of a primary source of overhead and instream cover for juvenile salmonids. The removal of riparian vegetation and IWM and the replacement of natural bank substrates with rock revetment can adversely affect important ecosystem functions.

Living space and food for terrestrial and aquatic invertebrates is lost, eliminating an important food source for juvenile salmonids. Loss of riparian vegetation and soft substrates reduces inputs of organic material to the stream ecosystem in the form of leaves, detritus, and woody debris, which can affect biological production at all trophic levels. The magnitude of these effects depends on the degree to which riparian vegetation and natural substrates are preserved or recovered during the life of the project.

Dikes, Levees, and Bank Revetment

The construction of levees and dikes to convert land for agricultural production tends to channelize riverine habitats and reduces channel migration and avulsion (McBain and Trush 2002). Reduced channel migration has eliminated off-channel habitats, reduced complex side channels, and reduced instream habitat complexity that all serve to provide suitable conditions for juvenile salmonids over a wide range of flow. Agricultural conversion has also directly reduced the amount of floodplains, and levees and dikes have further isolated historic floodplains from the channel. It is likely that the loss of floodplain habitats has substantially reduced food resources and refuge from predators for juvenile salmonids.

Angular rock (riprap) is used to armor the streambanks from erosive forces in the Restoration Area and throughout the Central Valley. Simple slopes protected with rock revetment generally create nearshore hydraulic conditions characterized by greater depths and faster, more homogeneous water velocities than occur along natural banks (USFWS 2000b, Garland et al. 2002). Higher water velocities typically inhibit deposition and retention of sediment and woody debris. These changes generally reduce the range of habitat conditions typically found along natural shorelines, especially by eliminating the shallow, slow-velocity river margins used by juvenile fish as refuge and escape from fast currents, deep water, and predators (USFWS 2000b).

The use of rock armoring also limits recruitment of IWM and greatly reduces, if not eliminates, the retention of IWM once it enters the river channel. Riprapping creates a relatively clean, smooth surface that diminishes the ability of IWM to become securely snagged and anchored by sediment. IWM tends to become only temporarily snagged along riprap, and generally moves downstream with subsequent high flows. Habitat value and ecological functioning aspects are thus greatly reduced, because wood needs to remain in place to generate maximum values to fish and wildlife (USFWS 2000b). Recruitment of IWM is limited to any eventual, long-term tree mortality and any abrasion and breakage that may occur during high flows (USFWS 2000b). Juvenile salmonids are likely being impacted by reductions, fragmentation, and general lack of connectedness of remaining nearshore refuge areas.

A separate but connected bypass system, consisting of the Chowchilla Bypass Channel, Eastside Bypass Channel, and Mariposa Bypass Channel, was constructed to divert and carry flood flows from the San Joaquin River and eastside tributaries upstream from the Merced River. These bypasses lack floodplain access, habitat structure, nearshore habitat and riparian habitat required by Chinook salmon.

Urbanization

CALFED (2000) estimated that wetted perimeter reductions in the Delta have decreased from between 25 and 45 percent since 1906. Historically, the San Francisco Estuary included more than 242,000 acres of tidally influenced bay-land habitats, and tidal marsh and tidal flats accounted for 98 percent of bay-land habitats. Today, only 70,000 acres of tidally influenced habitat remain (CALFED 2000). While historical uses of riparian areas (e.g., wood cutting, clearing for agricultural uses) have substantially decreased, urbanization still poses a serious threat to remaining riparian areas. Riversides are desirable places to locate homes, businesses, and industry.

4.2.7 High Water Temperatures

Juvenile Chinook salmon will grow at water temperatures ranging from 46.4°F to 77°F (8.0°C to 25°C) (Brett et al. 1982, Clarke and Shelbourn 1985) with optimum growth rates occurring at about 66.2°F (19.0°C) when fed maximal rations (Myrick and Cech 2001). However, juveniles exposed to low DO concentrations and pathogens may require even lower water temperatures for optimal growth rates (Myrick and Cech 2001). For example, Rich (1987), who reared juvenile salmon in surface water collected from the American River, determined that optimal growth rates for American River fish occurred at 59.5°F (15.3°C). It is also likely that the optimal temperature for growth is further reduced when food rations are low (Brett et al. 1969). Based on a model developed for sockeye salmon (*O. nerka*), Brett et al. (1982) determined that temperatures between 66.0°F and 68.9°F (18.9°C to 20.5°C) produced optimal growth for fish fed the maximum rations, but that temperatures of about 59°F (15°C) produced optimal growth for fish fed 60 percent of the maximum food rations.

Release temperatures from Friant Dam currently range from 48°F to 58°F (8.9°C to 14.4°C) and water temperatures are expected to be suitable for juvenile rearing except in the downstream reaches as air temperatures increase.

Unsuitably high water temperatures and exaggerated fluctuations in water temperature result from a combination of factors, including seasonally high air temperatures (May and June), low flow releases, groundwater pumping that eliminated the inflow of cool groundwater throughout the Restoration Area, removal of large woody riparian forests that provided shade, warm agricultural runoff, and warm flood flows from the Kings River through the James Bypass. It is also possible that high flow releases during summer and fall could exhaust the cold water pool in Millerton Lake and thereby cause release temperatures to substantially increase above 58°F (14.4°C).

Delta Conditions

Currently, there are no flow or water temperature standards to maintain suitable habitat for juvenile salmon in the lower San Joaquin River. Water temperatures in the San Joaquin River near Vernalis (DWR gage data) were usually below 65°F (18.3°C) from mid-April to mid-May when Vernalis flows were at least 3,500 cfs. Springtime water temperatures at Vernalis exceeded 65°F (18.3°C) during drought years (e.g., 1977 and 1987 to 1992) and when high flows entered the San Joaquin River from the James Bypass upstream from Newman during spring 1986. By the end of May, water temperatures typically ranged between 65°F and 70°F regardless of flow.

4.2.8 Harvest of Yearling-Sized Juveniles

Following reintroduction of spring-run Chinook salmon into the San Joaquin River, yearling salmon may be present in portions of the Restoration Area throughout the year. Yearling spring-run Chinook salmon (those adopting a stream-type life history strategy) typically range in length from about 80 to 150 mm (3 to 6 inches), depending on growth rate and freshwater residence time (Moyle 2002). Sport anglers may catch yearling Chinook salmon while fishing for trout or other game fish, likely resulting in injury or mortality due to hooking and handling. State fishing regulations specify bag limits for trout and salmon in the San Joaquin River, but size restrictions are not designated (DFG 2007b).

4.3 Ocean Phase

The survival of smolts entering the ocean during June and July is probably the most critical phase for salmon in the ocean (Pearcy 1992, Mantua et al. 1997, Quinn 2005). Marking studies suggest that about 59 to 77 percent of juvenile pink salmon (*O. gorbuscha*) died in their first 40 days at sea off the coast of British Columbia, whereas 78 to 95 percent of those that survived their first 40 days died over the next 410 days at sea (Parker 1968). Another marking study with chum salmon (*O. keta*) off the coast of Washington indicated that juvenile mortality averaged 31 to 46 percent per day during the first few days (Bax 1983).

The survival of smolts entering the ocean is highly correlated with ocean productivity as affected by freshwater outflow from the estuary. This, in turn, affects the availability of food resources at the interface between freshwater and saltwater, as well as coastal upwelling, ocean currents and El Niño events (Casillas 2007).

4.3.1 Inadequate Juvenile Food Availability

Long-term records indicate that there are 15- to 25-year cycles of warm and cool periods that strongly correlate with marine ecosystem productivity (Mantua et al. 1997; Hollowed et al. 2001). Cool productive cycles prevailed from 1947 through 1976, and a new cycle began in 1998, whereas warm unproductive cycles dominated from 1925 through 1946, and from 1977 through 1997 (Mantua et al. 1997; Mantua and Hare 2002). The coastal warming that occurred in the mid-1970s is believed to have caused increased stratification in the California Current; a sharper thermocline with less upwelling of nutrient-rich water; a reduction in the duration of upwelling; and a reduction in nutrients and/or zooplankton abundance carried by the California Current (Francis et al. 1998). In addition, the abundance of coastal euphausiids (*Thysanoessa spinifera*) declined whereas oceanic euphausiids (*T. pacifica*) increased (Francis et al. 1998). Such changes are thought to affect salmon early in their marine life history (Hare and Francis 1995), and coastal invertebrate species are important prey for ocean-type juveniles, such as Central Valley fall-run Chinook salmon.

The interface between the plume of freshwater outflow from the Columbia River and saltwater in the ocean is a highly productive area that is important to the survival of juvenile Chinook salmon and other salmonid species migrating into the ocean (Casillas

2007). Large freshwater plumes that extend well offshore 7 to 10 days after juvenile salmonids enter the ocean are highly correlated with higher numbers of returning adults 2 years later (Casillas 2007). The density of food organisms, particularly crustacean larvae, is unusually high at the freshwater-saltwater interface. It is likely that freshwater outflow from the San Francisco Estuary between May and July is also important to the survival of juvenile San Joaquin River salmon. The May through July period is probably important because that is when juvenile Chinook salmon entered the Gulf of the Farallones during spring 1997 (MacFarlane and Norton 2002). In the Gulf of the Farallones, the size of the plume would be controlled by inflow to the Delta from the Sacramento and San Joaquin basins as well as Delta exports, which can be as high as 35 percent of Delta inflow from February through June, and 65 percent of Delta inflow from July through January (SWRCB 1995 and draft 2006 Water Quality Control Plans for the San Francisco Bay/Sacramento-San Joaquin Delta Estuary).

Indicators of Ocean Productivity

Coastal waters off the Pacific Northwest are influenced by atmospheric conditions in the North Pacific Ocean, but also in equatorial waters, especially during El Niño events. Strong El Niño events result in the transport of warm equatorial waters northward along the coasts of Central America, Mexico, and California, and into the coastal waters off Oregon and Washington. These events affect weather in the Pacific Northwest, often resulting in stronger winter storms and transport of warm, offshore waters into the coastal zone. The transport of warm waters toward the coast, either from the south or from offshore, also results in the presence of unusual mixes of zooplankton and fish species.

The Pacific Decadal Oscillation (PDO) is a climate index based on patterns of variation in sea surface temperature of the North Pacific from 1900 to the present (Mantua et al. 1997). While derived from sea surface temperature data, the PDO index is well correlated with many records of North Pacific and Pacific Northwest climate and ecology, including sea level pressure, winter land-surface temperature and precipitation, and stream flow. The index is also correlated with salmon landings from Alaska, Washington, Oregon, and California.

Since 1955, the presence/absence of conditions resulting from the El Niño Southern Oscillation (ENSO) has been gauged using the Multivariate ENSO Index (MEI). Prior to 1977 (during the cool phase of the PDO), El Niño conditions were observed infrequently.

Both the PDO and MEI can be viewed as "leading indicators" of ocean productivity because after a persistent change in sign of either index, ocean conditions in the California Current soon begin to change. Most recently, in September 2005, the MEI appears to have signaled a return to warmer ocean conditions.

4.3.2 Marine Predation

Both bird and fish predators congregate at the freshwater-saltwater interface of the freshwater plume of the Columbia River where juvenile salmon feed (Casillas 2007). In spring 2003, there were many species of bird predators, including the common murre, sooty shearwater, Sabine's gull, western gull, rhinoceros auklet, black-foot albatross, and

red phalarope. In 2003, the common murre was the most common bird predator in May (47 percent), whereas the sooty shearwater was the most common bird predator in June (68 percent). Marine fish that intensively prey on juvenile salmon include Pacific hake (*Merluccius productus*), rockfish (*Sebastes spp.*), and to a lesser degree, jack mackerel (*Trachurus symmetricus*), Pacific mackerel (*Scombrus japonicus*), and spiny dogfish (*Squalus acanthias*). The abundance of bird and fish predators has been highly correlated with juvenile salmon abundance off the coast of Washington. However, the impact of predation on the number of returning adult salmon has not been quantified.

The primary marine mammals preying on salmonids are pinnipeds, including harbor seals (*Phoca vitulina*), California sea lions (*Zalophus californianus*), and Steller's sea lions (*Eumetopia jubatus*) (Spence et al. 1996). Pacific striped dolphin (*Lagenorhynchus obliquidens*) and killer whale (*Orcinus orca*) also prey on adult salmonids in the nearshore marine environment. Seal and sea lion predation is primarily in saltwater and estuarine environments, although they are known to travel well into freshwater after migrating fish. All of these predators are opportunists, searching out locations where juveniles and adults are most vulnerable.

4.3.3 Adult Commercial and Sport Harvest

Extensive ocean recreational and commercial troll fisheries for Chinook salmon exist along the Central California coast, and an inland recreational fishery exists in the Central Valley for Chinook salmon and steelhead. Ocean harvest of Central Valley Chinook salmon is estimated using an abundance index, called the Central Valley Index (CVI). The CVI is the ratio of Chinook salmon harvested south of Point Arena (where 85 percent of Central Valley Chinook salmon are caught) to the sum of the estimated escapements and harvest of Central Valley fish.

Ocean fisheries have affected the age structure of Central Valley spring-run Chinook salmon through targeting large fish for many years and reducing the numbers of 4- and 5-year-old fish (DFG 1998). Ocean harvest rates of Central Valley spring-run Chinook salmon are thought to be a function of the CVI (Good et al. 2005). Harvest rates of Central Valley spring-run Chinook salmon ranged from 55 percent to nearly 80 percent between 1970 and 1995, when harvest rates were adjusted to protect Sacramento River winter-run Chinook salmon. The drop in the CVI in 2001 as a result of high fall-run escapement to 27 percent also reduced harvest of Central Valley spring-run Chinook salmon.

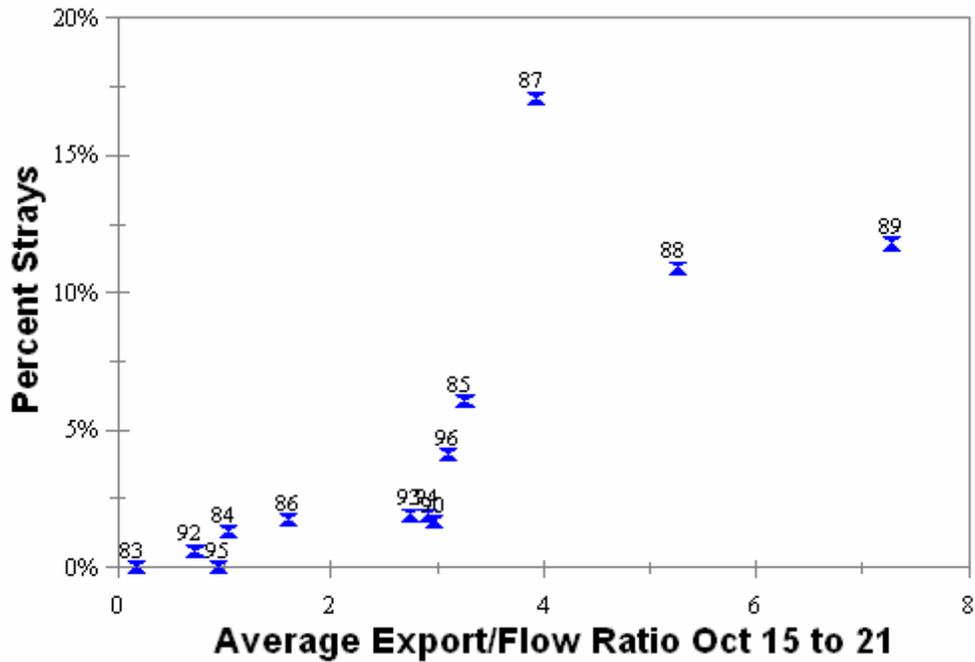
In-river recreational fisheries historically have taken Central Valley spring-run Chinook salmon throughout the species' range. During the summer, holding adult Central Valley spring-run Chinook salmon are targeted by anglers when they congregate in large pools. Poaching also occurs at fish ladders, and other areas where adults congregate; however, the significance of poaching on the adult population is unknown. Specific regulations for the protection of Central Valley spring-run Chinook salmon in Mill, Deer, Butte, and Big Chico creeks were added to existing DFG regulations in 1994. The current regulations, including those developed for Sacramento River winter-run Chinook salmon, provide some level of protection for spring-run fish (DFG 1998).

4.4 Adult Migration

Adult salmon will have to navigate approximately 270 miles from the ocean to their spawning habitat below Friant Dam. The number of Chinook salmon that successfully complete their migration will partly depend on environmental conditions that are needed for the fish to home to their natal stream as well as other factors, such as predation and harvest, that result in mortality.

4.4.1 Inadequate Flows and High Delta Export Rates

An important factor for successful upstream migration is sufficient flow throughout the migratory corridor that provide olfactory cues that allow the adult salmon to home to their natal stream. This has been a concern for adult fall-run Chinook salmon in the San Joaquin River basin since 1996 when Delta export rates at the CVP and SWP were increased to near maximum (about 9,600 cfs) to “make up” for reduced pumping rates during the spring outmigration period. When exports are high relative to San Joaquin River flows, it is likely that little, if any, San Joaquin River water reaches the San Francisco Bay where it may be needed to help guide the salmon back to their natal stream. An analysis by Mesick (2001b) of recovered adult salmon with coded wire tags (CWT) that had been reared at the Merced River Fish Facility and released in one of the San Joaquin tributaries suggests straying occurred when the ratio of exports to flows was high. The analysis indicates that during mid-October from 1987 through 1989, when export rates exceeded 400 percent of Vernalis flows, straying rates ranged between 11 percent and 17 percent (Figure 4-3). In contrast, straying rates were estimated to be less than 3 percent when Delta export rates were less than about 300 percent of San Joaquin River flow at Vernalis during mid-October.



Source: Mesick 2001b.

Notes:

1. Juveniles were released in the San Joaquin River basin and subsequently strayed to the Sacramento River and eastside tributary basins to spawn.
2. Average Export/Flow Ratio is based on the average ratio of the export rate at the CVP and SWP pumping facilities in the Delta compared to the flow rate in the San Joaquin River at Vernalis between 15 and 21 October, from 1983 to 1996.

Figure 4-3.
Estimated Percent of Adult Merced River Hatchery CWT Chinook Salmon Strays
Relative to Export to Flow Ratio

4.4.2 High Water Temperatures

In general, Chinook salmon appear capable of migrating upstream under a wide range of temperatures. Bell (1986) reported that salmon migrate upstream in water temperatures that range from 37°F (2.8°C) to 68°F (20°C). Bell (1986) reports that temperatures ranging between 37°F (2.8°C) and 55°F (12.8°C) are suitable for upstream migration of spring-run Chinook salmon, and between 50°F (10°C) and 66°F (18.9°C) for fall-run Chinook salmon. Marine (1992) reported a water temperature range of 43°F (6.1°C) to 57°F (13.9°C) as optimal for survival, maturation, and spawning for adult Chinook salmon.

4.4.3 Physical Barriers and Flow Diversion

Historically, adult spring-run Chinook salmon migrated as far upstream as Graveyard Meadows (Lee 1998). The amount of holding and spawning habitat available to spring-run Chinook salmon was reduced around 1920, when Kerckhoff Dam “blocked the spring-run salmon from their spawning areas upstream and seasonally reduced flows in about 14 miles of stream, below the dam, where there were pools in which the fish would have held over the summer” (DFG 1921, as cited in Yoshiyama et al. 1996). The completion of Friant Dam in 1941 blocked access to an additional 16 miles of habitat that was historically used by spring-run Chinook salmon for spawning, representing an estimated 36 percent loss of the historic spawning habitat (Hatton 1940, as cited in Yoshiyama et al. 1996).

Passage below Friant Dam during the 1940s was inhibited by low flows in the channel. In 1944 and 1947, DFG (1955) observed from 5,000 to 6,000 spring-run Chinook salmon migrating up the San Joaquin River as far as Mendota Dam in a flow that was estimated to be 100 cfs in the reach between Sack Dam and the confluence with the Merced River. DFG observed that “many of these fish have rubbed themselves raw going over the shallow sandbars” between Sack Dam and the confluence with the Merced River (a distance of approximately 50 miles). Such abrasions can increase the risk of mortality from disease for spring-run Chinook salmon, since they must hold in pools throughout the summer before spawning. Passage for the San Joaquin River adult spring-run salmon has been completely blocked in the Restoration Area since the 1950s, when the river was dewatered below Sack Dam except during uncontrolled flow releases in wet years.

The Settlement prescribes that passage will be restored at all structures that may impede the passage of adult Chinook salmon through the Restoration Area. Improvements will be made at the following structures during Phase 1:

- Mendota Dam – A bypass channel will be created around Mendota Pool (RM 205)
- Reach 4B headgate and Sand Slough control structures (RM 168.5)
- Arroyo Canal Water Diversion – Screens will be installed (RM 182)
- Sack Dam, a diversion dam for the Arroyo Canal (RM 182)
- Eastside Bypass structures (RM 138 and RM 168)

- Mariposa Bypass structures (RM 147.2)
- Salt and Mud sloughs – Seasonal barriers will be installed to prevent adult salmon from entering these false migration pathways

Improvements will be made at Chowchilla Bifurcation Structure (RM 216) during Phase 2. McBain and Trush (2002) identified at least one earthen diversion dam just downstream from Gravelly Ford (RM 227) that may be potential impediments to both upstream and downstream fish movement.

4.4.4 Delta Water Quality

Hallock et al. (1970) showed that radio-tagged adult fall-run Chinook salmon delayed their migration at Stockton whenever DO concentrations were less than 5 mg/L and/or water temperatures exceeded about 65°F (18.3°C) in October. Delaying the migration of adult fall-run Chinook salmon in the deepwater ship channel near Stockton may reduce gamete viability if the fish are exposed to high temperatures for prolonged periods. DFG reports that the quality and survival of eggs was poor from females exposed to water temperatures that exceeded 56 °F (13.3°C) (DFG 1992).

DO concentrations near Stockton in October were greater than 5 mg/L from 1983, when DWR began monitoring, to 1990, but were lower than 5 mg/L for most of October in 1991 and 1992. The Head of the Old River Barrier was installed in fall 1992, but it did not correct the problem until late October (Figure 4-4). In 1993, DO levels were low until about October 10, and it is likely that pulse flows that raised Vernalis flows to about 4,000 cfs on October 7 were responsible for increasing DO levels at Stockton (Figure 4-4). Similarly in 1994, DO levels were low until October 15, when pulse flows raised Vernalis flows to about 2,000 cfs (Figure 4-4). In 1995, DO levels were at least 6 mg/L in October when Vernalis flows ranged from about 3,000 cfs to 6,000 cfs through mid-October. DO levels were low or fluctuated greatly in 1996 until October 13, when pulse flow releases increased Vernalis flows from 2,000 to about 3,000 cfs (Figure 4-4).

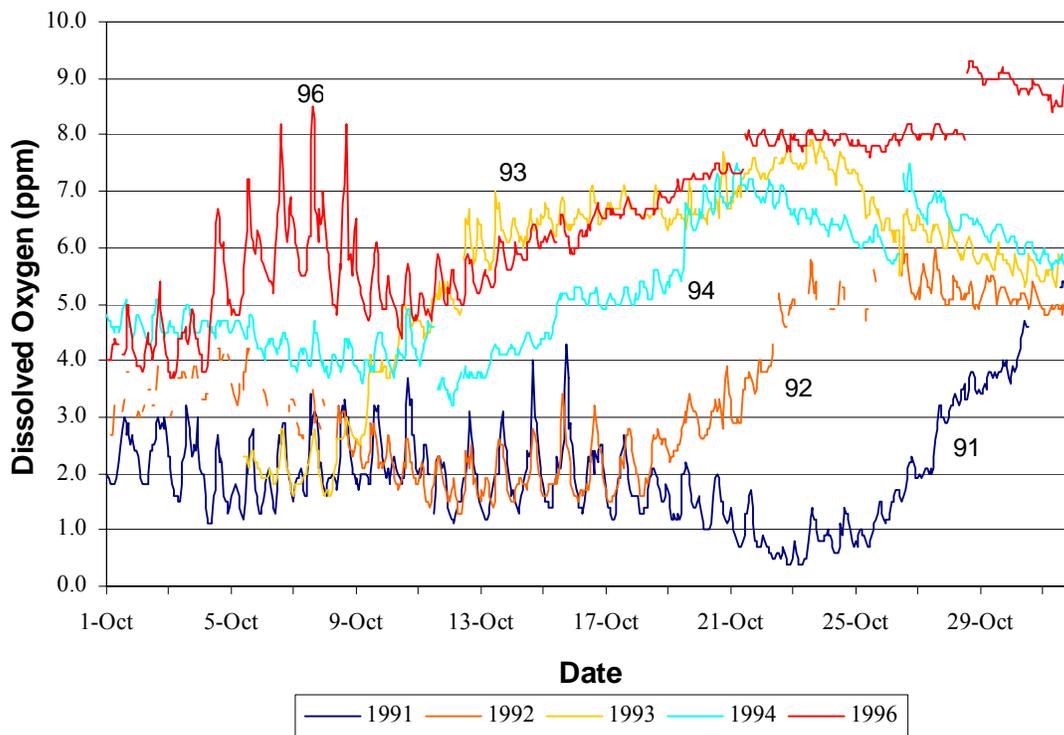


Figure 4-4.
Hourly Dissolved Oxygen Measurements at Burns Cut Off Road Monitoring Station During October in 1991 Through 1994 and 1996.

4.4.5 In-River Harvest

During the 1940s, DFG (1946) reported that low flows resulted in high rates of harvest and incidental mortality from spearing in the lower river. In 1944, approximately 200 people were observed spearing salmon at each sand bar in the lower river. Some people used pitch forks, which wounded many fish that probably died before spawning (DFG 1946). Although spearing is no longer legal, the illegal poaching of adult salmon will still be a concern.

Current bag limits specified by State fishing regulations allow legal catch throughout the year of one salmon in the San Joaquin River from Friant Dam downstream to the Highway 140 Bridge (DFG 2007b). Size restrictions, however, are not designated for salmon in any portion of the San Joaquin River. Downstream from the Highway 140 Bridge, one salmon may be harvested from January through October. During November and December, a zero bag limit for salmon is enforced downstream from the Highway 140 Bridge that requires any salmon caught during these months to be unharmed and not removed from the water.

4.5 Adult Holding

When adult spring-run Chinook salmon begin their migration to their natal streams, they are sexually immature, unable to spawn. After they arrive in their natal streams in the spring, they hold in deep pools through the summer, conserving energy until the fall when their gonads ripen and they spawn. Fall Chinook salmon generally do not hold in pools for long periods of time (>1 week), but they may briefly use large resting pools during upstream migration.

4.5.1 Historical Habitat in the San Joaquin River

Adult spring-run Chinook salmon held in pools above Friant Dam before its construction (DFG 1921, as cited in Yoshiyama et al. 1996), probably as far upstream as Mammoth Pool Reservoir (Yoshiyama et al. 1996). Hatton described “long, deep pools” in the canyon above Friant (1940, as cited in Yoshiyama et al. 1996). The amount of holding and spawning habitat available to spring-run Chinook salmon was reduced around 1920, when Kerckhoff Dam “blocked the spring-run salmon from their spawning areas upstream and seasonally dried up about 14 miles (22.5 km) of stream, below the dam, where there were pools in which the fish would have held over the summer” (DFG 1921, as cited in Yoshiyama et al. 1996). The completion of Friant Dam in 1941 further reduced the holding and spawning habitat available to spring-run Chinook salmon by completely blocking access to upstream areas.

4.5.2 Habitat Below Friant Dam

In July 1942, Clark (1943) observed an estimated 5,000 adult spring-run Chinook salmon holding in two large pools directly downstream from Friant Dam. He reported that the fish appeared to be in good condition, and that they held in large, quiet schools. Flow from the dam was approximately 1,500 cfs, and water temperatures reached a maximum of 72°F (22.2°C) in July. Several hundred yards downstream, there is another pool that has a maximum depth of 25 feet (8 meters) with an average depth of 11 feet (3 meters), with an approximate area of average depth of 93,000 square feet (8,600 square meters) (Stillwater Sciences 2003). Chinook generally do not feed while they hold; therefore, they can hold at very high densities. It is likely that these pools can hold up to about 20,000 adult spring-run Chinook salmon.

Although some fish may have held in pools downstream from Lanes Bridge, Clark (1943) concluded that the abundant spawning he observed in September and October in riffles between Friant Dam and Lanes Bridge were from fish holding in the pools below the dam that had moved back downstream to spawn.

4.5.3 Harvest

Current bag limits specified by State fishing regulations allow legal catch throughout the year of one salmon in the San Joaquin River from Friant Dam downstream to the Highway 140 Bridge (DFG 2007b).

Poaching of holding spring-run salmon remains a concern because fish are vulnerable for several months in a confined location at high densities. The banks of the pool below Friant Dam are fenced off, thus minimizing access for poachers. However, the North Fork Road Bridge downstream from the dam has a boat launch that provides access to the river where poachers could gain access to the pool.

4.5.4 High Water Temperatures

Water temperatures for adult Chinook salmon holding are reportedly optimal when less than 60.8°F (16°C), and lethal when above 80.6°F (27°C) (Moyle et al. 1995). Moyle et al. (1995) reported that spring-run Chinook salmon in the Sacramento River typically hold in pools that have temperatures below 69.8°F (21°C) to 77°F (25°C). In Butte Creek, prespawn adult mortalities were minimal when average daily temperatures were less than 66.9°F (19.4°C) with only brief periods of high temperatures up to about 70°F (21°C) in July between 2001 and 2004 (Ward et al. 2006). In 2003, 11,000 adults died before spawning, while over 6,000 survived to spawn in Butte Creek. Mortalities were attributed to high temperatures, large numbers of fish and outbreaks of two pathogens, Columnaris and Ich. Average daily temperatures exceeded 59°F (15°C) at all sites from late-June until the first week of September, exceeded 63.5°F (17.5°C) by July 12, and exceeded 68°F (20°C) for 7 days during the holding period at the uppermost holding pool (Quartz Bowl) in 2003 (Ward et al. 2004).

4.5.5 Disease

Diseases such as BKD, *Ceratomyxosis shasta* (C-shasta), columnaris, furunculosis, infectious hematopoietic necrosis, redmouth and black spot disease, whirling disease, and erythrocytic inclusion body syndrome are known, among others, to affect Chinook salmon (NMFS 1996, 1998). Many pathogens are ubiquitous along the northwestern Pacific coast of the United States in salmon populations. However, the pathogens are normally present at low levels and do not usually affect the host to the point of causing disease (Arkoosh 1998). Only when other stressors are present are there increased incidences of disease outbreaks. These stressors can include elevated water temperature, low DO, crowding, high levels of ammonia, and presence of pollutants (Wedemeyer 1974). The susceptibility of anadromous salmonids to these pathogens is also influenced by hydrological regime, behavior, and physiological changes associated with spawning activity.

Two extreme cases of disease-related fish kills occurred in the Klamath River and Butte Creek in 2003. In September 2002, 34,000 adult salmon, mostly Chinook, died in the lower 25 miles of the Klamath River, California due to a combination of low flows, high temperatures, and high infestation rates of Ich (*Ichthyophthirius multifiliis*) and/or columnaris (*Flavobacterium columnare*). Significant prespawning mortality of spring-run Chinook salmon also occurred in Butte Creek, California, during 2003 as a result of high temperatures and subsequent infection of columnaris and Ich (Ward et al. 2006).

4.5.6 Predation

Mammals may be an agent of mortality to salmonids in the Central Valley. Predators such as river otters (*Lutra Canadensis*), raccoons (*Procyon lotor*), striped skunk (*Mephitis mephitis*), and western spotted skunk (*Spilogale gracilis*) are common. Other mammals that take salmonids include badger (*Taxidea taxus*), bobcat (*Linx rufis*), coyote (*Canis latrans*), gray fox (*Urocyon cinereoargenteus*), long-tailed weasel (*Mustela frenata*), mink (*Mustela vison*), mountain lion (*Felis concolor*), red fox (*Vulpes vulpes*), and ringtail (*Bassariscus astutus*). These animals, especially river otters, are capable of removing large numbers of salmon and trout (Dolloff 1993). Mammals have the potential to consume large numbers of holding adults, but generally scavenge post-spawned salmon.

4.6 Spawning

Clark (1943) estimated that about 267,000 square feet (64 percent) of spawning habitat remained after Friant Dam had been constructed in 1941. Salmon were observed spawning in large numbers on all the riffles in the 10-mile reach between Friant Dam and Lanes Bridge in 1942. Since the 1940s, spawning habitat has been highly degraded by dams that block gravel recruitment, in-river gold and gravel mining, and water diversions that reduce flows and increase water temperatures. It is assumed that the Restoration Hydrographs will provide suitable water depths and velocities for spawning based on a Physical Habitat Simulation study conducted by USFWS in 1993.

4.6.1 Lack of Spawning-Sized Gravels

The abundance of spawning-sized gravels below Friant Dam has gradually decreased as a result of upstream dams blocking sediment recruitment and gravel mining from the river terrace and low-flow channel. The estimated average unimpaired coarse sediment supply for the mainstem San Joaquin River is approximately 48,600 cubic yards/year (Cain 1997). There is relatively little gravel recruitment from the tributaries below Friant Dam: Cottonwood Creek (RM 267.4) contributes about 55 cubic yards/year and Little Dry Creek (RM 261) contributes an average of about 335 cubic yards/year (Cain 1997).

An absence of gravel recruitment tends to reduce the amount of useable spawning habitat in three ways. First, without recruitment, uncontrolled high flow releases scour the gravel from the spawning beds so that they gradually become smaller in length and the depth of the gravel becomes shallower. Cain (1997) compared the 1939 and 1996 measurements of the channel thalweg elevation at seven cross sections in Reach 1A: at four cross sections, the thalweg elevation decreased by 4.5 to 7.0 feet whereas it increased by 0.8 to 3.2 feet at three cross sections. Second, smaller gravels tend to be mobilized at the highest rates, which causes the bed surface to armor with large rocks that can be too large for the salmon to move for redd construction. Both the reduction in spawning bed size and the armoring of the bed's surface has the effect of crowding spawners into the remaining usable spawning areas. Crowding is thought to increase the rate of redd superimposition, when spawners construct their redds on top of preexisting redds, thereby killing or burying some of the eggs in the pre-existing redds. The third

problem caused by reduced gravel recruitment is that uncontrolled scouring flows also erode sediment from the floodplains.

A reduction in upstream gravel supply can disrupt the balance between sediment supply and transport capacity, disturbing the longitudinal continuity of the river system and altering channel pattern (Kondolf and Swanson 1993, Kondolf 1997). The excess energy of sediment-starved water is typically expended on the bed, causing incision and likely channel narrowing. Sediment-starved channels can also respond through lateral migration into banks and floodplains, potentially causing greater rates of bank failure as the channel pattern adjusts to a new sediment supply and transport equilibrium (Simon 1995). Channel widening is a problem in some reaches of the Stanislaus River (Schneider 1999) and it appears to be a problem in Reach 1 of the Restoration Area (FMWG 2007). Bank erosion degrades the spawning habitat by reducing water depths and velocities and degrades the egg incubation habitat by increasing the rate that fine sediments are deposited on the spawning beds.

Instream aggregate extraction may have further reduced the amount of spawning-sized gravel in Reach 1A, where a majority of the salmon will be expected to spawn. In Reach 1A, Cain (1997) estimated that 1,562,000 cubic yards were removed from the active channel of the San Joaquin River between 1939 and 1989 (3,124 cubic yards/year), and 3,103,000 cubic yards were removed from the floodplain and terraces. Nine large captured mine pits occur from about 8.7 miles (RM 258.8) to about 34.3 miles (RM 233.2) below Friant Dam (Table 3-16 in McBain and Trush 2002); therefore, it is likely that many spawning beds were highly degraded by gravel mining.

During July 2007, the Fisheries Management Work Group observed one spawning bed with suitably sized gravels near the dam and three highly silted spawning beds during foot and canoe surveys of the first 5 miles of the low-flow channel below Friant Dam (RM 262.5 to RM 267.5) where a majority of the spring-run salmon would be expected to spawn. They also observed 22 potential spawning beds in the next 4.4-mile-long reach (RM 257.75 to RM 262.15) that had moderate levels of silt and suitably sized gravels for spawning. The median diameter (D_{50}) of the surface substrate at three of these riffles ranged between 40 and 47 mm based on pebble counts by Stillwater Sciences in 2002 (Table 3-7 in McBain and Trush 2002).

4.6.2 High Water Temperatures

Preferred spawning temperatures for spring-run and fall-run Chinook salmon are between 42°F (5.6°C) and 57°F (13.9°C) (Bell 1986). Temperatures above the preferred spawning range have been observed to increase the occurrence of abnormal fry and mortality, and lengthen the duration of the hatching period (Spence et al. 1996).

4.6.3 Hybridization Between Spring-Run and Fall-Run Salmon

Historically, spring-run Chinook salmon spawned in the upper watersheds whereas fall-run salmon were confined to the lower watersheds when fall flows dropped and barriers prevented their migration to the areas used by the spring-run salmon. Currently, with access to historical higher elevation spring-run spawning habitat blocked by Friant Dam, both runs would share the available spawning habitat downstream from Friant Dam,

posing the risk of hybridization. Forced coexistence of these two runs caused by substantial damming and loss of habitat in other river systems has led to concern for their genetic integrity (Cope and Slater 1957, Banks et al. 2000). However, despite spatial and temporal overlap of Chinook salmon spawning runs in the Central Valley, no evidence for natural hybridization among runs has been documented (Banks et al. 2000).

Genetic effects of run hybridization on Chinook salmon populations remain unclear. It is likely, however, that hybridization between spring-run and fall-run Chinook salmon in the San Joaquin River would influence the life history strategy adopted by hybrid progeny. Given the potential for water temperatures in large portions of the Restoration Area to exceed suitable limits during key periods of upstream migration (late summer and fall) and rearing (spring and early summer), altered run timing is of particular concern. To prevent spawning overlap by the two runs, it may be necessary to construct artificial barriers to separate spring-run and fall-run spawners.

4.6.4 Instream Flows

The relationship between instream flow and spawning habitat availability was modeled by USFWS (1994b). Although the study assessed spawning habitat availability for fall-run Chinook salmon, the authors agree that the relationships can be transferable to spring-run Chinook salmon. USFWS (1994b) found stream flows of 150 cfs to provide close to optimal spawning conditions in Reach 1A. Settlement flows for incubation range from 120 cfs to 350 cfs depending on water year type (Stipulation of Settlement, Exhibit B). Settlement flows appear adequate for incubation and emergence; however, this information should be taken cautiously, as it is extrapolated from fall-run Chinook salmon work conducted in 1993.

4.6.5 Harvest

Currently, fishing regulations in the San Joaquin River permit the harvest of one salmon year-round from Friant Dam downstream to the Highway 140 Bridge (Section 4.4.5); therefore, a majority of the spawning adults should be protected.

Poaching of adult fall-run Chinook salmon from their spawning beds is a common occurrence in the Stanislaus, Tuolumne, and Merced rivers based on reports from DFG wardens; however, the number of adult fish taken has not been estimated. Most poachers snag fish with large treble hooks, but others use gill nets to catch fish. It is likely that spring-run and fall-run salmon will be illegally harvested from the Restoration Area, but the likely extent of the problem in the Restoration Area is unknown.

4.7 Hatchery Impacts

Although the goal of the SJRRP is to restore naturally producing and self-sustaining populations of Chinook salmon and native fish species, salmon hatcheries may be used to help achieve this goal in three ways. First, spring-run Chinook salmon may be stocked with fish that are incubated and/or raised in a hatchery prior to release in the Restoration Area. Second, it is likely that large numbers of study fish will be needed for juvenile salmon survival studies and for calibrating rotary screw traps. Third, if monitoring determines that the natural production of juvenile salmon is too low during the relatively dry water year types (e.g., Critical Low and Critical High year types) when spring flows are either absent, or inadequate to sustain the Chinook salmon populations, hatchery fish may be used to supplement the population in those years. A long-term source of eggs for the hatchery will have to be identified to avoid sacrificing naturally produced San Joaquin River adult salmon.

By ensuring high survival of eggs, fry, and juveniles that typically experience high mortality in the wild, salmon hatcheries can produce more returning adults than would have occurred in the wild. The benefits of hatcheries include the continuation of harvest, reduction of anthropogenic impacts, and recovery of native stocks. In addition, hatcheries can help with native recovery by minimizing short-term extinction risks to depressed populations, maintaining safe population levels, and providing additional fish to help recovery.

The specific goals of each hatchery are as numerous as there are hatcheries; however, the goals can be placed into three categories:

- Hatcheries that strive to produce fish for harvest purposes, generally due to loss of habitat and mitigation
- Hatcheries that strive to conserve or recover depressed populations of salmonids
- Hatcheries that strive to create fish for harvest and to recover depressed populations

As many Pacific Coast salmon populations continue to decline, the use of hatcheries has been relied on as one of several options to recover such populations; however, there is controversy about the role of hatcheries in the recovery and supplementation of wild salmon stocks (Brannon et al. 2004). Recent literature suggests that hatchery programs have had negative impacts on wild fish due to genetic, domestication, physiological, behavioral, disease, and population level effects. Although the potential genetic and ecological risks of hatcheries are of concern, it is often difficult to demonstrate conclusively that adverse effects are actually occurring and, if so, how serious they are (DFG and NMFS 2001). Recent efforts to reform hatchery management, minimizing impacts to native salmonid populations, are ongoing. Development of hatchery management plans is a key component in hatchery reform.

4.8 Climate Change

The world is about 1.3°F (0.7°C) warmer today than a century ago. The latest computer models predict that, without drastic cutbacks in emissions of carbon dioxide and other gases released by the burning of fossil fuels, the average global surface temperature may rise by two or more degrees in the 21st century (IPCC 2001). Much of that increase will likely occur in the oceans, and evidence suggests that the most dramatic changes in ocean temperature are now occurring in the Pacific (Noakes 1998). The northwestern U.S. has warmed by between 1.3°F to 1.6°F (0.7°C and 0.9°C) during the 20th century (Battin et al. 2007).

Sea levels are expected to rise by 0.5 to 1.0 meters along the northeastern Pacific coasts in the next century, mainly due to warmer ocean temperatures, which lead to thermal expansion much the same way that hot air expands. This will cause increased sedimentation, erosion, coastal flooding, and permanent inundation of low-lying natural ecosystems (*e.g.*, salt marsh, riverine, mud flats) affecting salmonid primary constituent elements. Increased winter precipitation, decreased snowpack, permafrost degradation, and glacier retreat due to warmer temperatures will cause landslides in unstable mountainous regions, and destroy fish and wildlife habitat, including salmon-spawning streams. Glacier reduction could affect the flow and temperature of rivers and streams that depend on glacier water, with negative impacts on fish populations and the habitat that supports them.

Summer droughts along the south coast and in the interior of the northwest Pacific coastlines will mean decreased stream flow in those areas, decreasing salmonid survival and reducing water supplies in the dry summer season when irrigation and domestic water use are greatest. Global warming may also change the chemical composition of the water that fish inhabit: the amount of oxygen in the water may decline, while pollution, acidity, and salinity levels may increase. This will allow more invasive species to overtake native fish species and impact predator-prey relationships (Peterson and Kitchell 2001).

It is expected that Sierra snowpacks will decrease with global warming, and that the majority of runoff in California will shift from rainfall in the winter instead of melting snowpack in the mountains. This will alter river runoff patterns and transform the tributaries that feed the Central Valley from a spring/summer-snowmelt-dominated system to a winter-rain-dominated system. In addition, the cold snowmelt that furnishes the late spring and early summer runoff will be replaced by warmer precipitation runoff. This may truncate the period of time that suitable cold water conditions persist below existing reservoirs and dams because of the warmer inflow temperatures to the reservoir from rain runoff. Without the necessary cold water pool developed from melting snowpack filling reservoirs in spring and early summer, late summer and fall temperatures below reservoirs could potentially rise above thermal tolerances for juvenile and adult salmonids.

New efforts on salmonid habitat restoration will need to accommodate the imminent impact of climate change. Recent simulation studies indicate that climate change is bound to have a large negative impact on freshwater salmonid habitat. For instance, Battin et al. (2007) predict the combined effect of climate change and habitat restoration will be a change in salmonid population abundance with a spatial shift towards lower elevations preferred by “ocean-type” salmon runs such as fall-run Chinook salmon. An adaptive management approach will provide the flexibility to track significant changes in the life history of restored Chinook salmon challenged by the most human-induced rapid environmental change in the San Joaquin River watershed.

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5.0 Limiting Factors Analyses

Limiting factors are the physical, biological, or chemical conditions and associated ecological processes and interactions that influence the abundance and productivity of San Joaquin River adult salmon. An analysis of limiting factors is based on the relationships between fish production (either adults or juveniles), parental stock (spawner) abundance, and key environmental conditions over time. These analyses primarily identify the critical life history stage that affects the production of adults and determine the primary environmental factor(s) that affects the critical life history stage. Data are often lacking for many of the environmental factors that might affect the critical life history stage(s); thus, the analysis cannot identify all the limiting factors. Therefore, it is assumed that any stressor that affects the critical life history stage(s) should be considered a potential limiting factor (Appendices A and B).

An analysis of environmental factors that limit the production of Chinook salmon in the Restoration Area cannot directly assess existing conditions in the Restoration Area because of the current absence of Chinook salmon. Furthermore, spring-run Chinook salmon escapement data collected in the Restoration Area during the 1940s would not be useful for assessing the existing habitat due to the extensive degradation that has occurred in the Restoration Area and the Delta since the late 1940s. An alternative presented here used population trend analyses conducted for fall-run Chinook salmon populations in the lower Stanislaus, Tuolumne, and Merced rivers (hereafter referred to as the “Tributaries”). Habitat characteristics that affect Chinook salmon populations in the Tributaries are similar to those that will affect spring-run and fall-run populations in the San Joaquin River. For example, the Restoration Area and all of the Tributaries have been degraded by extensive gravel mining, loss of functional floodplain habitats, and contaminated agricultural runoff. As the SJRRP moves forward, data collected on Chinook salmon populations in the Restoration Area will be used to refine limiting factors analyses and conceptual models.

A few substantial differences between spring-run and fall-run Chinook salmon, and between the San Joaquin River and the Tributaries, cannot be assessed with limiting factors analyses based on the fall-run Chinook salmon populations in the Tributaries. Instead, each of these other factors was treated as if it would limit the production of Chinook salmon in the Restoration Area. Six limiting factors have been identified as potentially affecting salmon populations in the San Joaquin River but not the Tributaries:

- Flows, water temperature, and water quality in the Delta and lower San Joaquin River may affect migrating adult spring-run salmon from April through June. In contrast, adult fall-run salmon migrate from September through November.
- An absence of adult holding habitat with suitable water temperatures throughout the summer may affect the spring-run population. In contrast, fall-run Chinook

- salmon spawn during the fall within a few weeks after arriving at their spawning beds.
- High water temperatures in the spawning reach in September and October may prevent spring-run adults from spawning, or result in high rates of egg mortality. In contrast, some fall-run spawn in late November and December, when water temperatures are usually suitable regardless of flow.
 - The Tributaries have little functional floodplain habitat in their downstream reaches whereas spring-run fry successfully rear in downstream floodplains in some Central Valley rivers, such as the Sutter Bypass on Butte Creek. It is possible that Chinook fry will rapidly grow to a smolt size in restored floodplain and wetland habitats in the downstream reaches of the Restoration Area and, if true, many would migrate into the Delta before water temperatures become unsuitable in the lower Restoration Area reaches during May.
 - The relatively long and highly degraded migration corridor for San Joaquin River smolts will result in greater exposure to stressors, such as a greater number of predators and warmer spring water temperatures, than occurs for Stanislaus and Tuolumne river fish.
 - Spring-run redds may be more susceptible to redd superimposition than fall-run redds because the spring-run Chinook salmon spawn earlier than fall-run Chinook salmon.

5.1 Limiting Factors Analysis for the Tributaries

The limiting factors analyses described below for all three Tributaries (Mesick and Marston 2007b) indicate that fry and outmigrating smolts are the most critical life history stages, and that the magnitude and duration of winter and spring flows are the primary factors that control the production of smolt outmigrants and adult fall-run salmon. The analyses also suggest that the critical period for juvenile rearing and outmigration occurs from March through about mid-June. There is evidence to suggest that high flows produce large numbers of adult fish by the following:

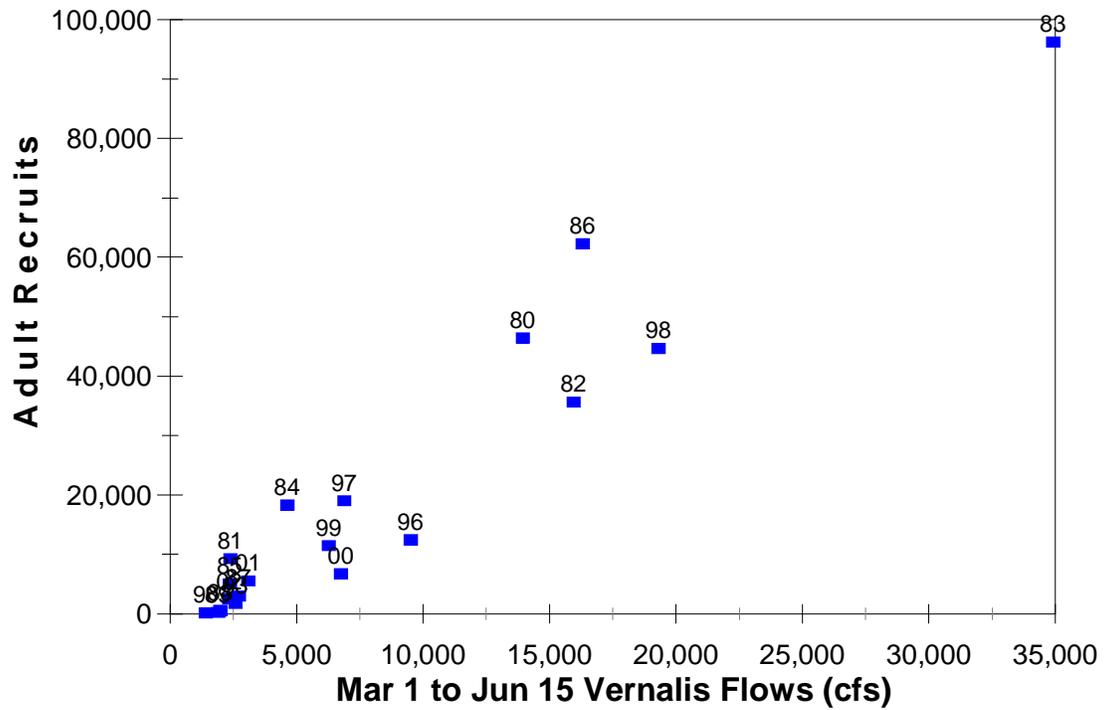
- Increasing the number of fry that survive to a smolt-size in the Tributaries from March through May
- Increasing the survival of smolts migrating through the Delta from April through mid-June (Section 4.2)
- Increasing the survival of smolts entering the ocean from May through July (Section 4.3)

These conclusions are based on correlation analyses between fish abundance and survival, mean flow magnitude, and spawner abundance. Two estimates of fish abundance and survival were used:

- Adult recruitment, which is the combined number of fish in the escapement and ocean harvest that are segregated into cohorts of same-aged fish
- Preliminary estimates of juvenile salmon abundance based on calibrated rotary screw trap catches (estimates are preliminary, because the trap efficiency models have not been finalized)
- CWT smolt survival studies (e.g., VAMP)

The analyses with adult recruitment for the lower Tuolumne River (Mesick and Marston 2007b) provide two key results. First, 93 percent of the variation ($R^2 = 0.93$) in the estimates of adult recruitment from 1980 to 2004 can be explained by the average flow rate in the San Joaquin River near Vernalis between March 1 and June 15 when the fish reared in the tributary and then migrated to the ocean as subyearling juveniles (Figure 5-1). The correlations with flows in the Tuolumne River, as measured at La Grange Dam were similar ($R^2 = 0.84$), which suggests that flow rates in the tributary and Delta are both important. Conversely, water quality indices (minimum DO and maximum water temperatures between May 1 and June 15) in the Stockton deepwater ship channel only explained between 10 percent and 24 percent of the variation in adult recruitment. Other variables, such as spawner abundance greater than 500 3-year-old equivalent adults, the mean Delta export rate during the April and May smolt outmigration period, and ocean productivity indices (PDO and upwelling) explained no more than 3 percent of the variation in adult recruitment.

San Joaquin River Restoration Program



Source: Mesick and Marston 2007b.

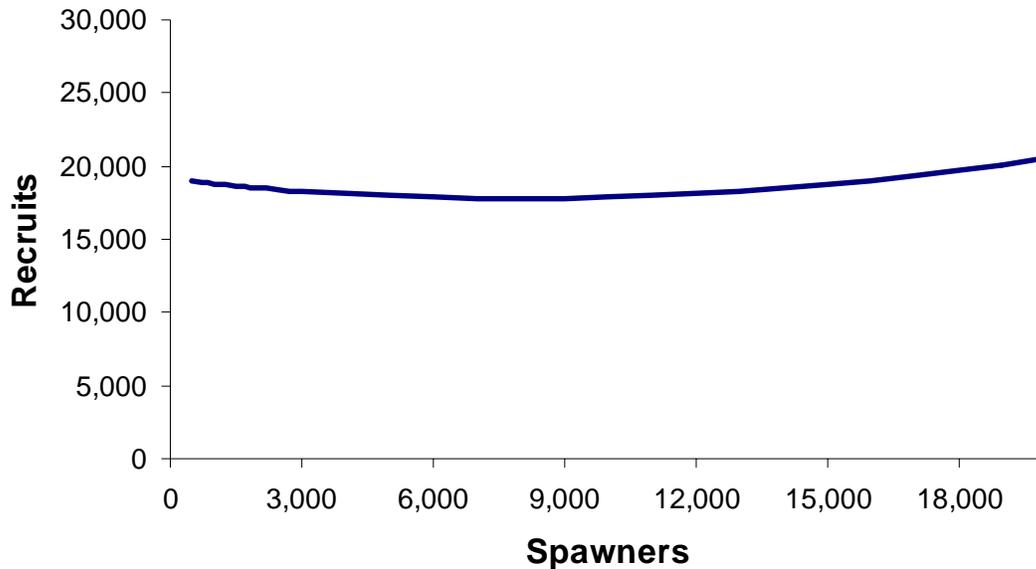
Note:

This analysis excludes recruitment estimates that were affected by a low number of spawners (< 500 Age 3 equivalent fish) to better illustrate the relationship with flow. The recruitment estimates are labeled according to the year when the fish outmigrated as smolts.

Flows are recorded from from March 1 to June 15 when the fish reared in the tributary and migrated to the ocean as subyearling juveniles

Figure 5-1.
Number of Fall-Run Chinook Salmon Recruits to Tuolumne River from 1980 to 2006 Relative to Mean Flow in the San Joaquin River at Vernalis

The second key result is that approximately 500 spawners can saturate the rearing habitat in the Tuolumne River based on the relationship between spawner abundance and recruitment shown in Figure 5-2. This suggests that neither spawner abundance (i.e., ocean harvest of adult fish) nor spawning habitat quality substantially affects adult recruitment, except during periods of prolonged drought when juvenile production drops to very low levels (e.g., 1987 through 1992).



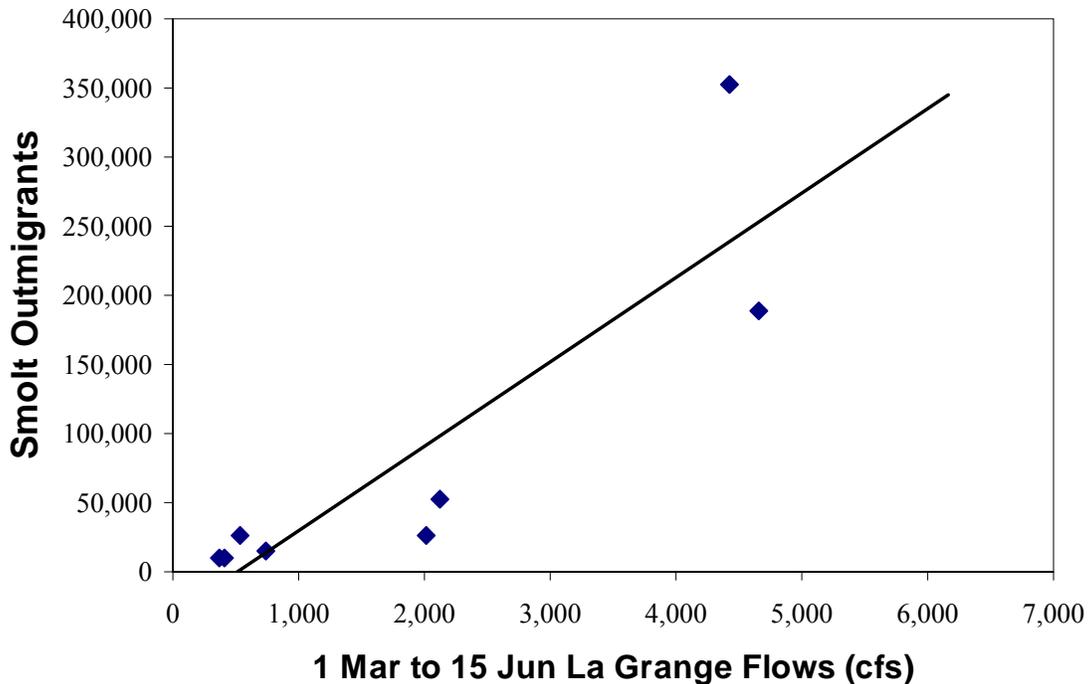
Source: Mesick and Marston 2007b.

Note:

A categorical variable called "Population Shift" was incorporated into the model to account for a shift in recruitment that occurred sometime between 1987 and 1994.

Figure 5-2.
Spawner-Recruit Relationship for Tuolumne River Based on Regression Model of Recruits, Quadratic Spawner Terms, and Mean Vernalis Flow of 7,000 cfs from March 1 Through June 15.

Rotary screw trap estimates of juvenile abundance indicate that winter and spring flows strongly affect juvenile survival in the Tributaries (Mesick and Marston 2007b). Mean spring flow releases in the Tuolumne River at La Grange from March 1 to June 15 are highly correlated ($\text{adj-R}^2 = 0.82$, $P = 0.0005$) with the number of Tuolumne River smolt outmigrants passing the Grayson traps at RM 5 (Figure 5-3). Furthermore, the production of smolt outmigrants in the Tuolumne River may be a key determinant of the number of adult recruits produced ($\text{adj-R}^2 = 0.96$, $P = 0.0004$) (Figure 5-4).

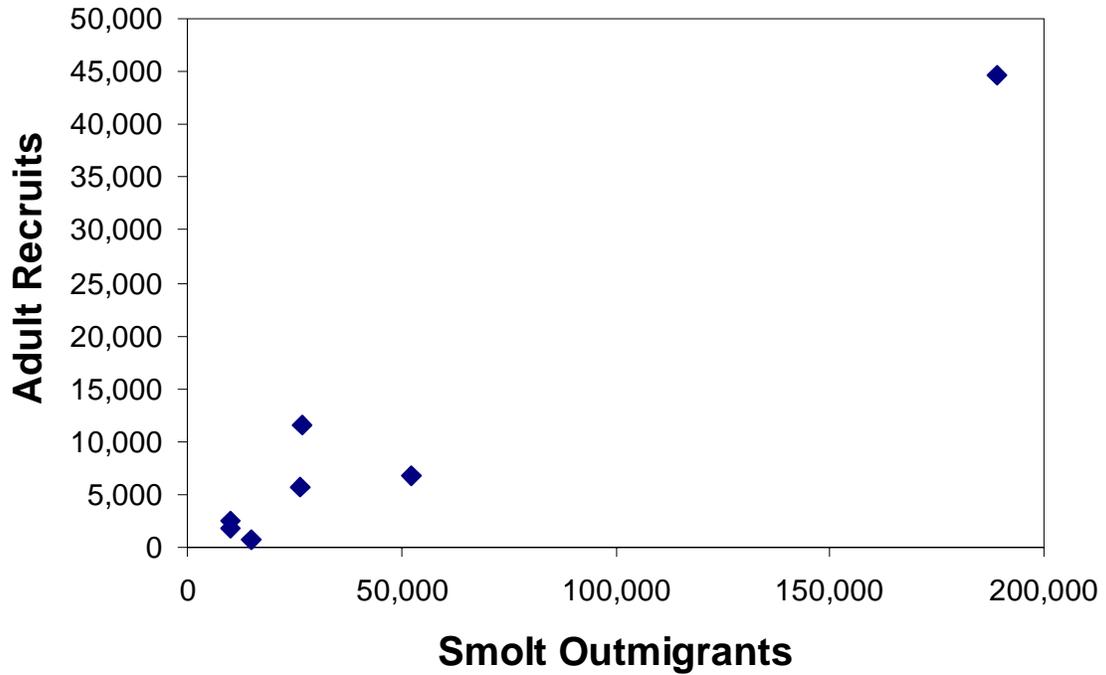


Source: Mesick and Marston 2007b.

Notes:

1. This analysis excludes smolt outmigrant estimates that were affected by a low number of spawners (< 700 Age 3 equivalent fish) to better illustrate the relationship with flow.
2. The regression model has an adj-R^2 of 0.82 and a probability level of 0.0005.
3. The screw trap estimates are preliminary because the trap efficiency models have not been finalized (DFG unpublished data).
4. Outmigrant fork length > 70 mm.

Figure 5-3.
Number of Smolt-Sized Chinook Salmon Outmigrants Passing the Grayson Rotary Screw Trap Site (RM 5) Plotted with Flows at La Grange Between March 1 and June 15 in Tuolumne River from 1998 to 2005



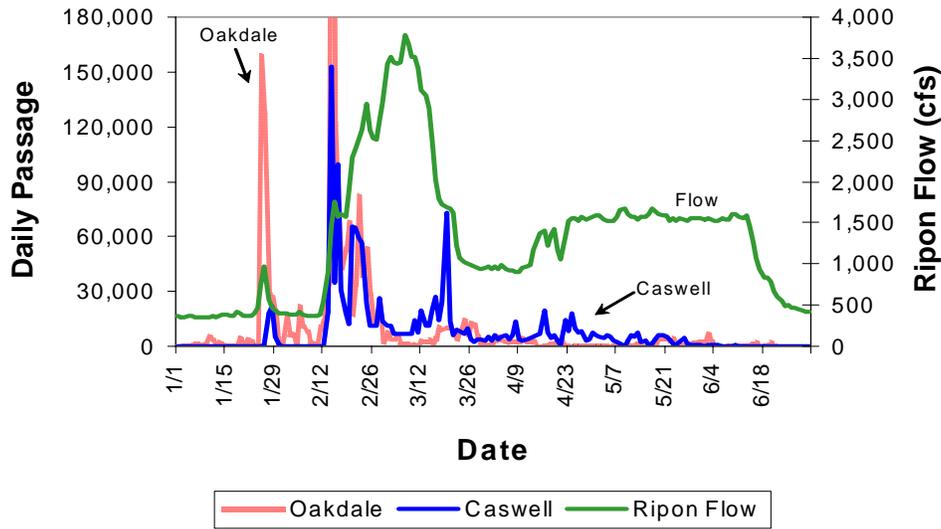
Source: Mesick and Marston 2007b.

Notes:

1. The screw trap estimates are preliminary because the trap efficiency models have not been finalized (DFG unpublished data).
2. Outmigrant fork length > 70 mm.

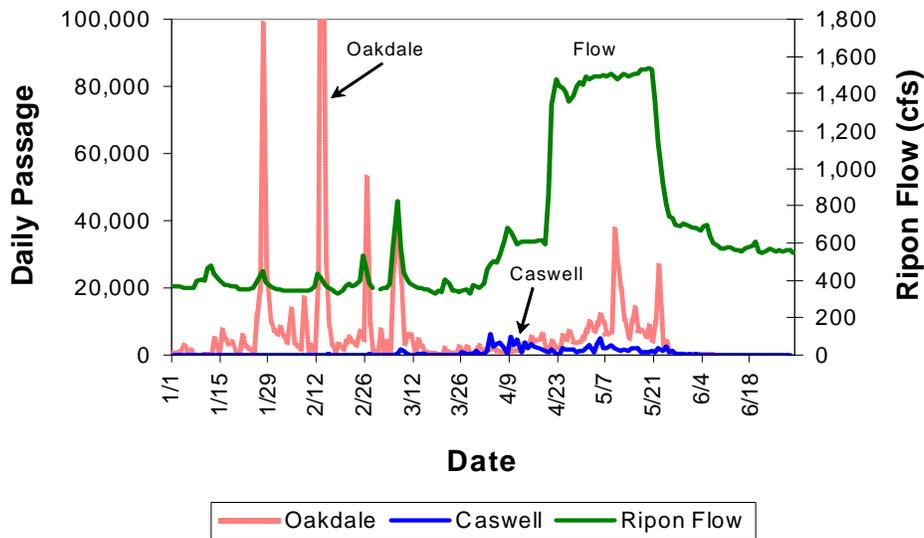
Figure 5-4.
Relationship Between Number of Smolt-Sized Chinook Salmon Outmigrants that Passed the Grayson Rotary Screw Trap Site (RM 5) and Adult Recruitment in Tuolumne River from 1998 to 2004

The rotary screw trap estimates of juvenile abundance also provide evidence that winter flows during February and March affect the production of smolt-sized juveniles in the Tributaries. Survival indices are computed by dividing the estimated number of fish caught at the terminal end of the river by the estimated number that were captured at the downstream end of the spawning grounds. In the Stanislaus River during 1998, 1999, and 2000 when flows were high between February and June, the number of juveniles that survived to a smolt size (fork length ≥ 70 mm) as they migrated through the lower river averaged 84 percent (range 74 percent to 95 percent). In addition, there were more smolt-sized fish near the terminal end of the river (Caswell State Park at RM 5) than near the lower end of the spawning grounds (Oakdale at RM 40) in April and early May, which suggests that juveniles were successfully rearing in the lower river. The spring 2000 data are shown in Figure 5-5. However, during 2001 to 2003, when flows were pulsed primarily between mid-April and mid-May, juvenile survival averaged 10 percent (range 7 percent to 11 percent), and there was no evidence of successful rearing in the lower river. The spring 2001 data are shown in Figure 5-6. These data suggest that flow during February and March might be an important determinant of the number of smolts that migrated from the Stanislaus River.



- Notes: 1. Overall juvenile survival between the Oakdale and Caswell traps was 74 percent in 2000.
 2. Fry- and parr-sized fish (<70 mm fork length) primarily migrate prior to April 1 whereas smolt-sized fish (\geq 70 mm fork length) primarily migrate after April 1.

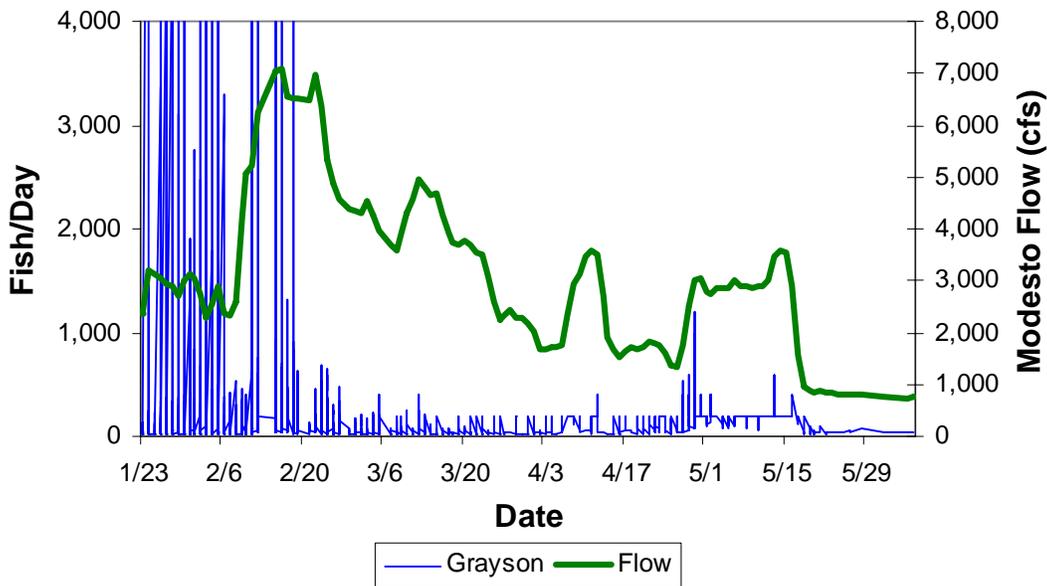
Figure 5-5.
Estimated Daily Passage of Chinook Salmon Fry and Smolt-Sized Outmigrants at Oakdale and Caswell Park Screw Traps Plotted with Mean Daily Flow at Ripon in Stanislaus River in 2000



- Note: 1. Overall juvenile survival between the Oakdale and Caswell traps was 11 percent in 2001.
 2. Fry- and parr-sized fish (<70 mm fork length) primarily migrate prior to April 1 whereas smolt-sized fish (\geq 70 mm fork length) primarily migrate after April 1.

Figure 5-6.
Estimated Daily Passage of Chinook Salmon Fry and Smolt-Sized Outmigrants at Oakdale and Caswell Park Screw Traps Plotted with Mean Daily Flow at Ripon in the Stanislaus River in 2001

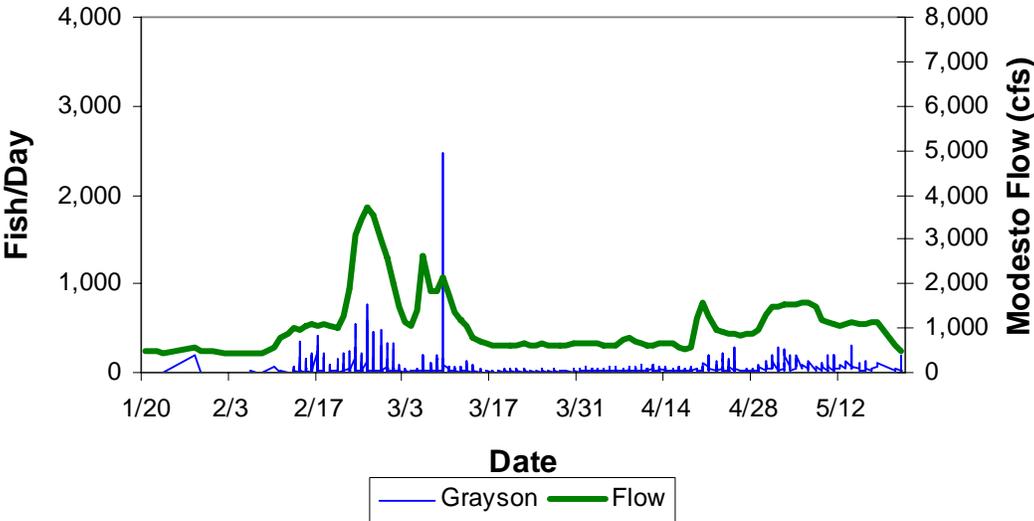
Rotary screw trap studies on the Tuolumne River from 1995 to 2005 were primarily focused on determining the number of smolt outmigrants as surveyed at the Shiloh (RM 3.4) and Grayson sites (RM 5.2). Therefore, it is not possible to compare the effect of February and March flows on the relative passage between an upstream and downstream screw trap as was done with the Stanislaus River data. However, screw trap surveys were conducted between January and late May at Grayson between 1998 and 2006 that show the same pattern observed at the Caswell State Park trap site in the Stanislaus River. Fry, parr, and smolt passage was high during wet years, such as 2000, when there were extended periods of high flows in February and March (Figure 5-7); moderate during dry years such as 2001 when there moderate periods of high flows in February and March (Figure 5-8); and low during dry years, such as 2002 when only base flows were released between February and early April (Figure 5-9).



- Note: 1. The total number of all sizes of juvenile outmigrants and smolt-sized (fork length ≥ 70 mm) outmigrants was 455,079 and 62,168, respectively.
3. Fry- and parr-sized fish (<70 mm fork length) primarily migrate prior to April 1 whereas smolt-sized fish (≥ 70 mm fork length) primarily migrate after April 1.
4. Number of fish is truncated at 4,000 fish per day

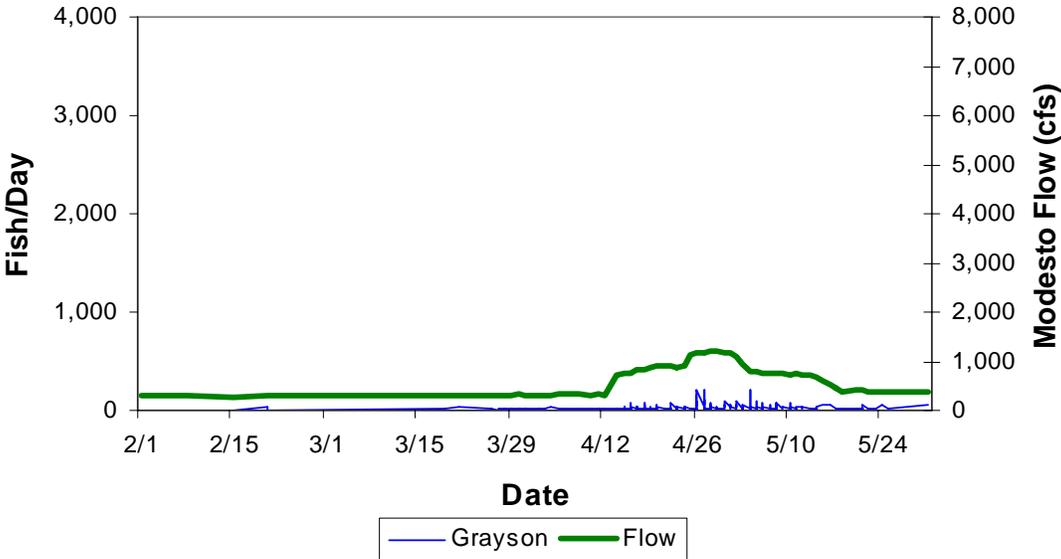
Figure 5-7.

Estimated Daily Passage of Chinook Salmon Fry and Smolt-Sized Outmigrants at Grayson Screw Trap Plotted with Mean Daily Flow at Modesto in Tuolumne River in 1999



- Note: 1. The total number of all sizes of juvenile outmigrants and smolt-sized (fork length > 70 mm) outmigrants was 111,254 and 34,824, respectively.
 2. Fry- and parr-sized fish (<70 mm fork length) primarily migrate prior to April 1 whereas smolt-sized fish (≥ 70 mm fork length) primarily migrate after April 1.

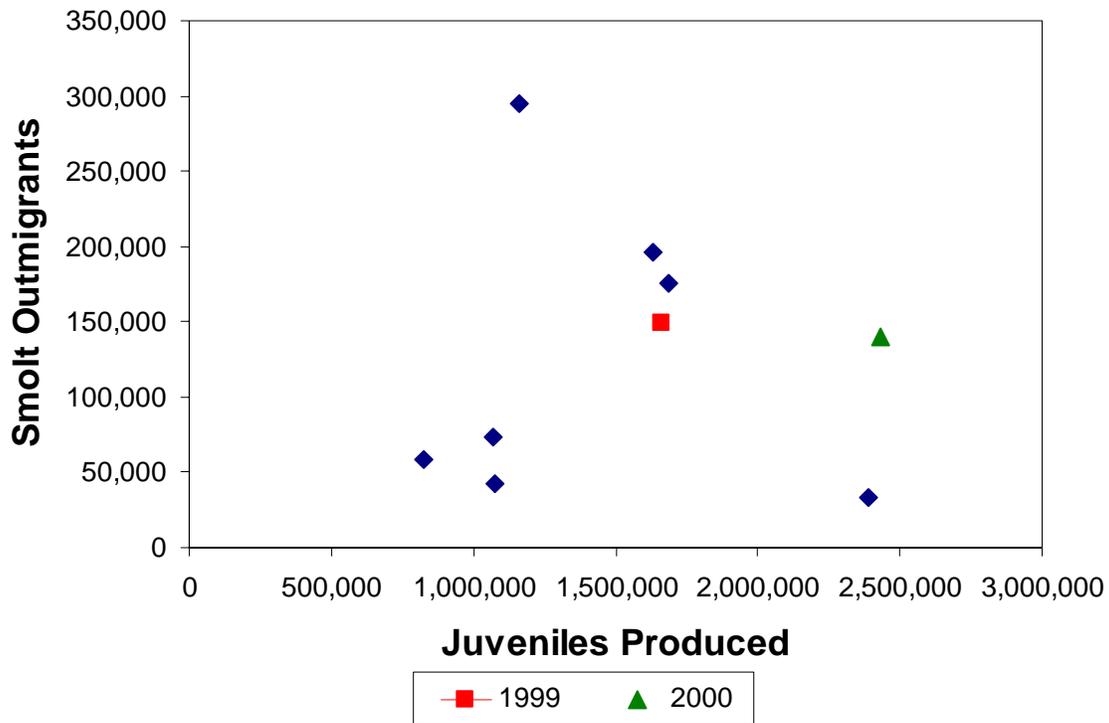
Figure 5-8.
Estimated Daily Passage of Chinook Salmon Fry and Smolt-Sized Outmigrants at Grayson Screw Trap Plotted with Mean Daily Flow at Modesto in Tuolumne River in 2001



- Note: 1. The total number of all sizes of juvenile outmigrants and smolt-sized (fork length ≥ 70 mm) outmigrants was 13,442 and 13,076, respectively.
 2. Fry- and parr-sized fish (<70 mm fork length) primarily migrate prior to April 1 whereas smolt-sized fish (≥ 70 mm fork length) primarily migrate after April 1.

Figure 5-9.
Estimated Daily Passage of Chinook Salmon Fry and Smolt-Sized Outmigrants at Grayson Screw Trap Plotted with Mean Daily Flow at Modesto in Tuolumne River in 2002

Spawning habitat was determined not to be a substantial limiting factor for the populations in the Tuolumne and Stanislaus rivers because despite the highly degraded spawning beds, relatively few spawners can produce many more fry than the existing rearing habitat in the Tributaries and Delta can support. In the Tuolumne River during 1999 and 2000 (the only years for which estimates are available), only 0.4 percent and 1.4 percent, respectively, of the estimated number of juveniles passing the upper rotary screw trap (RM 38.6) survived to a smolt-sized fish that passed the downstream trap (RM 5.2). After restoring 18 highly used spawning beds in the Stanislaus River in summer 1999 (Carl Mesick Consultants 2002b), juvenile production increased by 32 percent in spring 2000 compared to spring 1999; however, the increased juvenile production did not result in a corresponding increase in the number of smolts migrating from the river at Caswell Park (Figure 5-10) (Mesick and Marston 2007b).



Note: The Knights Ferry Gravel Replenishment Project (KFGRP) constructed 18 spawning beds in the Stanislaus River in summer 1999. In 1999 and 2000, the abundance of uvenile and smolts were affected by similar spring flows between March 1 and June 15 (1,515 cfs and 1,541 cfs, respectively from Goodwin Dam) and similar numbers of spawners (2,438 and 3,088 Age 3 equivalent fish, respectively).
Smolt outmigrant fork length \geq 70 mm

Figure 5-10.
Juvenile Production and Number of Smolt Outmigrants in Stanislaus River Before (1999) and After (2000) Spawning Gravel Enhancement Based on Abundance of Smolt Outmigrants at Caswell State Park Screw Traps (RM 5) Relative to Estimated Abundance of All Sizes of Juveniles that Passed the Oakdale Screw Trap (RM 40) from 1998 to 2004

5.2 Hypotheses

The limiting factors analyses provide evidence supporting three hypotheses regarding the production of adult spring-run and fall-run Chinook salmon in the San Joaquin River:

- The most critical life history stages are the rearing juveniles and outmigrating smolts
- The critical life history stages will be strongly affected by conditions in the Restoration Area, as well as conditions in the Delta, San Francisco Bay, and the Pacific Ocean
- The most important environmental factor that affects the critical life history stages is stream flow during the late winter and spring

The first hypothesis is that the most critical life history stages are the rearing juveniles and outmigrating smolts. Population analyses for fall-run Chinook salmon in the Stanislaus and Tuolumne rivers indicate that when relatively few adults spawn in highly degraded and sparse spawning habitats, many more fry are produced than can be supported by the rearing habitat. These analyses also indicate that the quality of the juvenile rearing and migratory habitats controls the production of adult salmon in these rivers, and it is likely that the same will be true for both spring-run and fall-run Chinook salmon in the Restoration Area. It is possible that the habitats for the other life history stages, such as adult migration, adult holding, and spawning could be critical for spring-run and fall-run salmon in the Restoration Area, if these habitats are completely absent or highly unsuitable.

The second hypothesis is that the critical life history stages will be strongly affected by conditions in the Restoration Area, as well as conditions in the Delta, San Francisco Bay, and the Pacific Ocean:

- In-river conditions have up to a 40-fold effect on smolt production in the Stanislaus and Tuolumne rivers (range from about 10,000 to 400,000 smolt-sized fish passing the rotary screw traps at the confluences between 1996 and 2005 (Figure 5-3)).
- Delta conditions have a 25-fold effect on smolt survival (range from about 2 percent to 50 percent between 1996 and 2005 (SJRG 2007)).
- Long-term cycles in marine ecosystem productivity have had up to a 2.4-fold effect on the production of adult Central Valley Chinook salmon since 1950 (Mesick and Marston 2007b).
- Ocean harvest of adult Central Valley salmon has about a 3-fold effect on escapement (harvest rates range from 26 percent to 79 percent from 1990 to 2005 (Mesick and Martson 2007b)).

Based on the above, it should be possible to achieve the Restoration Goal by improving fish habitats in the Restoration Area, although downstream and ocean conditions will influence the number of adult fish produced.

The third hypothesis is that the most important environmental factor that affects the critical life history stages is stream flow during the late winter and spring. Since the 1940s, fall-run Chinook salmon production in the Tributaries has been highest during wet years, characterized by high flows from February through June when juvenile salmon rear and migrate. Given that spring-run and fall-run salmon migrate at about the same time of year (Section 3.2.1), it is reasonable to assume that the beneficial effects of high flow would also apply to spring-run reintroduced into the Project Area. More generally, there is evidence to indicate that high flows over a prolonged period affect the survival of juvenile salmon in the rivers, Delta, and in the ocean:

- High flows between March and early June are highly correlated with the number of smolt outmigrants captured with rotary screw traps at the confluences of the Stanislaus and Tuolumne rivers (Section 5.1).
- High flows in April, May, and June are highly correlated with the survival of smolts migrating through the lower river (Section 5.1) and Delta (VAMP studies (SJRG 2007)).
- High estuary outflow in May, June, and July is correlated with the survival of smolts entering the ocean (Section 4.3).

There is uncertainty about the precise period of flows that will be necessary for the survival of juvenile spring-run Chinook salmon in the Restoration Area considering the likelihood that some of the juveniles would quickly grow to a subyearling smolt size (60-100 mm FL) in flooded wetland habitats in the downstream reaches and begin their outmigration toward the ocean in late-March through early May (Ward and McReynolds 2001, Ward et al. 2002), rather than April through mid-June when fall-run juveniles typically migrate in the Tributaries. Early migration by smolt-sized juveniles may also reduce the magnitude of the flows needed in the Delta and San Francisco Bay for their survival due to lower water temperatures and lower predation rates.

The precise mechanisms underlying the positive relationship between salmon production and flow are still under study. It is likely, however, that high flows increase smolt production and survival by improving or ameliorating a combination of potential secondary limiting factors, which include (1) food resources, (2) predation, (3) disease, (4) water temperatures, (5) contaminants, (6) water quality, (7) harvest, and/or (8) entrainment.

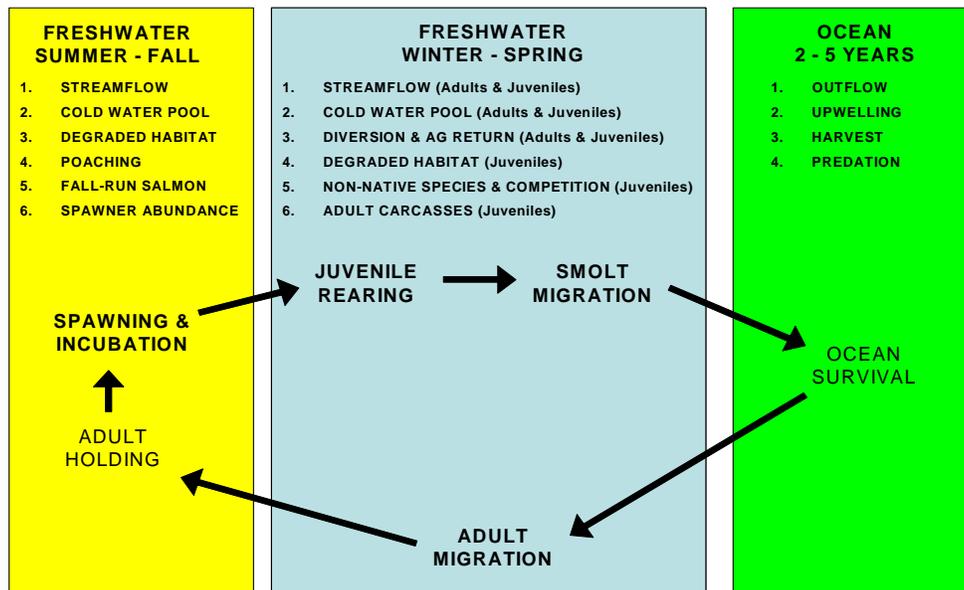
6.0 Conceptual Models

The following conceptual models represent the FMWG understanding of how the previously identified limiting factors may affect each life history stage of spring-run and fall-run Chinook salmon in the San Joaquin River basin. The FMWG recognizes that it is possible that not all limiting factors have been identified, and that the identified limiting factors may not be fully understood. Recognizing these uncertainties, the conceptual models will be developed into a series of testable hypotheses and appropriate studies described in the SJRRP Adaptive Management Plan to help evaluate the effectiveness of all restoration and management actions implemented to achieve the Restoration Goal.

Appendices A and B of this document summarize the potential limiting factors for each life-history stage of spring-run and fall-run Chinook salmon, respectively, and compare restoration targets with existing conditions in the San Joaquin River basin. Also presented is an assessment of how serious the impact of a given factor is likely to be, and how this likelihood may change with water year type. The conceptual models assume that all actions prescribed in the Settlement, such as screening the bypass channels and improving passage conditions, will be implemented. The adaptive management strategy will include monitoring to determine the effectiveness all actions, including those described in the Settlement.

6.1 Spring-Run Chinook Salmon

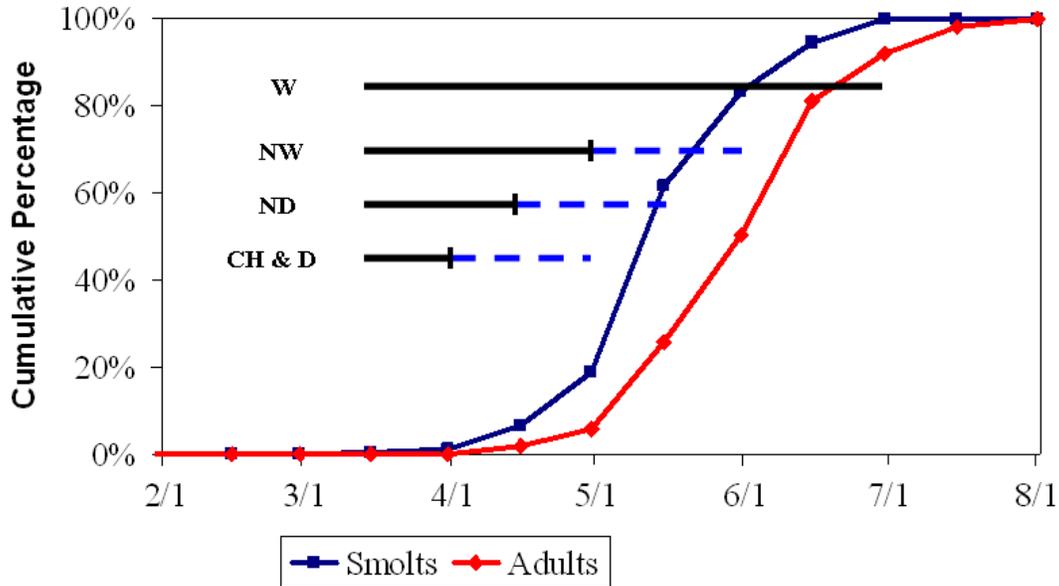
The abundance of adult spring-run Chinook salmon that return to spawn in the Restoration Area will probably be affected by numerous factors, only some of which will be under the control of the SJRRP whereas other factors will be outside the control of the SJRRP (Figure 6-1). Potential limiting factors that the SJRRP will have some control over include the following:



Note: The life stages in bold type are assumed to be the most critical for achieving the Restoration Goal.

Figure 6-1.
Overall Conceptual Model for San Joaquin River Spring-Run Chinook Salmon

- Restoration Flows** – Truncated spring pulse flows specified in the Settlement’s Restoration Hydrographs may protect no more than 83 percent of the migrating smolt-sized juveniles (≥ 70 mm FL) and no more than 50 percent of the migrating adults during all but wet years, assuming that the Restoration Hydrographs can be shifted up to 4 weeks, and that the reintroduced San Joaquin fish behave similarly to those that rear in the upper reaches of Butte Creek in the Sacramento Basin (Figure 6-2). A primary concern is whether it will be possible to shift the Restoration Hydrographs into May to protect a greater number of migrating adults and juvenile salmon, while providing at least periodic floodplain inundation during the March through May rearing period, maintaining suitable water temperatures for juvenile and adult salmon (target $\leq 68^{\circ}\text{F}$ (20°C)), without exhausting the cold water pool in Millerton Lake. Extending the high flow period into May and June would probably increase smolt production and survival by improving or ameliorating a combination of factors, which include food availability, predation, disease, water temperatures, contaminants, water quality, harvest, and entrainment. However, it is also possible that many fry will migrate to the downstream reaches of the Restoration Area where they will rapidly grow to a smolt size in restored floodplain and wetland habitats prior to May. If true, pulse flows between February and April may produce a sufficient number of smolts to sustain the salmon populations.



Sources: Hill and Webber 1999, Ward and McReynolds 2001, Ward et al. 2002, CDFG 1998.

Notes:

1. The solid black horizontal lines represent the release period for spring pulse flows as prescribed in the Settlement during Critical High (CH), Dry (D), Normal Dry (ND), Normal Wet (NW) and Wet (W) years. No spring pulse flows would be released during Critical Low years.
2. The dashed blue horizontal lines represent the maximum flexibility to shift the flow schedule as prescribed by the Settlement.

Figure 6-2.
Relationship Between Timing of Settlement Spring Pulse Flows and Mean Cumulative Percentage of Fish Passage for Butte Creek Subyearling Spring-Run Smolts and Historical Populations of Adult Spring-Run Chinook Salmon in the Sacramento Basin

- **Cold Water Pool** – The volume of the cold water pool in Millerton Lake (Section 4.1.2) may be insufficient to provide the prescribed summer and fall flow releases and maintain suitable water temperatures for holding adult spring-run salmon during the summer (target < 70°F) and incubating salmon eggs during the fall (target < 58°F).
- **Habitat Degradation Within the Restoration Area** – The highly degraded channel and floodplain morphology, loss of native riparian vegetation, and exotic species below Friant Dam to the confluence with the Merced River may result in high rates of mortality for juvenile salmon. In addition, the lack of gravel recruitment from lateral and upstream sources and high flow events (e.g., 1997) have gradually scoured away the spawning gravels immediately downstream from Friant Dam.

Spawner Abundance – Legal and illegal harvest of yearling juveniles and spawning adults may substantially limit adult recruitment, particularly if escapements are low. In addition, conditions that result in low production of juvenile salmon will limit the number of adult fish that return to spawn 2 to 4 years later.

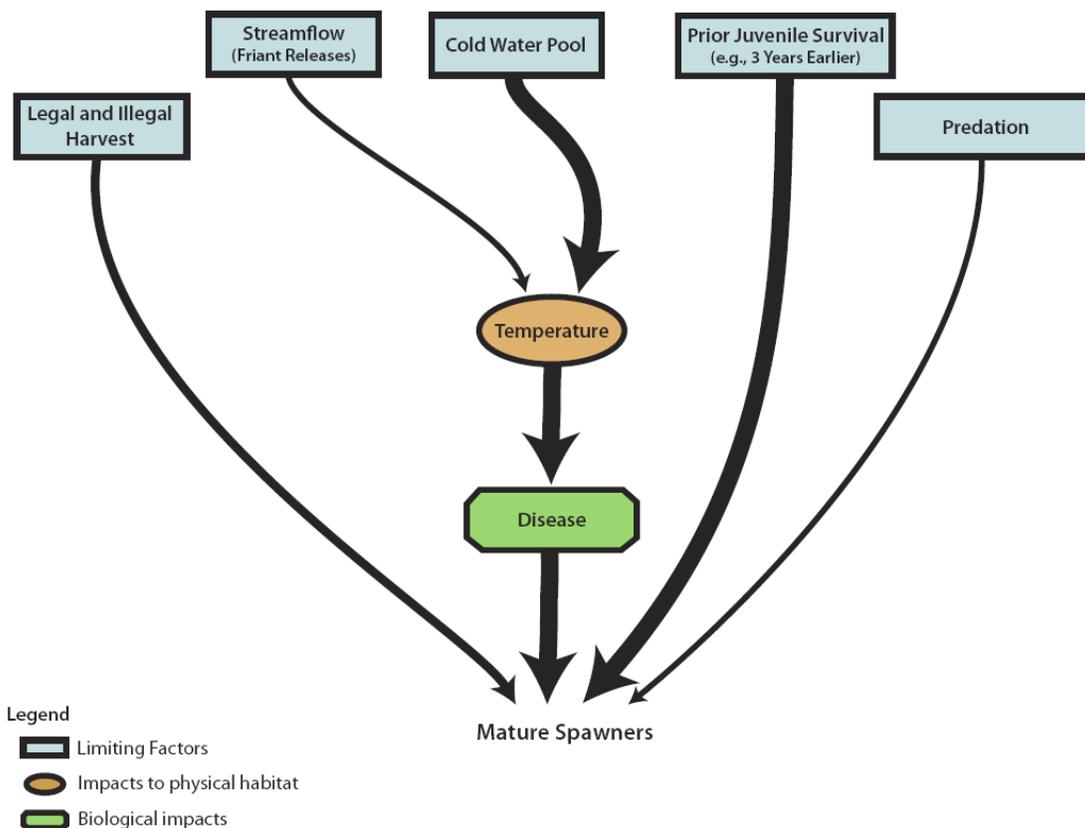
Factors outside the control of the SJRRP that have been identified include the following:

- **Streamflow Releases Outside the Restoration Area** – Stream flow releases in the Stanislaus, Tuolumne, and Merced rivers that contribute to flows in the mainstem San Joaquin River, Delta, and San Francisco Estuary are expected to affect the survival of rearing and migrating juvenile salmon and the survival and homing ability of adult salmon.
- **Habitat Degradation Outside the Restoration Area** – The highly degraded channel and floodplain morphology, loss of native riparian vegetation, and exotic species below the confluence with the Merced River, the Delta, and San Francisco Estuary are expected to substantially reduce the survival of rearing and migrating juvenile salmon.
- **Water Quality** – Pesticides and other contaminants may substantially reduce the food resources needed by juvenile salmon within and below the Restoration Area, and to a lesser degree, result in direct mortality of juvenile salmon. In addition, poor water quality (e.g., low DO and high ammonia concentrations) in the mainstem channel may affect the survival of juvenile, and to a lesser degree, adult salmon.
- **Delta Exports** – Springtime Delta exports at the CVP and SWP pumping facilities affect entrainment of juvenile fish as well as reducing flow in the Stockton deepwater ship channel and the amount of freshwater outflow into the ocean, all of which affect the survival of juvenile salmon and the ability of adult salmon to home to the Restoration Area.
- **Ocean Productivity** – Ocean productivity (food resources), as affected by upwelling, coastal currents, El Niño events, and the amount of freshwater outflow from the San Francisco Bay, will affect the survival of juvenile and adult salmon;
- **Climate Changes** – Climate changes are expected to affect inland water temperatures, hydrographs (i.e., floodplain inundation), and ocean productivity conditions, and therefore, affect the survival of juvenile and adult salmon; and
- **Harvest and Predation in the Ocean** – Harvest of adult salmon and predation of juvenile and adult salmon in the ocean affect the number of adults that return to spawn, which may affect the number of juveniles produced during the following spring.

The following are potential mechanisms by which the above limiting factors are expected to affect each life history stage of spring-run Chinook salmon, including adult holding, spawning, juvenile rearing, smolt migration, ocean survival, ocean harvest, and adult migration. Potential benefits and impacts of hatcheries and climate change are also discussed in terms of overall population effects.

6.1.1 Adult Holding

There are currently several holding pools below Friant Dam that were extensively used by spring-run salmon during the 1940s. It is likely that these pools can sustain at least 20,000 fish. However, there are concerns that high water temperatures, and to a lesser degree, predation and harvest (legal and illegal) may affect the ability of spring-run salmon to hold in these pools (Figure 6-3). The number of spawners is also substantially affected by the survival of the fish when they were juveniles, 2 to 5 years earlier.



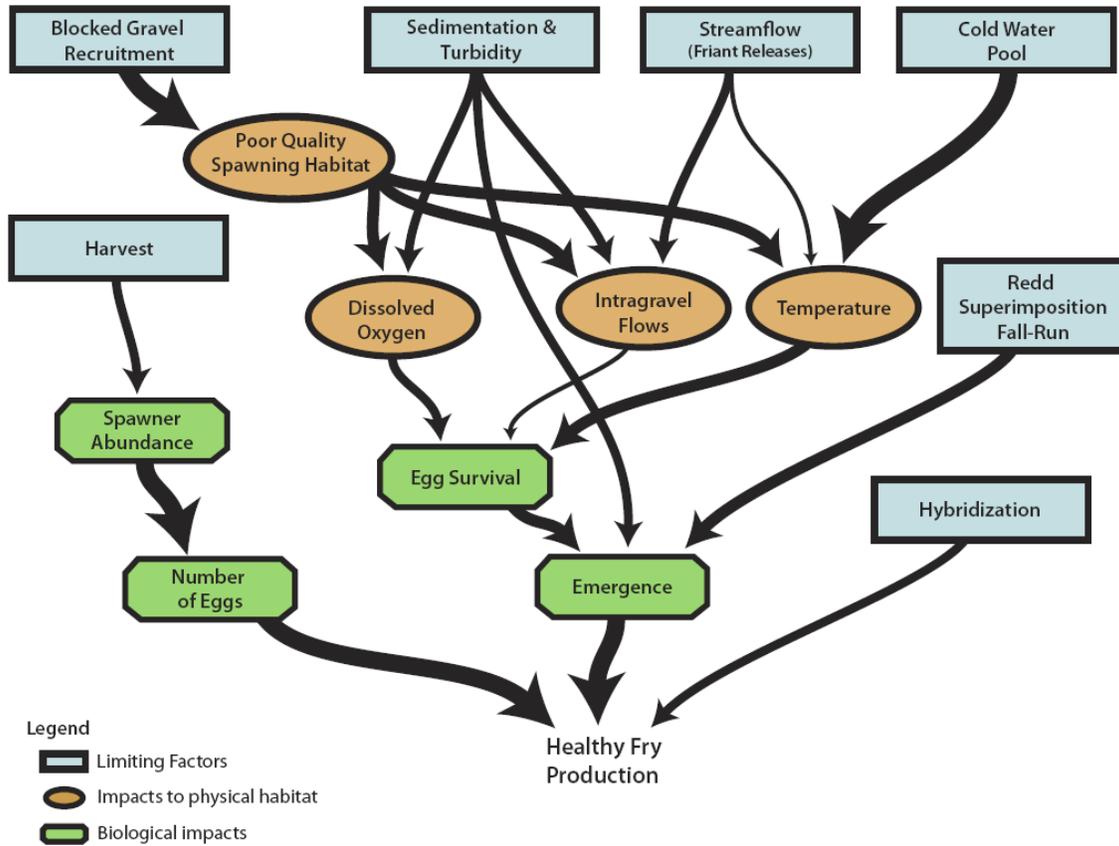
Note: The width of the arrows indicates the relative importance of each mechanism.

Figure 6-3.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Holding Adult Spring-Run Salmon

- **High Water Temperatures** (Section 4.5.4) – If the cold water pool in Millerton Lake is exhausted as a result of increased summer and fall flows, the temperature of the release flows could exceed suitable levels for holding adults. If temperatures become unsuitably high, disease may become a likely cause of mortality (Section 4.5.5).
- **Harvest** (Section 4.5.3) – Adults will be susceptible to legal and illegal harvest while they hold in the pools below the dam. If escapements are too low to saturate the rearing habitat with juvenile fish, the harvest of adult spawners from the holding pools could become a substantial limiting factor.
- **Predation** (Section 4.5.6) – Mammals have the potential to consume large numbers of spawners, but generally scavenge post-spawned fish. It is assumed that predation of holding adults will not have a population level effect. Therefore, predation will not be directly evaluated unless routine monitoring indicates that adult mortality rates during the holding period are higher than expected.

6.1.2 Spawning and Egg Incubation

Spring-run Chinook salmon will probably spawn in the reach immediately downstream from Friant Dam, where water temperatures should be suitable for spring-run spawning and egg incubation between August and January. However, there are only a few, highly silted beds in this reach because Friant Dam has blocked most of the gravel recruitment, and high flows since 1950 have scoured the gravels from these beds (Section 4.6.1). It is likely that the adults would be forced to spawn in either the highly degraded habitats immediately below the dam or in the downstream habitats where egg survival and alevin emergence could be highly impaired by high water temperatures. Another substantial concern is that the increased summer and fall flows required by the Settlement may exhaust the cold water pool in Millerton Lake such that water temperatures of the release flows become unsuitable for adult spawners and egg incubation (Sections 4.1.2 and 4.6.2) (Figure 6-4). Other concerns include sedimentation of spawning gravels, turbid storm runoff during egg incubation, redd superimposition by fall-run salmon, hybridization with fall-run salmon, and legal and illegal harvest of adults (Figure 6-4).



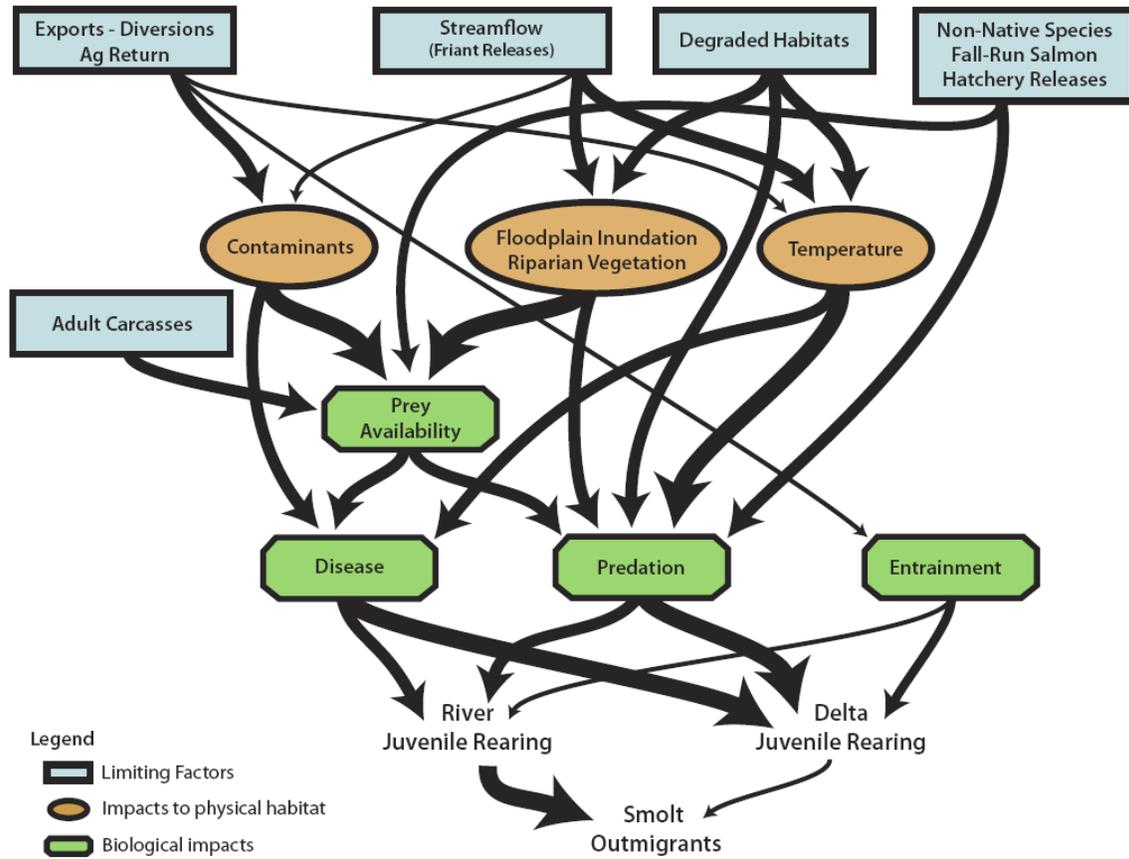
Note: The width of the arrows indicates the relative importance of each mechanism.

Figure 6-4.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Spawning and Incubation Habitat for Spring-Run Salmon.

- **Redd Superimposition by Fall-Run Salmon** (Section 4.1.3) – Fall-run salmon will probably spawn at the same locations where spring-run fish spawn; thus, there is potential that spring-run redds would be superimposed by fall-run spawners, thereby killing spring-run eggs, especially when fall-run escapements are high.
- **Hybridization with Fall-Run Salmon** (Section 4.6.3) – A small percentage of fall-run salmon will probably spawn at the same time and location as spring-run fish, so there is potential for hybridization. Hybridization may result in fish with migratory behaviors that might not be viable in the San Joaquin River basin. For example, hybridization between fall-run and spring-run salmon in the Feather River Hatchery has resulted in adult fish that primarily migrate during the summer (current passage rates at RBDD as shown in Figure 3-4).
- **Sedimentation and Turbid Storm Runoff** (Section 4.1.1) – High permeability measurements made in Reach 1A in 2002 (McBain and Trush 2002) suggest that sedimentation has not adversely affected spawning habitat quality at those locations. However, turbid storm runoff may cause egg mortality, particularly if ground-disturbing activities (e.g., construction or bank erosion) occur near Friant Dam or in one of the upper tributaries (e.g., Cottonwood Creek). It is possible that coating eggs with clay-sized particles suffocates the embryos, or at least stunts their growth.
- **Harvest** (Section 4.6.5) – Adults will be susceptible to legal and illegal harvest particularly while they spawn on shallow gravel beds. If escapements are too low to saturate the rearing habitat with juvenile fish, the harvest of adult spawners from the spawning beds could be a substantial limiting factor.

6.1.3 Juvenile Rearing

Juvenile salmon that rear in the upper SJRRP reaches and begin their downstream migration in May and June are expected to be substantially impacted by the truncated spring Restoration Hydrographs prescribed in the Settlement, the highly degraded physical habitats within and downstream from the Restoration Area, and exotic species that potentially compete for food or prey on juvenile salmon (Figure 6-5). The primary mechanisms by which these factors will affect the production of salmon smolts are probably linked to reduced food resources, temperature-increased metabolic demands, and abnormally high rates of predation and disease. In the upstream reaches, it is likely that the combined effects of limited food resources and low water temperatures will result in slow growth rates for juvenile salmon that delay the onset of smoltification until late spring (May and June) when downstream conditions in the Delta are usually unsuitable for migrating smolts. In the downstream reaches, the lack of inundated floodplain and wetland habitats from late January through early May may limit their survival.



Note: The width of the arrows indicates the relative importance of each mechanism.

Figure 6-5.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Production of Juvenile Spring-Run Salmon in the San Joaquin River.

The following summarizes the key mechanisms by which the limiting factors may affect the survival of rearing juvenile spring-run salmon.

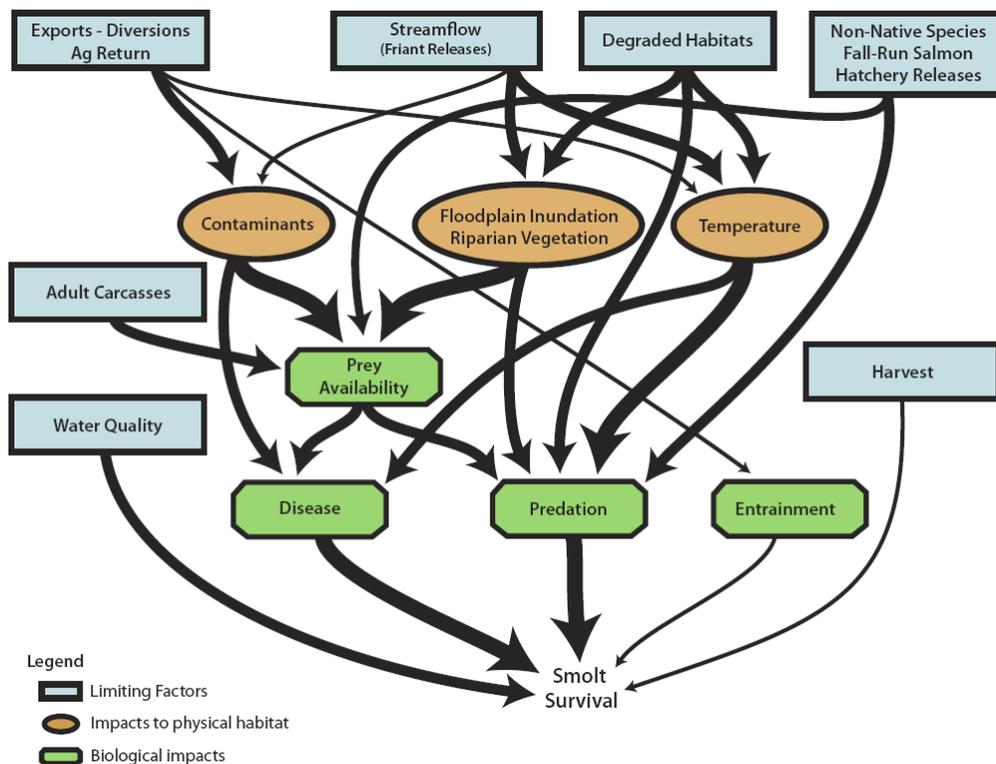
- **Inadequate Food Resources** can result from many potential causes:
 - Reduced magnitude and duration of winter and spring flows (presumably February through mid-June) reduces floodplain inundation that provides food organisms and organic detritus supporting the food web for juvenile salmon (Section 4.2.1).
 - Pesticides and other contaminants may reduce the abundance of food organisms (Section 4.2.1).
 - Elevated water temperatures may increase food requirements beyond the amount available to juvenile salmon (Section 4.2.7).
 - Levees, dikes, and dredger tailings reduce floodplain habitat inundation that provides food organisms and organic detritus supporting the food web for juvenile salmon (Section 4.2.6).
 - Low numbers of adult salmon carcasses will reduce food resources for juveniles (Section 4.2.1). This will be a particular problem for the first few years before adults begin to return.
 - Loss of riparian vegetation on floodplain and wetland habitats reduces the input of organic detritus that drives the juvenile salmon's food web (Section 4.2.1).
 - Nonnative invasive species include plants that may not augment the salmon's food supply (Section 4.2.1). Invertebrate species, such as Asiatic freshwater clams, and fish, such as centrarchids, may compete with salmon for food.
 - Competition with other native fish species, including fall-run juveniles may reduce food resources for spring-run juveniles (Section 4.2.1).
 - Potential competition with hatchery-reared juveniles released into the Restoration Area may reduce food resources for juveniles (Section 4.2.1). Initially, releases of hatchery-reared Chinook salmon in the Restoration Area will be limited to survival studies and releases to calibrate rotary screw traps. It is assumed that impacts of limited releases for study purposes will be minimal.
 - Intermittent flows in bypass channels used as rearing areas may reduce food resources (Section 4.2.1). Typically when floodplains or bypass channels become inundated, there is an initial pulse in terrestrial food resources followed by a gradual increase in aquatic food resources.

- Sedimentation and gravel extraction affects the composition of the invertebrate community, although it is unknown whether the change in species composition substantially affects the availability of food for juvenile salmon (Section 4.2.1).
- **Predation** – Predation by native and introduced fish species can be abnormally high when flows are confined to the low-flow channel and water temperatures are high (Section 4.2.3).
 - Key predators are thought to include Sacramento pikeminnow, which feeds all year, striped bass, which typically begins migrating into tributary habitats in April, and introduced centrarchids, when they begin feeding in April or May as water temperatures rise. These fish tend to utilize dredged habitats in the Restoration Area and Delta, including captured mine pits, the Stockton Deepwater Ship Channel, and canals leading to the CVP and SWP pumping facilities. Nonnative submerged aquatic vegetation provides habitat for nonnative predators.
 - Another concern is that hatchery-reared catchable-sized trout may be released into the Restoration Area during the winter and spring rearing period and that these fish would prey on juvenile salmon. However, it is DFG policy not to release hatchery reared trout into rivers with anadromous fish populations.
- **Disease** – Disease may be a substantial source of mortality when food resources are low, water quality is poor, and/or water temperatures are high (Section 4.2.2).
- **Entrainment** – The bifurcation structures in the Restoration Area will be screened as directed by the Settlement; however, it is uncertain whether the screens will be fully effective (Section 4.2.5). Large unscreened diversions, such as those of the West Stanislaus Irrigation District, Patterson Water Company, and El Solyo Water Company, may entrain a substantial number of fry and parr. There is no information on entrainment rates at the numerous small diversions throughout the basin.
- **Degraded Physical Habitat** (Section 4.2.6) – Loss of connected floodplain habitats, in-river gravel extraction, blocked sediment recruitment by upstream dams, bank stabilization, and reduced recruitment of LWD reduce the suitability of the habitats used by parr-sized juveniles (50-80 mm FL) for feeding stations and predator refuge.
- **Contaminants** (Section 4.2.6) – It is assumed that contaminants do not directly cause juvenile mortality, but rather have indirect effects by reducing food resources or accelerating disease infestation rates.
- **High Water Temperatures** – Water temperatures that exceed 77°F (>25°C) (Section 4.2.7) in late spring may cause juvenile mortality. However, it is assumed that juvenile salmon die from other factors, such as predation, disease, or starvation, as water temperatures approach lethal levels.

6.1.4 Smolt Migration

The likely causes of mortality for migrating sub-yearling smolts are expected to be similar to those for rearing juveniles, including the truncated spring hydrographs prescribed in the Settlement, the highly degraded physical habitats within and downstream from the Restoration Area, and exotic species that potentially compete for food or prey on juvenile salmon (Figure 6-6). However, it is likely that the negative impacts of high water temperatures, contaminants, water quality (e.g., ammonia near wastewater treatment plants, DO concentrations in the deepwater ship channel), entrainment, and predation will be much worse for juveniles that slowly grow to a smolt size in the upper reaches and then outmigrate between April and mid-June compared to those that rapidly grow in warmer downstream reaches and then outmigrate between late March and early May. Another problem that may affect smolts is sport harvest.

The relative importance of these stressors may partially depend on whether the smolts migrate through the natural channels or bypass channels. It is expected that predation will be a greater problem in the natural channel compared to the bypass channels, which would only receive intermittent flows during the migratory period. In contrast, the bypass channels may have higher water temperatures that would improve the growth of fry between January and April, but negatively impact smolts migrating in May and June.

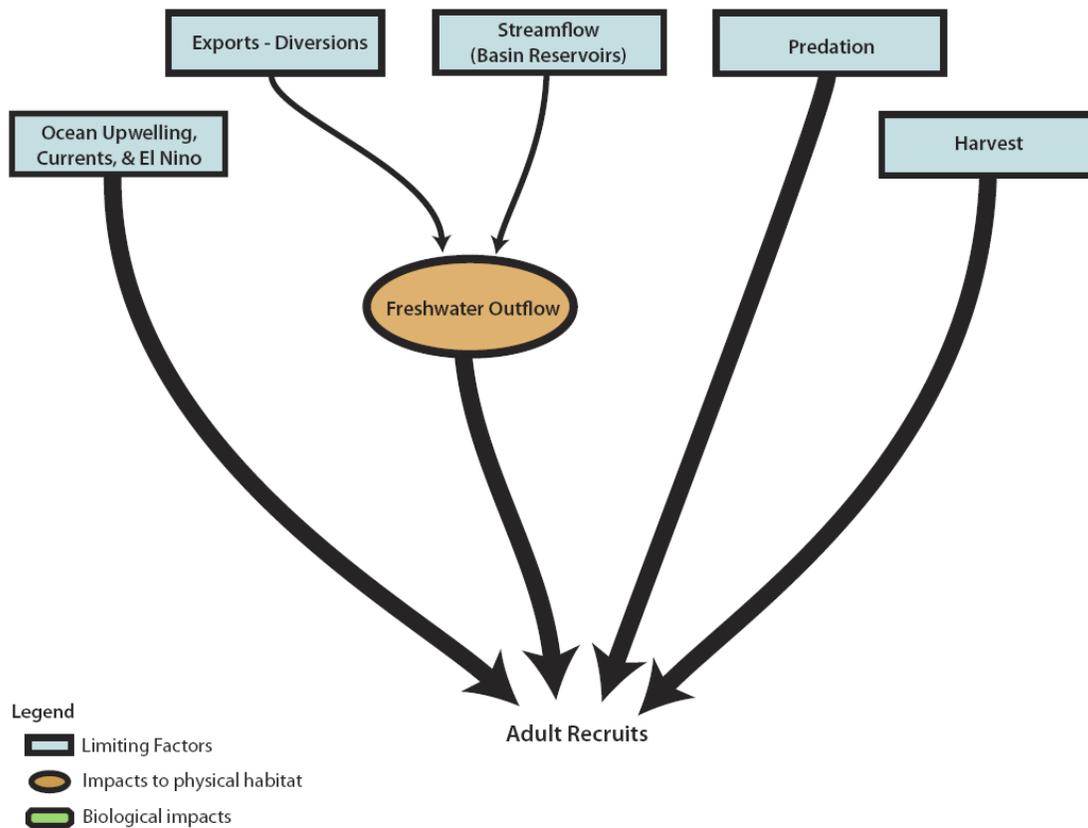


Note: The width of the arrows indicates the relative importance of each mechanism.

Figure 6-6.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Survival of Migrating San Joaquin River Spring-Run Salmon Smolts

6.1.5 Ocean Survival

The survival of smolts entering the ocean during June and July is probably the most critical phase for salmon in the ocean (Figure 6-7). Freshwater outflow from the estuary is highly correlated with smolt survival and the availability of food resources at the interface between freshwater and saltwater (Section 4.3). Coastal upwelling, ocean currents, and El Niño events also affect ocean productivity and the survival of smolts entering the ocean (Section 4.3). Indices of ocean productivity conditions will be incorporated into the assessment of adult salmon production in the Restoration Area.



Note: The width of the arrows indicates the relative importance of each mechanism.

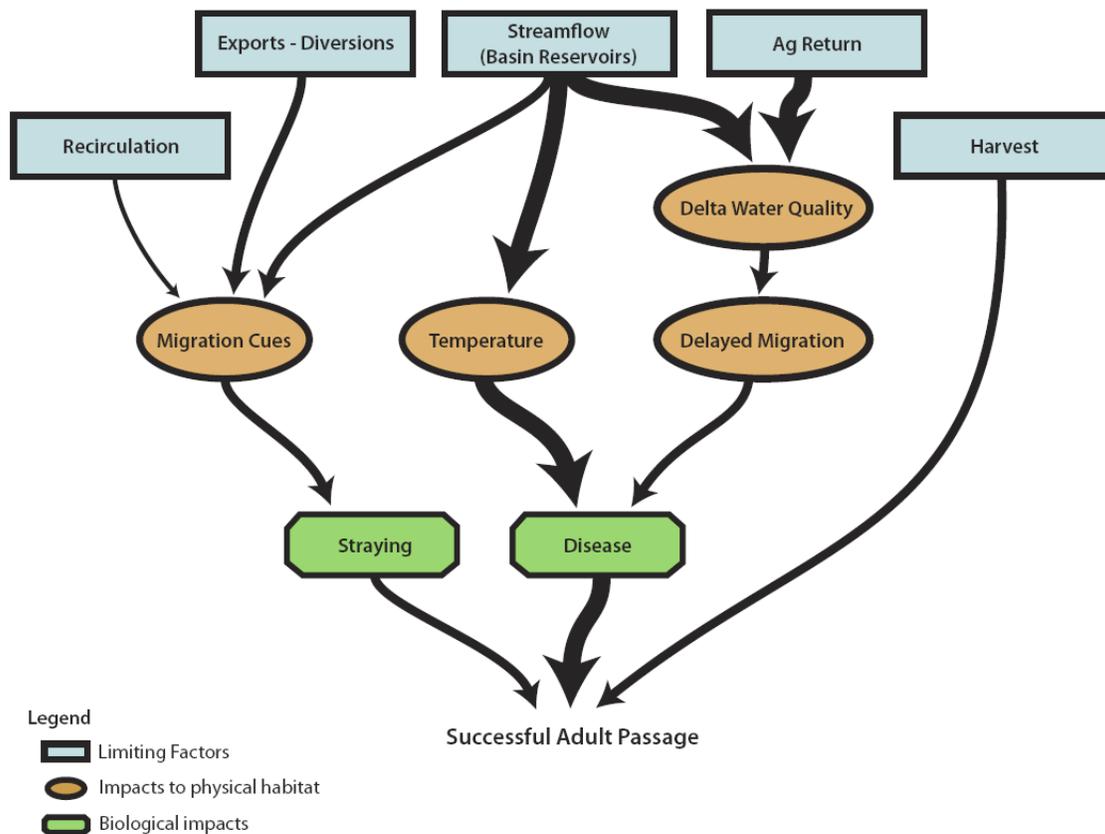
Figure 6-7.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Survival of San Joaquin River Spring-Run Salmon in the Ocean

6.1.6 Ocean Harvest

It is anticipated that ocean harvest rates (Section 4.3.3) will have population level effects whenever harvest rates reduce escapement to the point that there are too few spawners to saturate the habitat with juveniles (Figure 6-7). Estimates of ocean harvest rates will be incorporated into the assessment of adult salmon production in the Restoration Area.

6.1.7 Adult Migration

Conditions in the lower Restoration Area (Reaches 3 through 5) and the Delta may affect adults in terms of passage and straying rates (Sections 3.3 and 4.4). The most significant concern is that when the spring-pulse flows cease, water temperatures will become unsuitable and the adults will succumb to disease or other sources of mortality (Figure 6-8). It is also important to remember that the conditions that affect juvenile survival in freshwater (Sections 6.1.3 and 6.1.4) and ocean habitats (Section 6.1.5) also affect the number of adults that return to spawn.



Note: The width of the arrows indicates the relative importance of each mechanism.

Figure 6-8.
Possible Limiting Factors, Impacts to Physical Habitats, and Biological Impacts that May Affect Survival of Migrating Adult San Joaquin River Spring-Run Salmon.

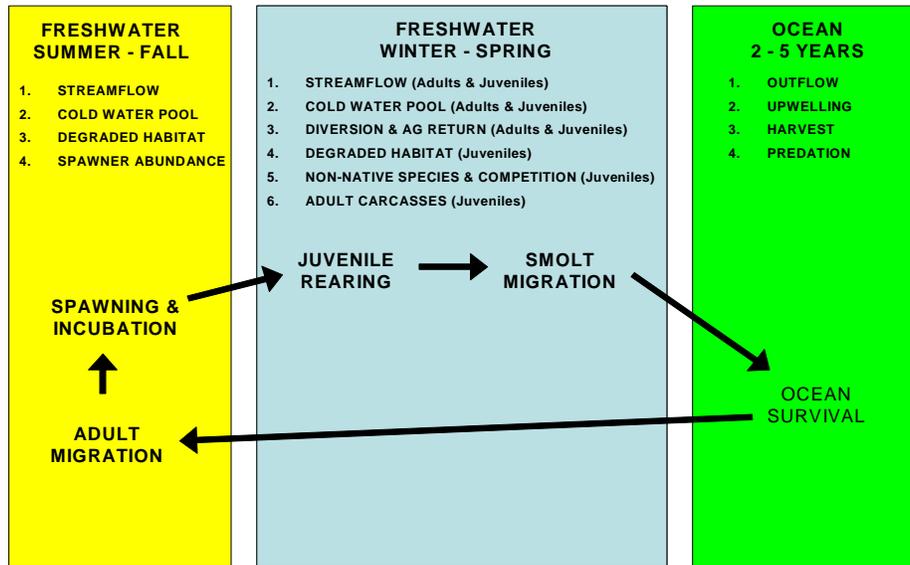
- It is likely that without spring pulse flow releases, water temperatures will become high enough (70 to 80°F) in late spring and early summer to cause high rates of adult mortality due to disease (Sections 4.4.2, 4.5.4, and 4.5.5). It is unlikely that suboptimal water temperatures would affect gamete viability because the fish migrate when they are sexually immature.
- Low DO concentrations and possibly high water temperatures may delay passage for adults in the deepwater ship channel near Stockton, particularly when the tributary pulse flows cease in mid- to late May (Section 4.4.4) and thereby worsen high temperature-related impacts.
- High export rates relative to flows (export rates ≥ 400 percent of Vernalis flows) can cause up to 20 percent of adult San Joaquin salmon to stray to the Sacramento River basin (Section 4.4.1).
- Legal and illegal harvest of adult fish in freshwater habitats (Section 4.4.5) may result in an inadequate number of spawners to saturate the rearing habitat with juveniles.

6.1.8 Hatcheries

Hatcheries can benefit or impact the natural salmon population depending on how they are operated (Section 4.7). Potential beneficial uses of hatcheries include (1) incubating eggs from a source population of spring-run Chinook salmon for the purposes of reintroduction, (2) sustaining the salmon populations (i.e., genetic diversity) during drought conditions when flows are not sufficient for juvenile survival, and (3) providing fish for rotary screw trap calibration studies and smolt survival studies that identify high priority restoration projects, passage problems, and critical flow periods. Potential negative impacts to the natural population include genetic contamination (i.e., decreased fitness), sources of disease, and competition with naturally produced juveniles.

6.2 Fall-Run Chinook Salmon

The environmental factors that are likely to affect the production of fall-run Chinook salmon are nearly identical to those that affect spring-run Chinook salmon, with a few exceptions. The primary difference is that adult fall-run do not require summer holding habitat, because they mostly migrate in October and November and then spawn shortly thereafter. The key management issues are whether the cold water pool in Millerton Lake will be sufficient to restore naturally reproducing populations of both salmon runs.



Note: The life stages in bold type are assumed to be the most critical for achieving the Restoration Goal.

Figure 6-9.
Overall Conceptual Model for San Joaquin River Fall-Run Chinook Salmon

6.2.1 Spawning

Adult fall-run Chinook salmon have nearly the same spawning habitat requirements as those described for spring-run fish and it is likely that they will utilize the same spawning beds after the spring-run have completed their spawning. It is possible that this overlap in habitat use will result in redd superimposition and hybridization impacts on the spring-run population (Sections 3.5 and 4.6).

6.2.2 Adult Migration

Adult fall-run Chinook salmon have nearly the same migration requirements as those described for spring-run fish, except that fall-run fish will mostly migrate in October and November when high flows will be needed to provide suitable water temperatures (Sections 3.3 and 4.4). The main concern is whether fall pulse flows of sufficient magnitude and duration to permit passage for migrating adult fall-run salmon would exhaust the cold water pool in Millerton Lake and thereby potentially increase the temperature of Friant releases above the levels needed to successfully incubate spring-run salmon eggs from August through December.

6.2.3 Juvenile Rearing

The limiting factors analyses suggest that juvenile survival in the Restoration Area will be an important determinant of adult production, and that there is potential for competition between juvenile spring-run salmon and juvenile fall-run salmon. Juveniles of both runs will probably utilize the same resources since their rearing periods are expected to overlap substantially (Section 3.2.1). It is possible that spring-run juveniles will have a competitive advantage over the fall-run juveniles for the limited food resources and habitats (Section 4.2.1), because they will emerge first and be slightly larger than the fall-run juveniles. However, it is also possible that large numbers of juvenile fall-run salmon could substantially reduce the survival of spring-run juveniles.

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7.0 Data Needs

The following are key information needs, and tasks required to address the needs for spring-run and fall-run Chinook salmon in the San Joaquin River and for downstream salmon populations.

7.1 Spring-Run Chinook Salmon

To effectively manage the recovery of a naturally reproducing spring-run Chinook salmon population, the following information should be considered:

- **Source Populations** – Identify potential source populations that are not genetically contaminated, migrate early, tolerate high water temperatures, and are sufficiently robust to provide fish for reintroduction to the San Joaquin River. It is desirable to obtain a source population of naturally reproducing fish to help ensure that the restored population eventually contributes to genetic diversity and, therefore, the recovery of the Central Valley spring-run Chinook salmon ESU as determined by NMFS (Lindley et al. 2007). In addition, it will be necessary to verify the anecdotal information that some populations, such as Butte Creek, migrate early and can tolerate relatively high water temperatures. It will also be necessary to determine whether the source population is sufficiently robust that transferring several hundred adults will not impact its sustainability.
- **Adult Fish Passage** – Evaluate the effects of the Restoration Flow releases, water temperatures, and Delta exports on adult fish passage.
 - Link Reclamation’s HEC-5Q River temperature model for the Restoration Area with the HEC 5Q CALFED temperature model for the lower San Joaquin River below the confluence of the Merced River to estimate water temperature at 20-mile intervals throughout the migratory corridor (Friant Dam to Dos Reis) in hourly time steps for spring-run adults (April 15 to June 15) and fall-run adults (September 1 to November 15) for each Restoration Flow Hydrograph.
 - Develop a quantitative model of the relationship of the effects of flow, water temperature, DO concentrations in the deepwater ship channel, and Delta export rates on straying rates and gamete viability of adult spring-run salmon. Use existing data to estimate straying rates and gamete viability relative to flow and water temperatures. Utilize the CALFED-sponsored water temperature model for the San Joaquin River below the confluence of the Merced River.
 - Evaluate adult passage relative to potential barriers and structural improvements to be implemented in the Restoration Area.

- Determine the impact of altered groundwater inflow on water temperatures and flow in the adult migration corridor.
- **Spawning Habitat Assessment** – Determine the distribution and quality of spawning habitat below Friant Dam:
 - Survey the location of spawning habitats.
 - Obtain and analyze sediment bulk samples from likely spawning beds located throughout the 10-mile-long reach immediately below Friant Dam.
 - Measure sedimentation rates and turbidity in the primary spawning reach during the spring-run spawning period.
- **Holding Habitat** – Evaluate the effects of the Restoration Flow releases and water temperatures on the suitability of holding habitat:
 - Use the SJRRP water temperature model to estimate the water temperature at one-mile intervals for the 5-mile-long reach immediately below Friant Dam in 6-hour time steps from April 15 to August 31 for each Restoration Flow Hydrograph.
 - Determine temperature tolerances for holding adult spring-run for each potential source population.
- **Cold Water Pool** – Evaluate the effects of the Restoration Flow releases and water diversions on the size of the cold water pool in Millerton Lake and the suitability of the release temperatures for spring-run spawning habitat:
 - Use the Program’s water temperature model to estimate the water temperature of the release flows from Friant Dam in 6-hour time steps from April 15 to December 31 for each Restoration Flow Hydrograph.
 - Evaluate the benefits of installing temperature control devices on release and diversion structures to conserve the volume of the cold water pool in Millerton Lake.
- **Spawning/Incubation** – Evaluate the effects of the Restoration Flow releases and water temperatures on spawning and egg incubation habitats. Evaluate how redd superimposition from fall run spawners may affect the production of juvenile spring-run salmon.

- Use the SJRRP water temperature model to estimate the water temperature at one-mile intervals for the 5-mile-long reach immediately below Friant Dam in six-hour time steps from September 1 to December 31 for each Restoration Flow Hydrograph.
- Evaluate the benefits of installing temperature control devices on release and diversion structures to conserve the volume of the cold water pool in Millerton Lake.
- Determine temperature tolerances for adult spring-run spawners for each potential source population.
- Develop a quantitative model of the relationship between flow, water temperature, the amount of suitable spawning habitat, redd superimposition with and without fall-run Chinook salmon, and the expected maximum number of fry that could be produced.
- **Poaching** – Estimate how poaching may impact the abundance of spring-run spawners in the San Joaquin River:
 - Assess the effects of legal and illegal harvest of Chinook salmon and other fish.
- **Juvenile Survival** – Evaluate how the Restoration Flow releases and water temperatures will affect the number of spring-run juveniles that survive to a smolt size in the San Joaquin River:
 - Use the SJRRP water temperature model to estimate the water temperature at ten-mile intervals throughout Reach 1 in 6-hour time steps from March 1 to May 31 for each Restoration Flow Hydrograph.
 - Estimate the impact of altered groundwater inflow on water temperatures and flow in rearing habitats.
 - Estimate the benefits of restoring channel width, channel depth, and widths of mature riparian tree forests or wetland habitats on water temperatures throughout the Restoration Area.
 - Survey the size, location, and potential for predation at the in-river gravel excavation sites in the Restoration Area.
 - Develop a quantitative model to compare the effects of flow, water temperature, and other potential stressors for juveniles rearing in the upper reaches with those rearing in the lower reaches. Stressors evaluated should include food resources, predation, disease, contamination, and entrainment.

- **Smolt Survival** – Evaluate how Restoration Flow releases and water temperatures will affect the survival of spring-run smolts migrating from the San Joaquin River:
 - Link Reclamation’s HEC-5Q River temperature model for the Restoration Area with the HEC 5Q CALFED temperature model for the lower San Joaquin River below the confluence of the Merced River to estimate the water temperature at 20-mile intervals throughout the migratory corridor (Friant Dam to Dos Reis) in 6-hour time steps for smolt outmigrants (March 15 to June 15) for each Restoration Flow Hydrograph.
 - Determine the impact of altered groundwater inflow on water temperatures and flow in juvenile migration corridors.
 - Estimate the benefits of restoring channel width, channel depth, and widths of mature riparian tree forests or wetland habitats on water temperatures throughout the Restoration Area.
 - Survey the size, location, and potential for predation at the in-river pits and other gravel excavation sites in the Restoration Area.
 - Develop suitability criteria for juvenile spring-run for each potential source population.
 - Develop a quantitative model of the effects of flow, water temperature, and smolt survival between Friant Dam and the confluence with the Merced River.
- **Food Availability** – Evaluate how the Restoration Flows, water temperatures, floodplain inundation, exotic species, contaminants, channel morphology, and fine sediments affect food availability for juvenile salmon:
 - Survey the location of functional and diked floodplain habitats, wetland habitats, exotic plant and fish species, agricultural lands that discharge irrigation runoff into the river, and fine sediment sources between Friant Dam and the confluence with the Merced River.
 - Update the hydraulic and digital terrain models used to evaluate relationships between flow and floodplain inundation.
 - Develop a quantitative food supply model that includes the effects of flow, nutrients, floodplain inundation, wetland habitat inundation, native and exotic riparian vegetation, instream production, channel morphology, and reservoir (Millerton Lake) production.

- **Limiting Factors Assessment** – Evaluate the relative importance of unscreened diversions, predators in captured mine pits and other degraded habitats, starvation, contamination, and disease to juvenile mortality in the San Joaquin River:
 - Survey the unscreened diversions, predators and their habitats, contaminated agricultural runoff, and riparian vegetation on functional floodplains.
 - Incorporate the results of these studies into the quantitative model.
- **Delta Survival** – Evaluate the effects of flow, water temperature, exports, the Head of the Old River Barrier, water quality and ocean-vessel traffic in the deepwater ship channel, and conditions in the Old River channel on the survival of spring-run smolts in the Delta. Evaluate the effects of ocean conditions on the survival of San Joaquin River salmon smolts:
 - Incorporate the results of the VAMP studies into the quantitative model.
 - Incorporate the results of ongoing ocean studies.
- **Quantitative Models** – Predict the abundance of adult spring-run salmon in the San Joaquin River below Friant Dam using the quantitative models developed for the above tasks.

7.2 Fall-Run Chinook Salmon

To effectively manage the recovery of a naturally reproducing fall-run Chinook salmon population, the following information should be considered:

- **Adult Fish Passage and Gamete Viability** – Evaluate the effects of the Restoration Flow releases, water temperatures, and Delta exports on adult fish passage and gamete viability:
 - Same tasks as for spring-run salmon.
 - Assess gamete viability at the Merced River hatchery relative to flow releases, Delta exports, and water temperatures in the river and Delta.
- **Spawning Habitat** – Determine the distribution and quality of spawning habitat below Friant Dam.
 - Same tasks as for spring-run salmon.

- **Cold Water Pool** – Evaluate the effects of the Restoration Flow releases and water diversions on the size of the cold water pool in Millerton Lake and the suitability of the release temperatures for spring-run spawning habitat. Determine if it is necessary to enhance spawning habitat downstream from Friant Dam where water temperatures will be suitable under the Restoration Flows. Determine if it is necessary to block fall-run spawners from spring-run spawning areas to prevent superimposition on spring-run salmon redds.
 - Same tasks as for spring-run salmon.
- **Spawning/Incubation.** – Evaluate the effects of the Restoration Flow releases and water temperatures on spawning and egg incubation habitats:
 - Same tasks as for spring-run salmon.
- **Juvenile Survival** – Evaluate how Restoration Flow releases and water temperatures will affect the number of fall-run juveniles that survive to a smolt size in the San Joaquin River:
 - Same tasks as for spring-run salmon.
- **Smolt Survival** – Evaluate how the Restoration Flow releases and water temperatures will affect the survival of fall-run smolts migrating from the San Joaquin River:
 - Same tasks as for spring-run salmon.
- **Food Availability** – Evaluate how the Restoration Flows, water temperatures, floodplain inundation, exotic species, contaminants, channel morphology, and fine sediments affect food availability for juvenile salmon:
 - Same tasks as for spring-run salmon.
- **Juvenile Mortality** – Evaluate the relative importance of unscreened diversions, predators in captured mine pits and other degraded habitats, starvation, contamination, and disease to juvenile mortality in the San Joaquin River:
 - Same tasks as for spring-run salmon.
- **Smolt Survival** – Evaluate the effects of flow, water temperature, exports, the Head of the Old River Barrier, water quality and ocean-vessel traffic in the deepwater ship channel, and conditions in the Old River channel on the survival of spring-run smolts in the Delta:
 - Same tasks as for spring-run salmon.
- **Adult Abundance.** Predict the abundance of adult fall-run salmon in the San Joaquin River below Friant Dam using the quantitative models developed for the above tasks.

7.3 Impacts on Downstream Chinook Salmon Populations

Evaluate potential effects of restoration flows on water temperatures in the lower San Joaquin River below the confluence with the Merced River and the potential negative impacts on fall-run salmon populations in the Merced, Tuolumne, and Stanislaus rivers:

- Link Reclamation's HEC-5Q River temperature model for the Restoration Area with the HEC 5Q CALFED temperature model for the lower San Joaquin River below the confluence of the Merced River to estimate the effect of the Restoration Flow releases on the water temperatures at 20-mile intervals between the Merced River confluence and Dos Reis, in hourly time steps from April 1 to June 15 and from September 1 through November 15 for each Restoration Flow Hydrograph.

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8.0 References

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Conceptual Models of Stressors and Limiting Factors for San Joaquin River Chinook Salmon

Appendices Potential Stressors, Suitable and Existing Conditions



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Appendix A Potential Stressors, Restoration Targets, Existing Conditions, Likely Impact Levels, Certainty of Impact on Spring-Run Chinook Salmon

Appendix B Potential Stressors, Restoration Targets, Existing Conditions, Likely Impact Levels, Certainty of Impact on Fall-Run Chinook Salmon

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Appendix A
Potential Stressors, Restoration Targets,
Existing Conditions, Likely Impact Levels,
Certainty of Impact on Spring-Run Chinook
Salmon

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Table A-1.

Egg Survival and Emergence Life Stage (Section 4.1)

SUITABLE CONDITIONS Spawning: Late August to October Incubation to Emergence: Late August to February		EXISTING CONDITIONS					
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range		Impact Level	Level of Certainty
				Location	Frequency		
Blocked Gravel Recruitment	Numerous beds of suitably sized gravel.	Reach 1A, particularly in the 5-mile reach below Friant Dam	Few spawning beds in 5-mile reach below Friant Dam.	Rive Miles 262.5 to 267.5		High	High
Elevated Water Temperature	<57°F.	Reach 1A, particularly in the 5-mile reach below Friant Dam	When Millerton's cold water pool becomes depleted, the temperature of releases flows would exceed 57°F. Even with suitable release flow temperatures, flows may be too low to provide suitable water temperatures below Rive Miles 262.5, where spawning beds exist.		>77% of years	High	Medium
Excessive Sedimentation	Fines < 10% in spawning beds.	Reach 1A, particularly in the 5-mile reach below Friant Dam	Most spawning beds appear to be highly silted between Rive Miles 262.5 and 267.5, but fine concentrations have not been measured with sieve analysis.	Rive Miles 262.5 to 267.5		Medium	Low
Turbidity	Low turbidity during incubation.	Reach 1A, particularly in the 5-mile reach below Friant Dam	Unknown	Unknown	Unknown	Unknown	
Redd Superimposition	No redd superimposition between spring-run and fall-run Chinook salmon.	Reach 1A, particularly in the 5-mile reach below Friant Dam	Likely high impacts because redd superimposition rates by fall-run salmon are high in the Tributaries	Highest rates likely near Friant Dam	Depends on fall-run escapement	Medium	Medium
Hybridization	No hybridization between spring-run and fall-run Chinook salmon.	Reach 1A, particularly in the 5-mile reach below Friant Dam	Possibly high because spring-run and fall-run salmon do not coexist in the Central Valley unless their populations are segregated.	Unknown	Unknown	Medium	Low

Potential stressors, suitable and existing conditions in the San Joaquin Basin, likely impact levels, certainty of the impact on spring-run Chinook salmon
Appendix A.

**Table A-2.
Juvenile Rearing and Migration (Sections 4.2 and 6.1)**

PREFERRED CONDITIONS December to June				SPRING PULSE FLOWS PRESCRIBED IN SETTLEMENT			
Stressor	Water Year Type	Timing	Magnitude	Timing Potential to Extend Period by 3 Weeks	Magnitude (cubic feet per second)	Impact Level	Level of Certainty
Truncated Springtime Pulse Flows	Wet Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 4,000 2,000 350	Low	High
	Normal-Wet Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 4,000 350 350	Medium	Medium
	Normal-Dry Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 350 350 350	High	High
	Dry Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 350 350 350 350	High	High
	Critical-High Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 200 200 215 255	High	High
	Critical-Low Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	130 150 150 190 230	High	High

**Table A-3.
Juvenile Rearing and Migration (Sections 4.2)**

PREFERRED CONDITIONS December to June		EXISTING CONDITIONS					Impact Level	Level of Certainty
Stressor	Restoration Targets	Water Year Type	Location	Observed Range	Outside Preferred Range			
					Location	Frequency		
Excessively High Water Temperature	<60 - 69°F (depending on food resources and water quality).	Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	High
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
Inadequate Food Resources	Prolonged inundation of floodplains and wetlands, productive riparian vegetation, numerous adult salmon carcasses, suitable water temperatures, low contaminants, natural channel morphology, low sedimentation, low competition with juvenile spring-run salmon and trout, and few exotic species.	Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
Predation	Natural channel morphology, natural hydrograph, and few exotic fish or planted catchable-sized trout.	Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
Disease	Low infestation rates when water temperatures, water quality, and food resources are adequate.	Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium

**Table A-3.
Juvenile Rearing and Migration (Section 4.2) (cont'd)**

PREFERRED CONDITIONS December to June		EXISTING CONDITIONS					Impact Level	Level of Certainty
Stressor	Restoration Targets	Water Year Type	Location	Observed Range	Outside Preferred Range			
					Location	Frequency		
Entrainment	No diversions during rearing and migration or intakes have effective screens.	All	San Joaquin River: Friant Dam downstream to Chipps Island	Unknown	Unknown	Unknown	Unknown	
Degraded Physical Habitat	Natural channel morphology, natural riverbed sediment sizes, and natural levels of large woody debris.	All	Friant to Estuary	A majority of the physical habitat has been degraded in the San Joaquin River basin from numerous causes, including blocked sediment recruitment by Friant Dam, dikes and levees that separate the low flow channel from floodplains, in-channel gravel excavation, sedimentation, removal of riparian vegetation, and channel dredging in the deepwater ship channel.			High	Low
Contaminants	Zero to low level contaminants, including pesticides, herbicides, and certain trace elements (e.g., selenium).	All	Friant to Estuary	Contaminant levels are high in the San Joaquin River basin, particularly during spraying to control West Nile Virus.			High	Medium
Sport Harvest	Low harvest rates of yearling-sized Chinook salmon.	All	Friant to Estuary, but primarily Reach 1A during summer and fall	It is assumed that subyearling smolts primarily contribute to adult recruitment; therefore, the harvest of yearlings will not have a substantial impact.			Low	Low
Water Quality	High concentrations of dissolved oxygen and low levels of ammonia.	All	City of Stockton to Jersey Point	Dissolved oxygen levels frequently become unsuitable in the deepwater ship channel when Vernalis flows fall below 2,000 cubic feet per second and outflow from wastewater treatment plants can increase ammonia levels.			Medium	Medium

**Table A-4.
 Ocean Survival (Section 4.3)**

PREFERRED CONDITIONS Juvenile Entry: May to July Adults: 2-5 Years		EXISTING CONDITIONS					
Stressor	Restoration Target	Location	Observed Conditions	Outside Preferred Range		Impact Level	Level of Certainty
				Location	Frequency		
Inadequate Food Resources	Upwelling of nutrient-rich water in coastal waters and large freshwater plumes from the San Francisco Estuary.	Ocean	There are 15- to 25-year cycles of productive and unproductive periods. An unproductive period began in 2002. There were unusually poor conditions in 2005 that are affecting 2007 escapements. Freshwater plumes are probably relatively small except during uncontrolled releases in wet years due to high rates of diversions throughout the Central Valley and high rates of exports at the Central Valley Project and State Water Project pumping facilities.			Medium	High
Predation	Natural rates of predation by fish, bird, and mammal.	Ocean	Predation levels are highest when juvenile abundance is high.			Unknown	
Adult Commercial and Sport Harvest	Harvest should not reduce escapement to the point that the rearing habitat cannot be saturated with juvenile salmon. The rearing habitat in the San Joaquin River Tributaries can be saturated by the progeny of about 500 adult fall-run Chinook salmon.	Ocean	The Central Valley Index of ocean harvest ranged from 26% to 62% between 2000 and 2006. The nonlanded fishing mortality south of Horse Mountain, Humboldt County, California, ranged between 9% to 15% (mean 11.8%) of the commercial catch and between 11% and 16% (mean 12.2%) of the sport catch.			Unknown	

**Table A-5.
Adult Migration (Section 4.4)**

PREFERRED CONDITIONS Spawning: Late August to October Incubation to Emergence: Late August to February		EXISTING CONDITIONS					
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range		Impact Level	Level of Certainty
				Location	Frequency		
Inadequate Flows and High Delta Exports	Sufficient flow from Friant Dam must reach the San Francisco Estuary to provide migratory cues to allow adult salmon to home to the San Joaquin River. Delta exports should not exceed 400% of flows.	Friant to Estuary	It is likely that straying rates will increase when spring pulse flows cease and Delta exports exceed about 400% of flow releases.		To be modeled	Medium	High
Excessively High Water Temperatures	< 56°F.	Friant to Delta	It is likely that water temperatures in the lower Restoration Area and Delta will exceed suitable levels when spring pulse flows cease.		To be modeled	High	High
Structural Barriers	Suitable water depths > 1 foot and screened diversion canals.	Restoration Area	The Settlement prescribes that passage should be restored at all structures in the Restoration Area.			Low	High
Water Quality	Dissolved oxygen > 5 parts per million.	Delta	When spring pulse flows from Friant and the Tributaries cease, dissolved oxygen levels in the deepwater ship channel may drop to unsuitable levels.		To be modeled	Medium	High
Harvest	Harvest rates that do not result in an inadequate number of spawners to saturate the rearing habitat with juvenile salmon.	Friant to Estuary	Harvest is currently legal in the San Joaquin River from January through October. Neither legal or illegal harvest has been quantified.			Unknown	

**Table A-6.
Adult Holding (Section 4.5)**

PREFERRED CONDITIONS Mid-April to September		EXISTING CONDITIONS				Impact Level	Level of Certainty
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range			
				Location	Frequency		
Deep Pools	There are several large, deep pools in the 1-mile-long reach immediately downstream from Friant Dam.		The holding pools should be adequate to support up to about 18,000 adult salmon.	None		Low	High
Excessively High Water Temperatures	< 70°F.	Five-mile reach below Friant Dam	The temperature of the release flows at Friant Dam is expected to remain at suitable levels throughout holding.		Rare	Low	High
Legal and Illegal Harvest	Harvest rates that do not result in an inadequate number of spawners to saturate the rearing habitat with juvenile salmon.	Five-mile reach below Friant Dam	Harvest is currently legal throughout the year from Friant Dam downstream to the Highway 140 Bridge. Neither legal or illegal harvest has been quantified.			Unknown	Harvest
Disease	Low infestation rates associated with suitable levels of water temperatures, dissolved oxygen, crowding, ammonia, and contaminants.	Five-mile reach below Friant Dam	The conditions that may cause disease during holding are not expected to occur.		Rare	Low	High
Predation	Low predation rates typically occur when water temperatures, dissolved oxygen, crowding, ammonia, and contaminants are at suitable levels.	Five-mile reach below Friant Dam	The conditions that may cause predation during holding are not expected to occur.		Rare	Low	High

Potential stressors, suitable and existing conditions in the San Joaquin Basin, likely impact levels, certainty of the impact on spring-run Chinook salmon

Appendix A.

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Appendix B
Potential Stressors, Restoration Targets,
Existing Conditions, Likely Impact Levels,
Certainty of Impact on Fall-Run Chinook
Salmon

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**Table B-1.
Egg Survival and Emergence Life Stage (Section 4.1)**

SUITABLE CONDITIONS Spawning: Late October to Mid-December Incubation to Emergence: Late October to March		EXISTING CONDITIONS					
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range		Impact Level	Level of Certainty
				Location	Frequency		
Blocked Gravel Recruitment	Numerous beds of suitably sized gravel.	Reach 1A, particularly in the 10-mile reach below Friant Dam	Few spawning beds in 5-mile reach below Friant Dam.	River Miles 262.5 to 267.5		High	High
Elevated Water Temperature	<57°F.	Reach 1A, particularly in the 10-mile reach below Friant Dam	When Millerton's cold water pool becomes depleted, the temperature of releases flows would exceed 57°F. Even with suitable release flow temperatures, flows may be too low to provide suitable water temperatures below River Mile 262.5 where spawning beds exist.		>77% of years	High	Medium
Excessive Sedimentation	Fines < 10% in spawning beds.	Reach 1A, particularly in the 10-mile reach below Friant Dam	Most spawning beds appear to be highly silted between River Miles 262.5 and 267.5, but fine concentrations have not been measured.	River Miles 262.5 to 267.5		Medium	Low
Turbidity	Low turbidity during incubation.	Reach 1A, particularly in the 10-mile reach below Friant Dam	Unknown	Unknown	Unknown	Unknown	

**Table B-2.
Juvenile Rearing and Migration (Sections 4.2 and 6.1)**

PREFERRED CONDITIONS January to June				SPRING PULSE FLOWS PRESCRIBED IN SETTLEMENT			
Stressor	Water Year Type	Timing	Magnitude	Timing Potential to Extend Period by 3 Weeks	Magnitude	Impact Level	Level of Certainty
Truncated Springtime Pulse Flows	Wet Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 4,000 2,000 350	Low	High
	Normal-Wet Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 4,000 350 350	Medium	Medium
	Normal-Dry Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 2,500 350 350 350	High	High
	Dry Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 350 350 350 350	High	High
	Critical-High Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	1,500 200 200 215 255	High	High
	Critical-Low Years	Feb 15 – Jun 15	Sufficient to inundate floodplains and maintain water temperatures below 69°F in all reaches.	Mar 16 – 31 Apr 1-15 Apr16-30 May 1- Jun 30 Jul 1- Aug 31	130 150 150 190 230	High	High

**Table B-3.
Juvenile Rearing and Migration (Section 4.2)**

PREFERRED CONDITIONS January to June		EXISTING CONDITIONS					Impact Level	Level of Certainty
Stressor	Restoration Targets	Water Year Type	Location	Observed Range	Outside Preferred Range			
					Location	Frequency		
Excessively High Water Temperature	≤60 - 69°F (depending on food resources and water quality).	Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	High
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
Inadequate Food Resources	Prolonged inundation of floodplains and wetlands, productive riparian vegetation, numerous adult salmon carcasses, suitable water temperatures, low contaminants, natural channel morphology, low sedimentation, low competition with juvenile fall-run salmon and trout, and few exotic species.	Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	High
		Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
Predation	Natural channel morphology, natural hydrograph, and few exotic fish or planted catchable-sized trout.	Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium
		Normal-Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
Disease	Low infestation rates when water temperatures, water quality, and food resources are adequate	Dry Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-High Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Critical-Low Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	High	Medium
		Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Low	Medium
		Normal-Wet Years	Friant to Estuary	To be modeled	To be modeled	To be modeled	Medium	Medium

**Table B-3.
Juvenile Rearing and Migration (Section 4.2) (cont'd)**

PREFERRED CONDITIONS January to June		EXISTING CONDITIONS					Impact Level	Level of Certainty
Stressor	Restoration Targets	Water Year Type	Location	Observed Range	Outside Preferred Range			
Entrainment	No diversions during rearing and migration or intakes have effective screens.	All	San Joaquin River: Friant Dam downstream to Chipps Island	Unknown	Unknown	Unknown	Unknown	Unknown
Degraded Physical Habitat	Natural channel morphology, natural riverbed sediment sizes, and natural levels of large woody debris.	All	Friant to Estuary	A majority of the physical habitat has been degraded in the San Joaquin River basin from numerous causes, including blocked sediment recruitment by Friant Dam, dikes and levees that separate the low flow channel from floodplains, in-channel gravel excavation, sedimentation, removal of riparian vegetation, and channel dredging in the deepwater ship channel.			High	Low
Contaminants	Zero to low level contaminants, including pesticides, herbicides, and certain trace elements (e.g., selenium).	All	Friant to Estuary	Contaminant levels are high in the San Joaquin River basin, particularly during spraying to control West Nile virus.			High	Medium
Sport Harvest	Low harvest rates of yearling-sized Chinook salmon.	All	Friant to Estuary, but primarily Reach 1A during summer and fall	It is assumed that subyearling smolts primarily contribute to adult recruitment; therefore, the harvest of yearlings will not have a substantial impact.			Low	Low
Water Quality	High concentrations of dissolved oxygen and low levels of ammonia.	All	City of Stockton to Jersey Point	Dissolved oxygen levels frequently become unsuitable in the deepwater ship channel when Vernalis flows fall below 2,000 cubic feet per second and outflow from wastewater treatment plants can increase ammonia levels.			Medium	Medium

**Table B-4.
Ocean Survival (Section 4.3)**

PREFERRED CONDITIONS Juvenile Entry: May to July Adults: 2-5 Years		EXISTING CONDITIONS				Impact Level	Level of Certainty
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range			
				Location	Frequency		
Inadequate Food Resources	Upwelling of nutrient-rich water in coastal waters and large freshwater plumes from the San Francisco Estuary.	Ocean	There are 15- to 25-year cycles of productive and unproductive periods. An unproductive period began in 2002. There were unusually poor conditions in 2005 that are affecting 2007 escapements. Freshwater plumes are probably relatively small except during uncontrolled releases in wet years due to high rates of diversions throughout the Central Valley and high rates of exports at the Central Valley Project and State Water Project pumping facilities.			Medium	High
Predation	Natural rates of predation by fish, bird, and mammal.	Ocean	Predation levels are highest when juvenile abundance is high.			Unknown	
Adult Commercial and Sport Harvest	Harvest should not reduce escapement to the point that the rearing habitat cannot be saturated with juvenile salmon. The rearing habitat in the San Joaquin River Tributaries can be saturated by the progeny of about 500 adult fall-run Chinook salmon.	Ocean	The Central Valley Index of ocean harvest ranged from 26% to 62% between 2000 and 2006. The non-landed fishing mortality south of Horse Mountain, Humboldt County, California, ranged between 9% to 15% (mean 11.8%) of the commercial catch and between 11% and 16% (mean 12.2%) of the sport catch.			Unknown	

**Table B-5.
Adult Migration (Section 4.4)**

PREFERRED CONDITIONS Late September to Early December		EXISTING CONDITIONS				Impact level	Level of Certainty
Stressor	Restoration Targets	Location	Observed Conditions	Outside Preferred Range			
				Location	Frequency		
Inadequate Flows and High Delta Exports	Sufficient flow from Friant Dam must reach the San Francisco Estuary to provide migratory cues to allow adult salmon to home to the San Joaquin River. Delta exports should not exceed 400% of flows for at least a 10-day period in mid-October.	Friant Dam Releases	Dry to wet year attraction flows from Nov 1-10: 700 cubic feet per second			Medium	Medium
			Critical-high year attraction flows from Nov 1-10: 400 cubic feet per second			Medium	Medium
			Critical-low year attraction flows from Nov 1-6: 130 cubic feet per second			Medium	Medium
Excessively High Water Temperatures	< 66°F	Friant to Delta	It is likely that water temperatures in the lower Restoration Area and Delta will exceed suitable levels without fall pulse flows.		All Years typically until mid-November	High	High
Structural Barriers	Suitable water depths > 1 foot and screened diversion canals.	Restoration Area	The Settlement prescribes that passage should be restored at all structures in the Restoration Area.			Low	High
Water Quality	Dissolved oxygen > 5 parts per million.	Delta	When spring pulse flows from Friant and the Tributaries cease, dissolved oxygen levels in the deepwater ship channel may drop to unsuitable levels.		To be modeled	Medium	High
Harvest	Harvest rates that do not result in an inadequate number of spawners to saturate the rearing habitat with juvenile salmon.	Friant to Estuary	Harvest is not legal in the San Joaquin River during November and December. Illegal harvest has been quantified.			Unknown	